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ENVIRONMENTAL PROTECTION AGENCY

Radiation Office

XUM

INTERNATIONAL NUMERICAL MULTIPLE AND SUBMULTIPLE PREFIXES

Multiples and submultiples	Prefixes	Symbols	Pronunciations
10^{12}	tera	T	tĕr'a
10^{9}	giga	G	jĭ'ga
10^{6}	mega	M	mĕg'a
10^{3}	kilo	k	kĭl'o
10^{2}	hecto	h	hĕk'to
10	deka	da	dĕk'a
10^{-1}	deci	d	dĕs'i
10^{-2}	centi	c	sĕn'ti
10^{-3}	milli	m	mĭl'i
10^{-6}	micro	μ	mi'kro
10^{-9}	nano	n	năn'o
10^{-12}	pico	p	pē'ko
10^{-15}	femto	f	fĕm'to
10^{-18}	atto	a	ăt'to

SYMBOLS, UNITS, AND EQUIVALENTS

Symbol	Unit	Equivalent
Å	angstrom	10^{-10} meter
a	annum, year	
BeV	billion electron volts	GeV
Ci	curie	3.7×10^{10} dps
cm	centimeter(s)	0.394 inch
cpm	counts per minute	
dpm	disintegrations per minute	
dps	disintegrations per second	
eV	electron volt	1.6×10^{-12} ergs
g	gram(s)	
GeV	giga electron volts	1.6×10^{-3} ergs
kg	kilogram(s)	1,000 g = 2.205 lb.
km^2	square kilometer(s)	
kVp	kilovolt peak	
m^3	cubic meter(s)	
mA	milliampere(s)	
mCi/mi^2	millicuries per square mile	0.386 nCi/m^2 (mCi/km^2)
MeV	million (mega) electron volts	1.6×10^{-6} ergs
mg	milligram(s)	
mi^2	square mile(s)	
ml	milliliter(s)	
mm	millimeter(s)	
nCi/m^2	nanocuries per square meter	2.59 mCi/mi^2
pCi	picocurie(s)	10^{-12} curie = 2.22 dpm
R	roentgen	
rad	unit of absorbed radiation dose	100 ergs/g

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RADIOLOGICAL HEALTH DATA AND REPORTS

Volume 12, Number 1, January 1971

Radiological Health Data and Reports, a monthly publication of the Environmental Protection Agency, presents data and reports provided by Federal, State, and foreign governmental agencies and other cooperating organizations. Pertinent original data and interpretive manuscripts are invited from investigators.

In August 1959, the President directed the Secretary of Health, Education, and Welfare to intensify Departmental activities in the field of radiological health. The Department was assigned responsibility within the Executive Branch for the collation, analysis, and interpretation of data on environmental radiation levels. This responsibility was delegated to the Bureau of Radiological Health, Public Health Service. Pursuant to the Reorganization Plan No. 3 of 1970, effective December 2, 1970, this responsibility was transferred to the Radiation Office of the Environmental Protection Agency which was established by this reorganization.

The Federal agencies listed below appoint their representatives to a Board of Editorial Advisors. Members of the Board advise on general publications policy; secure appropriate data and manuscripts from their agencies; and review those contents which relate to the special functions of their agencies.

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ENVIRONMENTAL PROTECTION AGENCY
William D. Ruckelshaus, Administrator

Reports

Estimates of Potential Doses to Various Organs from X-radiation Emissions from Color Television Picture Tubes

Robert H. Neill, Harry D. Youmans, and *John L. Wyatt*[1]

Using the exposure rates determined in the metropolitan Washington color television survey, doses to various organs of children and adults were calculated. For a television receiver with a picture tube emitting 0.5 mR/h (the maximum exposure rate allowed under the Department of Health, Education, and Welfare performance standards) and with other specified assumptions, the calculated dose to a child ranged from 5 mrem/a (female gonads) to 42 mrem/a (skin). For an adult, the doses were 2 mrem/a to 18 mrem/a, respectively. Using the same assumptions but for a picture tube emitting 0.043 mR/h (the average exposure rate observed in the metropolitan Washington survey), the respective doses were 0.4 mrem/a to 3.6 mrem/a for a child and 0.2 mrem/a to 1.6 mrem/a for an adult.

This document presents estimates of organ doses that could result from color television receiver x-ray emissions under two operating conditions. The first condition of operation assumes that the picture tube emits 0.5 mR/h at 5 cm from the front face, which is the maximum allowable level under the Department of Health, Education, and Welfare (DHEW) performance standard (*1*). The second assumes that the front face of the picture tube emits x radiation at 0.043 mR/h, the average rate observed in the metropolitan Washington survey (*2*). Since all receivers emitting radiation in excess of the allowable level were corrected during the survey, the dose estimates represent the dose that would have occurred before correction. All calculations are based on the viewing habits indicated in the metropolitan Washington survey.

Types of exposure

While there are various ways that individuals can be exposed to x radiation emanating from color television receivers, two major types of exposure to people are:

1. picture tube face radiation (primarily a source of exposure to those watching television),
2. radiation from the shunt regulator tube, the rectifier tube, and other parts of the picture tube (possible sources of exposure to all persons in the room with the television receiver).

Dose estimates are presented only for the former case. None are presented for the latter due to lack of applicable data, since such estimates would involve the dose to an individual located at a specified distance from a television set operating for a specified length of time and emitting beams of specified directions, areas, and shapes.

Viewing population

While the Electronics Industries Association (EIA) has estimated that 95 percent of the approximately 60 million households in the United States have television receivers, only about 39 percent have color television (*3*). In view of the current production rate of 500,000 color television receivers per month and projections for the future, it is reasonable to presume that virtually the entire population will be color television viewers.

Although we give no consideration to multiple sources of exposure, it is interesting to note that the EIA estimates that 29 percent of the household

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in the United States now have more than one television receiver.

Assumptions

The following assumptions are necessary to these calculations:

1. Hours per year that an individual spends viewing television

On the basis of the metropolitan Washington survey, values of 1,100 hours per year (3 h/day) for children 14 years of age and under and 950 h/a (2.6 h/day) for people over 14 years old were selected as being typical (*2*). While these figures are similar to the 1,000 h/a used by Braestrup (*4*), they are somewhat lower than those recently reported by the A. C. Nielsen Company (*5*).

2. Distance from television receiver

The limited data available on viewing distances from the Washington survey (*2*) indicate typical figures of 165 cm (approximately 5 feet, 5 inches) for children 14 and under and 250 cm (approximately 8 feet, 2 inches) for the remainder of the viewing population.

3. Reduction in intensity with distance

The distance factor expresses the reduction in intensity as a function of distance from the screen. The reduction first begins to follow the inverse square law at a distance of about 75 cm, because the screen is not an isotropic point source. Data used in this report were taken from Wang (*6*) (figure 1) and represent the decrease in exposure rate along the central axis of a 25-inch, color television tube. No attempt has been made to take into account the fact that some people sit slightly to one side of the receiver.

These data are presented in figure 1. Curves for exposure reduction at a 30° angle from the central axis of the 25-inch, black and white tube are presented for comparison. While it is recognized that many color television receivers have smaller picture tubes, the only available data are for the larger sizes. Radiation levels will, in general, fall off more rapidly with distance from a small screen receiver than from a larger one. The distance factors in x-ray intensity associated with viewing distances of 165 cm and

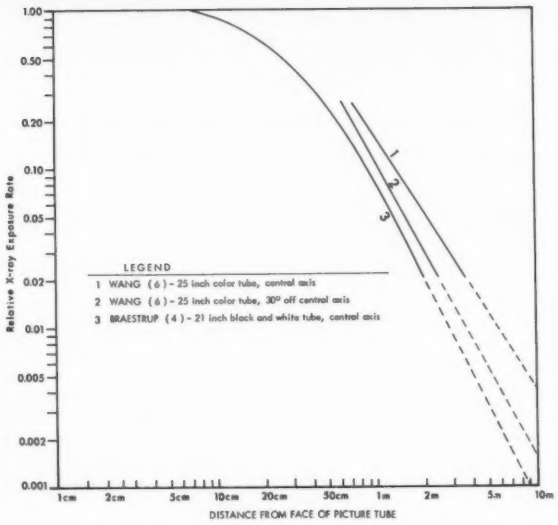

Figure 1. Reduction in x-ray intensity as a function of distance

250 cm for children and adults were 0.070 and 0.036 cm.

4. Backscatter factor

Radiation dosage in a sample of tissue results not only from photons traveling in the primary beam and impinging upon the sample, but also from photons and electrons scattered backward from tissue at a greater depth than the sample under consideration. The increase in dosage from this latter effect is represented by the tissue backscatter factor. In this report, a tissue-backscatter factor of 1.24 was used. This value applies to a 17.5 cm diameter beam with a half-value layer of 1.0 mm of aluminum (*7*).

5. Exposure-to-dose conversion factor

The following conversion factor in muscle at 30 keV was used (*8*).

$\bar{f} = 0.91$ rad/R

6. Organs and depth dose factors

The organs considered are those discussed by Davis, Youmans, Wolff, and Mills (*9*). Bone marrow and central nervous system are not included because of attenuation afforded by bone, large volume, wide distribution in the body, and highly nonuniform dose distribution.

Radiation dosage at a given depth in the body may be less than that at the surface of the skin because of the shielding afforded by intervening layers of tissue. This decrease in

radiation dosage with depth is represented by the depth dose factor.

The depth dose factors were taken from Mills (*10*) except where indicated. They are appropriate for use with any x-ray beam of spectral characteristics similar to those of radiations from television receivers (*11*). Values used were skin, 0.95; adult thyroids, 0.56; juvenile thyroids, 0.70 (an estimate based on the assumption of 8 mm depth below surface); male gonads, 0.41; female gonads, 0.11 (*4*); lens of eye, 0.80.

7. Quality factor

A quality factor of 1 mrem/mrad was used (*12*) throughout the calculations.

Computational method

The estimated average annual dose is computed as follows:

$$D = X\,T\,d\,B\,\bar{f}\,F\,Q = 1.13\,X\,T\,d\,F$$

where

D = estimated average annual dose (mrem/a),
X = exposure rate at 5 cm from the front face of the picture tube (mR/h),
T = annual viewing time (h/a),
d = distance factor,
B = backscatter factor (1.24),
$\bar{f}$ = conversion factor (0.91 mrad/mR),
F = depth dose factor, and
Q = quality factor (1 mrem/mrad).

Discussion

Table 1 presents estimates of the annual dose to children and adults which could result if x-ray emissions from the front face of the picture tube are 0.5 mR/h, the maximum allowable under the DHEW performance standard.

The average front face exposure rate observed from the 1,124 color television receivers measured in the metropolitan Washington survey was 0.043 mR/h. The distribution is presented in figure 2.

Estimates of the average annual dose for both children and adults in households surveyed in the metropolitan Washington survey are shown in table 2. The difference is attributed primarily to differences in viewing habits of the two groups; children watch longer and sit closer than adults. The estimates are average doses which are based on average viewing habits and an average dose rate.

Table 1. Estimated average dose from the front face of picture tubes emitting 0.5 mR/h

Organ	Children less than 15 years of age (mrem/a)	People 15 years of age and older (mrem/a)
Skin	42	18
Thyroid	30	11
Male gonads	18	8
Female gonads	5	2
Lens of eye	34	15

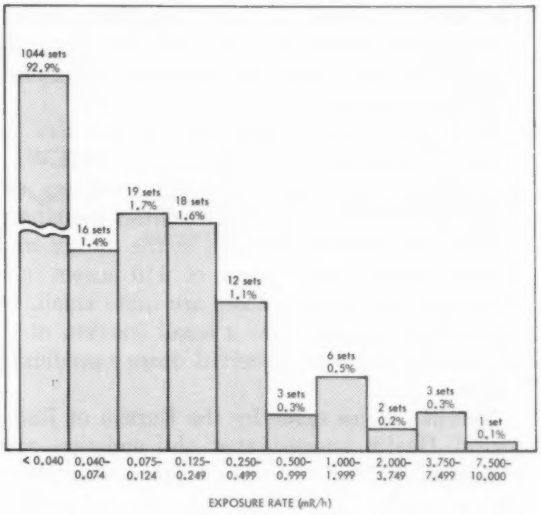

Figure 2. Front-face exposure rates observed from 1,124 color television receivers in the metropolitan Washington survey

While estimates of maximum doses are fraught with uncertainties due to the paucity of physical measurements and variability of viewing habits, we have attempted a semi-quantitative estimate as follows: about 1.33 percent (15/1,124) of the color television receivers measured in the metropolitan Washington survey emitted radiation from the front face of the picture tube at 0.5 mR/h or

Table 2. Estimated average dose from the front face of picture tubes emitting radiation at the average rate observed in the Washington metropolitan survey (0.043 mR/h)

Organ	Children less than 15 years of age (mrem/a)	People 15 years of age and older (mrem/a)
Skin	3.6	1.6
Thyroid	2.6	.9
Male gonads	1.5	.7
Female gonads	.4	.2
Lens of eye	3.0	1.3

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more. The average dose rate for these 15 receivers was 2.7 mR/h. This would suggest that 1.33 percent of the viewing population might receive doses which are larger than those presented in table 2 by a factor of about 60 and about 5½ times the doses shown in table 1. Data collected during the survey indicate that 9 percent of the children viewed television at a distance of 4 feet or less. The dose estimates may be increased by a factor of about 2 for this group. These data also indicated that over 20 percent of children watched television about 2,000 hours per year. For this group, the calculated dose could be increased by another factor of almost 2.

It is interesting to note that the Ad Hoc Advisory Committee to the Secretary, DHEW, on x radiation from color television receivers estimated on the basis of the metropolitan Washington survey that 140,000 children in the nation could receive annual skin doses of 110 mrem (*13*). Although the average doses are quite small, the calculations suggest that a small fraction of the population may have received doses considerably larger.

Recent studies made by the Bureau of Radiological Health suggest that the emission rates published in 1968 from the Metropolitan Washington survey (*2*) may be high due to a lack of correction for energy dependence of the survey meters used. The appropriate correction factors, which are not presently available, may result in the published average emission rates being high by a factor of 1 to 2, and the maximum emission rates being high by a factor of 3. The average and maximum doses as reported herein, reflect those doses which would have occurred at the emission rates stated in the Washington survey report and under the assumed conditions.

Although only 1.33 percent of the receivers in the Washington, D.C., survey emitted 0.5 mR/h or more from the front face, 5.9 percent emitted 0.5 mR/h or more at some location 5 cm from the surface of the receiver. The Dade County Department of Health in Florida reported that, of 507 color receivers surveyed between 1967 and 1970, 9.1 percent emitted over 0.5 mR/h at some location.[4] In contrast, the Suffolk County Department of Health in New York in a revised statistical report on 5,429 color receivers surveyed between January 1968 and January 1970 obtained 16.2 percent for the percentage of receivers emitting over 0.5 mR/h at some point (*14*).

Due to the lack of applicable data, no allowance has been made for the fact that many people view the television screen from an angle or that children often lie on their stomachs to watch television while many adults prop their feet up on a stool. Viewing from an angle would cause a reduction in the dose to all organs (*6*), whereas lying face down or propping one's feet on a stool would tend to cause a reduction in gonadal dose.

For comparison purposes, the appropriate guidelines of the Federal Radiation Council (FRC) (*15, 16*) and the International Commission on Radiological Protection (ICRP) (*12, 17*) for limiting doses to the public are shown in table 3. The values shown are exclusive of doses from natural background and medical radiation exposure.

Table 3. Recommendations of the FRC and ICRP for limiting radiation exposure[a]

Organ	Individuals (mrem/a)	Populations (mrem/a)
Skin	3,000 (*12*)	
Thyroid	1,500 (*15*)	500 (*15*)
Gonads	500 (*12,16*)	[b]170 (*16*)
Lens of eye	1,500 (*15*)	

[a] These recommendations exclude occupational exposure, natural background and medical radiation exposure.
[b] Stated as 5 rem/30 years.

While the above guidance for populations is not directly comparable to the dosage estimates in this paper, it is included to indicate the relative magnitudes. The National Council on Radiation Protection and Measurements has recommended that students under 18 years of age not receive an additional exposure of more than 100 mrem/a as a result of educational activities. The average annual dose from natural background has been estimated to be between 90 and 200 mrem/a in most of the United States (*18*).

It is well recognized that such recommendations are based upon the number of individuals that may be exposed as well as the magnitude of exposure. Also, any exposure involves some degree of risk, and unnecessary exposure should be

[4] Unpublished synopsis of the results of the Dade County surveys for 1967–1970, April 15, 1970.

avoided regardless of amount. This is especially true in a time when the availability of products which produce radiation continues to increase and the knowledge of synergistic responses remains inadequate. There is evidence to indicate a general downward trend in radiation exposure rates from color television receivers since 1965. Figures 3 and 4, prepared from data published in the Washington survey, show this downward trend. The same trend can be noted in the results of surveys performed by the Dade County, Fla., Department of Public Health, from which figure 5 was taken. The decrease may be attributed to improved design, improved quality control procedures or increased awareness on the part of service repairmen of proper voltage settings or a combination of all three.

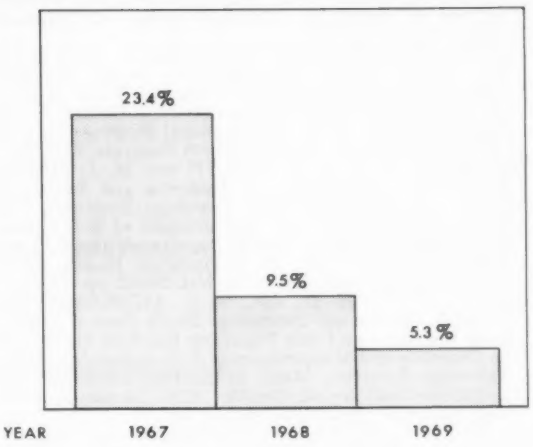

Figure 5. Percentage of receivers emitting 0.5 mR/h or more observed in the Dade County, Florida surveys (by year of survey)

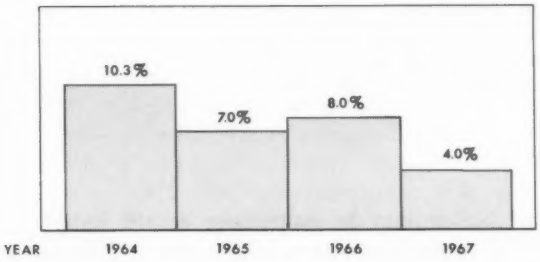

Figure 3. Percentage of color television receivers observed in the metropolitan Washington survey emitting 0.5 mR/h or more (classified by date of purchase)

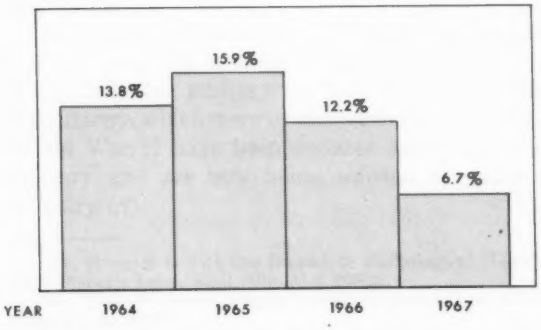

Figure 4. Percentage of color television receivers observed in the metropolitan Washington survey emitting 0.25 mR/h or more (classified by date of purchase)

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Surveys of Radium Contamination in Aircraft Instrument Repair Facilities 1966–1967

Gail D. Schmidt,[1] *Jerome A. Halperin,*[2] *Eric Geiger,*[3] and *Walter Livingstone*[4]

A survey of 30 aircraft instrument repair facilities revealed many cases of significant radium contamination, resulting from repair and dial stripping operations on radium luminous dial instruments. Five out of 15 facilities surveyed in Wichita, Kans., showed alpha contamination exceeding 200,000 dpm/100 cm², with a maximum of 1,650,000 dpm/100 cm². The direct gamma radiation ranged from background to 60 mR/h. Whole body counts of 33 persons working in these facilities showed radium body burdens in two people that were about 40 percent of the maximum permissible body burden. A similar survey of 15 repair facilities in Dade County, Florida, showed lower alpha contamination levels by a factor of 10.

Based upon these findings, a radiation control program including contamination guides has been recommended for instrument repair facilities.

United States production of radium-226 has amounted to about 200 grams and about 1,800 grams have been imported (*1*). A substantial portion of this was utilized during World War II in radioactive self-luminous compounds for instruments, dials, and markers. Although production and use data are not available, estimates place the quantity of radium thus used at several hundred grams (*2*).

One major use of radium luminous compounds was to illuminate aircraft instruments (figure 1), particularly for military aircraft. Many of these instruments which were produced during and after World War II have been declared surplus by the military and are now being utilized by private industry (*2*).

Figure 1. Radium dial instrument

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[2] Mr. Halperin is radiological health representative, EHS, Chicago, Ill. 60607.
[3] Mr. Geiger is with the Eberline Instrument Corporation, Santa Fe, N. Mex.
[4] Mr. Livingstone is with Dade County Department of Public Health, Miami, Fla.

To refurbish these instruments, the dials containing radium are removed, the old paint stripped from the dial, and the dial repainted. These tear down and dial stripping operations have produced significant contamination in the facilities where this type of work is done. These aircraft instruments containing radium are not labeled, and very few of the workers in these facilities are aware that a radiation hazard is involved. A study of luminising facilities in the United Kingdom has shown a correlation between facility contamination levels and radium body burdens of the workers (*3*).

Radium does not come under the regulatory authority of the Atomic Energy Commission and has only recently been regulated in those States where appropriate legislation exists.

Facilities which repair instruments for flight use must be certified by the Federal Aviation Agency (FAA) but no provision for radium control is included in the certification. The FAA has, however, recently issued a circular advising repair stations of the radiation hazard associated with the repair and maintenance of radium dial instruments (*4*).

Instrument repair and dial stripping operations

In order to evaluate the radiation hazard and to develop a recommended radiation control program for aircraft instrument repair facilities, two studies were conducted for the U.S. Public Health Service. The first study was conducted under contract[5] by the Eberline Instrument Corporation, Santa Fe, N. Mex., with the cooperation and assistance of the Kansas State Department of Health and the Wichita-Sedgwick County Health Department. The second study was conducted by the Florida State Board of Health and the Dade County Department of Public Health in cooperation with the Bureau of Radiological Health, U.S. Public Health Service. Each study involved the survey of 15 facilities. In the Eberline contract study, 14 facilities were located in the Wichita, Kans., area, which is primarily a light aircraft fabrication area and the 15th facility in New Mexico.

The 15 facilities in the Florida study are located in the Dade County, Fla., area and are primarily associated with providing service to commercial and private airlines.

The nature of the operations in these facilities can be broken down into the following areas:

Dial stripping and refinishing. After removal from the instrument, the dial faces are generally stripped. The first step involves soaking in a solvent, which contains halogenated hydrocarbons, chlorinated hydrocarbons, phenol and/or similar solvents. The dial face is then scrubbed with a brush to remove adhering deposits, rinsed in a detergent to remove oily deposits, and prepared for refinishing. The refinishing with nonradioactive paints is carried out with a silk screen stencil that is specific for a particular dial. Thus it is necessary for a facility to accumulate a large number of silk screens to match the many instrument types. In many cases, the old silk screens are contaminated with radium from former radium paintings and are a source of considerable contamination.

Instrument disassembly and repair. Instruments are disassembled in one area prior to transfer to a repair bench. The instrument disassembly area is another major source of contamination, since the radium paint is frequently chipped or cracked during this operation. Following repair, the instruments are tested and recalibrated.

Storage. Since there are many types and models of aircraft instruments and surplus sales generally involve a large group of instruments, most facilities have a large stock of used and surplus instruments and dials.

Other operations. Included in these surveys were several facilities, including distributors or manufacturers which did not currently work on radium dial instruments but had detectable contamination from past operations.

The 30 facilities in this study carried out the following operations on radium dial instruments.

Dial stripping and refinishing	10
Instrument disassembly and repair	21
Storage	approximately 25
Other	4

Twenty-two of these facilities were certified by the FAA (*5*). The distribution of these facilities by the FAA certification speciality of instrument,

[5] Contract # PH-86-67-66.

radio and accessory repair are shown in table 1 along with the total number of such certified stations in the United States (*5*). These repair specialties are not mutually exclusive since a given facility may be certified in several specialties. Those facilities surveyed but not FAA certified, include dial strippers and refinishers, distributors, and others which do no actual repair of flight instruments.

Table 1. Facilities certified by the Federal Aviation Agency

Repair specialty	Total FAA certified stations	FAA certified stations in this study
Instrument	386	21
Radio	559	10
Accessory	486	10
Total facilities		22

Evaluation of radiation hazards

Surface contamination levels. The results of the field surveys of the 30 facilities involved are summarized in table 2. Total alpha contamination was determined using portable alpha survey meters (gas flow and alpha scintillation). These instruments were calibrated on each scale with standard plutonium alpha sources.

Significant amounts of removable long half-life alpha contamination were found in all cases where dial stripping was currently being performed or had been performed in the past. Lower levels of fixed and removable alpha contamination were found in instrument disassembly and storage areas. Dial refinishing areas showed some contamination which could be attributed to old silk screens and equipment that had been used in the past when dials were painted with radium luminous compounds.

Each facility will be described in terms of its maximum contamination since this parameter can readily be compared between facilities. The use of an average contamination value would be more appropriate; however, this is difficult to obtain where surveys have been conducted by different organizations, and a few high values can significantly affect the average, as will be shown later.

Figures 2, 3, and 4 show the distribution of facilities by maximum total and maximum removable alpha contamination on floors and benches.

Table 2. Contamination levels in aircraft instrument repair facilities[a]

Facility		Area	Maximum alpha contamination (dpm/100 cm²)	
Code	Type of operation		Total	Removable (long-lived)
Light aircraft fabrication area				
A	Stripping and repair	Benches	200,000	170
		Floors	100,000	270
B	Stripping and repair	Benches	167,000	2,100
		Floors	334,000	1,000
C	Stripping	Benches	[b]330,000	2,700
		Floors	[c]16,500	1,400
D	Stripping and repair	Benches	>330,000	500
		Floors	100,000	<200
E	Stripping and repair	Benches	16,700	190
		Floors	6,700	750
F	Stripping and repair	Benches	3,300	<100
		Floors	5,000	<100
Commercial aircraft area				
1	Stripping and repair	Benches	170,000	5,000
		Floors	4,700	1,850
2	Repair	Benches	136,000	350
		Floors	20,000	220
3	Repair	Benches	20,000	280
		Floors	68,000	790
4	Stripping and repair	Benches	34,000	280
		Floors	34,000	54
5	Stripping and repair	Benches	1,400	430
		Floors	2,700	24

[a] Of the facilities not specifically listed here, 13 had no contamination above background and the other 6 had only spotty contamination.
[b] Sink had a maximum value of 1,650,000 dpm/100 cm².
[c] Floor had been recently tiled, concrete under the tile may be more highly contaminated.

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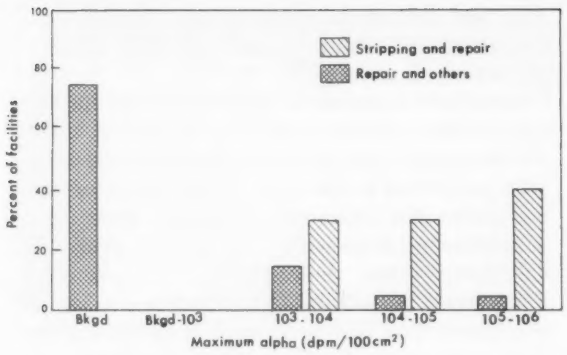

Figure 2. Distribution of facilities by maximum total alpha contamination on benches

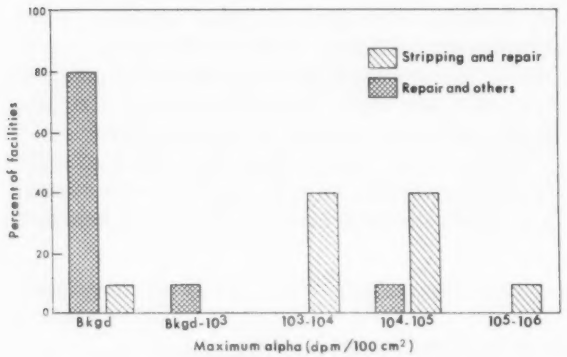

Figure 3. Distribution of facilities by maximum total alpha contamination on floors

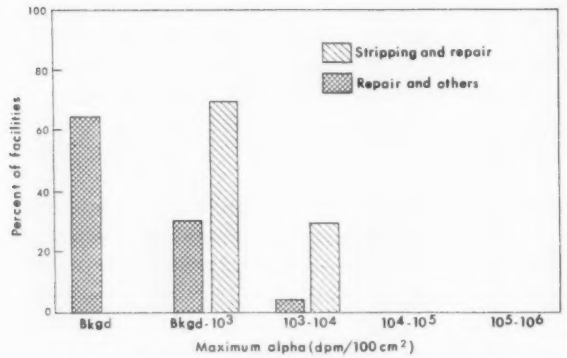

Figure 4. Distribution of facilities by total alpha removable contamination on floors and benches

It is observed that those facilities which performed dial stripping operations have significantly higher levels of contamination than those facilities which do not strip dials.

Radon air levels. Radon levels were measured by radon film badge (*6*) in the light aircraft facilities and by alpha counting of air filter samples for radon daughters (*7*) in the commercial airline facilities. Radon film badges worn for several months in facilities A, B, and H, generally indicated levels less than 30 pCi/liter with none exceeding 100 pCi/liter, the recommended occupational concentration guide for radon-222 in air (*8*). Radon film badges worn in facility C, where operations involved dial stripping and refinishing only, gave an average concentration of 370 pCi/liter (range, 40–1,400 pCi/liter) or almost four times the recommended occupational concentration guide in air. A large number of radium dials were stored and stripped in facility C, and this apparently accounts for the higher radon levels measured in this facility.

The radon levels in the commercial airline facilities did not exceed approximately $\frac{1}{10}$ of the recommended occupational concentration guide for air in either the storage or the work areas.

Airborne alpha levels. Long-lived alpha-particle concentrations in air were determined in facilities A, B, C, and H by filtration. The filters were counted with a large area, gas-flow proportional counter 5 days or more after collection so that only long half-life alpha-particle emitters were measured.

Table 3 summarizes the airborne concentrations of long half-life alpha-particle emitters for facility B, which was doing a large amount of disassembly work during the study, but no stripping of dials, and for facility C, which was doing a large amount of dial stripping work but no disassembly of

Table 3. Summary of long half-life alpha air concentrations—facilities B and C

Location	Number of samples	Average air concentration (pCi/m³)
Facility B		
Teardown area	9	1.13
Old stripping area	9	.09
Facility C		
Stripping area	13	2.5
Refinishing room	3	.53

instruments during the study. The concentrations were generally less than 10 percent of the recommended concentration guide of 30 pCi/m³ for occupational exposure to radium-226 (*8*).

Air samples from facility C indicated higher airborne concentrations of long half-life alpha emitters than in facility B, but in no case did the concentration approach the concentration guide. The maximum concentration measured was 7.4 pCi/m³ which is 25 percent of the concentration guide in air if it is assumed that all of the activity is radium-226 (no daughters present).

Air samples were collected in facility A and H over a period of several months. Concentrations of long half-life alpha emitters in these facilities were generally below 1 percent of the concentration guide for occupational exposure.

External hazards. Gamma radiation levels above 2 mR/h were frequently observed. However, these were generally in storage areas, and radiation levels in work areas were localized or the area had a limited occupancy time. The storage area was usually a drawer or box where dial faces were stored or shelves where complete instruments were stored. Radiation levels to which a significant portion of the body could be exposed were generally 5 mR/h or less with one exception; a desk drawer containing a large number of dials gave a reading of 13 mR/h at 30 cm (1 foot) from the top of the drawer.

Significant levels of gamma radiation were also present in areas where paint and residues from the dial stripping operation had accumulated. The distribution of the facilities by maximum gamma levels and work area gamma levels are shown in figures 5 and 6, respectively. Again it is observed that radiation levels are generally greatest in those facilities that strip dials.

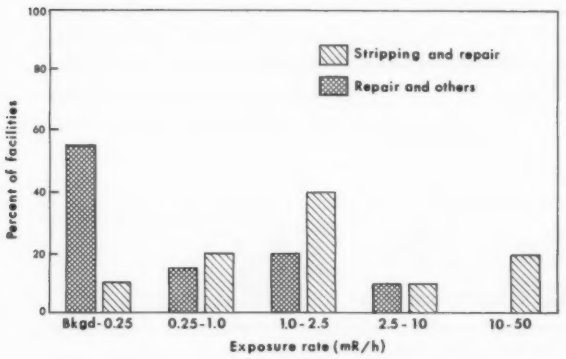

Figure 5. Distribution of facilities by maximum gamma radiation

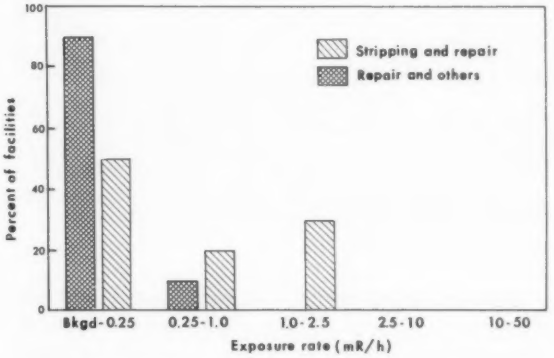

Figure 6. Distribution of facilities by work area gamma levels

Beta-gamma film badges were worn for 2 months in facilities A, B, and C, with biweekly exchange. Beta-gamma film badges were worn in facility H for 6 months with monthly exchange. All of these badges showed exposures of less than 10 mR gamma and less than 10 mrad beta exposure (the minimum detectable level for the film badge).

Internal hazards. The mobile whole body counting facility of Helgeson Nuclear Services, Inc., was used to determine the radium content of selected workers from each participating company in the light aircraft fabrication area. The detector was a 20 cm (8 in.) diameter by 10 cm (4 in.) thick NaI(Tl) crystal with a cesium-137 resolution of 7.6 percent. A shadow shield was used to reduce the background.

The whole-body counter was located at the Wichita-Sedgwick County Health Department on July 5, 6, and 7, 1967, and employees of the participating companies were scheduled to report there for counting. Two individuals who had not worked with radium were counted as controls. After showering carefully and putting on a clean pair of coveralls, all persons were counted for 20 minutes. The body burden calculation was based on a 30-percent retention of the radon gas from radium in the body (*9*) and the 1.76 MeV gamma of bismuth-214. Only data for the two persons from facility C (a dial stripping operation), whose body burdens were 41 and 43 nCi of radium, are considered significant (table 4).

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Table 4. Radium-226 body burden assessments

Facility	Number of individuals	Whole body count [a](nCi)	Urinalysis (pCi/24 h)
A	4	≤ 9 ± 9	≤0.03 ± 0.12
B	4	≤ 7 ± 8	.08 ± .08
C	1	41 ± 10	.66 ± .20
	1	43 ± 10	.23 ± .12
	1	6 ± 9	.16 ± .10
D	1	10 ± 10	
	5	≤ 5 ± 8	
E	1	14 ± 10	
	1	10 ± 10	
	7	≤ 6 ± 10	
6	1	[b]2 ± 11	

[a] Values listed are individual body burdens or the maximum value when given for a number of individuals.
[b] Body burden based upon radon breath measurement.

Urine samples were also collected and analyzed for radium. The results for these samples (table 4) were all less than 0.2 pCi/24 h, except for samples from the two individuals at facility C who had significant body burdens. The radioactive material in these urine samples was positively identified as radium-226 by alpha spectrometry.

One individual in the "commercial airline area" was assessed for radium body burden by radon breath sampling. The results were negative as shown in table 4. This individual has worked in instrument repair and dial painting for 25 years and repairs watches and paints hands and faces with radium luminous compounds in his home workshop.

Contamination study. In order to thoroughly evaluate the radiation hazards involved, a detailed survey was made of the contamination levels in facilities A, B, and C. Table 5 summarizes the total alpha contamination of these facilities (the data in table 5 is from a survey in June 1967 and differs slightly from the data in table 2 which includes data from the initial survey in December 1966). The average gross alpha contamination level of 50,000 dpm/100 cm² for facility C is a factor of six higher than A or B. If the maximum value of 1,650,000 dpm/100 cm² is not included in the average, the difference is only a factor of about 2.4.

Since these averages appeared subject to considerable fluctuation, the "area integral" was calculated as an attempt to derive a more meaningful measure of facility contamination. The surface contamination measurements were plotted on a facility floor plan and isopleths defining areas of essentially equal contamination were drawn. The "area integral" contamination value is then the summation over the facility floor area of the contamination levels times the fractional area involved. The "area integral" value of 4,100 dpm/100 cm² for facility C is twice that of facility A or B. Graphically, facility C appeared more highly contaminated with general levels of about 1,000 dpm/100 cm², whereas facilities A and B had general levels of about 50–200 dpm/100 cm². The "area integral" values noted above (table 5) result primarily from the limited areas of high contamination in the stripping and disassembly area.

It was expected at the onset of these studies that the data might help answer the question—what is an acceptable contamination level? As noted above, only facility C had individuals with positive radium body burdens. Thus, the conclusions that can be drawn with regard to a relation between contamination and internal radiation hazard is rather limited. The negative data do, however, allow one to draw the general conclusion that work under these conditions has not

Table 5. Total alpha contamination characteristics of facilities A, B, and C[a]

Contamination characteristic	Total alpha[b] (dpm/100 cm²)		
	A	B	C
Maximum	100,000	82,000	1,650,000
Gross average	8,300 (67)	7,900 (67)	50,000 (53)
Average of values > 3,000	53,400 (10)	20,600 (25)	123,000 (21)
			[c]47,000 (20)
Average of values ≤ 3,000	450 (57)	310 (42)	1,600 (32)
Area integral[d]	2,300	1,800	4,100

[a] The data here are from a survey in June 19, 1967, while the data in table 1 include a December 1966 survey.
[b] Value in parenthesis is the number of survey readings.
[c] Average value with the maximum level excluded.
[d] The area integral was determined by drawing contamination contours on the facility floor plan and summing the product of contamination level times fractional area.

resulted in internal exposure exceeding the maximum permissible body burden.

Facility C, which had personnel with detectable radium body burdens, also had the highest levels of contamination. These individuals had painted with radium luminous compounds in the past and thus the source of internal body burden is uncertain. However, it appears that alpha contamination levels exceeding 200,000 dpm/100 cm² total-maximum and 50,000 dpm/100 cm² total-average are unsatisfactory and may lead to unnecessary body burdens. Therefore, it appears appropriate to arbitrarily take $\frac{1}{10}$ of these alpha values or

25,000 dpm/100 cm² total-maximum
5,000 dpm/100 cm² total-average
500 dpm/100 cm² removable-average (comparable value)

to be acceptable contamination guides.

These levels are more conservative than those recommended by Duggan and Godfrey (*4*) and comparable to those recommended by Schmidt (*10*). The Kansas State Department of Health has adopted "working levels of contamination in any aircraft instrument repair facility" of 10,000 dpm/100 cm² maximum alpha, 2,500 dpm/100 cm² average alpha and 500 dpm/100 cm² of removable alpha. It is felt that the levels recommended here are attainable and appear to be appropriate for contamination control purposes.

The "area integral" contamination level, as derived above, may be expected to show a better correlation with internal exposure than maximum and average values. A recommended "area integral" limit has not been made because of the experimental nature of this concept and our limited experience in using this parameter.

Industry limitations

The aircraft instrument repair facilities surveyed have not shown the ability to decontaminate their facilities and establish effective control programs. An initial survey by the Kansas State Department of Health revealed extensive levels of contamination, and the facilities were directed to decontaminate and to initiate adequate radiation control programs. The survey 6 months later indicated no significant decrease in contamination levels. Most of the facilities surveyed had no instrumentation for detection of radiation. Only two had any alpha instrumentation, and only three had beta-gamma instruments.

In general, personnel working in these facilities have not received even basic training in radiological safety or contamination control.

The disposal of waste from the dial stripping operation is also a problem. Until recently, the waste was burned in a hood, dumped on the ground outside, released to the sewer, or retained in a hold tank and subsequently taken to the sewage treatment plant. Personnel at these facilities are not familiar with the Department of Transportation regulations pertaining to shipment of radioactive waste. Most facilities need assistance in packaging, monitoring, labeling, and disposing of their radium waste.

Both surveys revealed that radium or radioactive luminous paint was not being currently used to refinish aircraft instrument dials (except for infrequent individual requests). The problems of contamination are thus associated only with the surplus dials which are being reconditioned and existing dials on operating aircraft. The survey in the Miami area indicated that none of the commercial airlines serving that area are now using radioactive luminous dials and that most airlines had ceased this practice several years ago. The dials are now being repainted with fluorescent and phosphorescent finishes and similar compounds which are activated by ultraviolet or black light and contain nonradioactive materials. Matte white is also used as a nonluminous finish.

Moreover, the problem may be a diminishing one in that most of the instruments that are worth rehabilitating have already had the radium dial replaced with a non-radium dial. Approximately 5 percent of the aircraft instruments in facilities A and C contained radium dials. In many cases, the dial stripping operation can be eliminated entirely without any serious economic impact. However, these facilities are still faced with the problem of decontamination of their facility. In some cases, they can accomplish this decontamination within their own capabilities, provided they obtain suitable instrumentation.

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Conclusions

1. Aircraft instrument repair facilities should obtain radiation protection services on a consultant basis if such competence is not available within the company. Such services should be considered mandatory for facilities that strip surplus radioactive dials.

2. Total alpha contamination in aircraft instrument repair facilities should be restricted to levels less than 25,000 dpm/100 cm^2 maximum total 5,000 dpm/100 cm^2 average total and 500 dpm/100 cm^2 removable.

Surface contamination limits should be applied with professional judgment and used only as guides for indicating good housekeeping and general adequacy of the radiation control program. Compliance with contamination guides should not be used as evidence that exposure of persons to internal sources of radiation is within the prescribed standards. Biological sampling or whole body counting should be used to ascertain that the internal exposures are acceptable. It is recommended that contamination levels, especially removable activity, be kept as low as practicable.

3. Dial faces containing radium and aircraft instruments containing radium dials should be segregated using a beta-gamma survey meter from those that do not contain radium. Visual segregation of radium luminous dials is not accurate nor acceptable.

4. The stripping of dials containing radium paint should be discontinued in all facilities where suitable alternatives can be found. If dial stripping must be continued, the procedure recommended by Geiger (*11*) should be used.

5. Instrument disassembly should be done on a plastic sheet or bench paper which can be rolled up and discarded when it becomes contaminated.

6. Urinalysis should initially be done monthly on persons engaged in dial stripping operations involving radioactive material. If three consecutive monthly samples are negative, the collection frequency may be reduced to not less than once every year. If urinalysis provides a positive indication of radium in the body and this result is confirmed by a resample, the person should be whole body counted.

7. Air samples should be collected in areas where instruments are stored, disassembled, or dials are stripped, unless data over a period of several weeks indicate that concentrations in these areas are unlikely to exceed 10 percent of the maximum permissible concentration for occupational exposure.

8. Radium waste should be confined to as small a volume as practical. This waste should be packaged in plastic-lined metal drums. Liquid waste should be evaporated or packed with sufficient absorbent so that all the liquid would be absorbed if the primary container should leak. If an area within the State has been approved by the appropriate regulatory agency for disposal of radium waste, this facility should be used. If such a facility is not available within the State, a commercial waste disposal service should be encouraged to establish a collection and temporary storage facility within the State and provide assistance to each facility to assure compliance with the packaging, labeling, and shipping requirements of the Department of Transportation.

9. Other State or local radiological health agencies having aircraft instrument repair facilities should initiate surveys to determine the extent of radiation contamination and personnel exposure.

10. State radiological health agencies may also wish to consider conducting or sponsoring training programs in radiation protection for persons working in aircraft instrument repair facilities.

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Environmental Survey of Uranium Mill Tailings Pile, Mexican Hat, Utah

Robert N. Snelling[1]

At the request of the Navajo Tribal Council through the PHS Division of Indian Health at Window Rock, Ariz., an environmental radiological survey was conducted on the A–Z Minerals Corporation uranium tailings pile, Mexican Hat, Utah, in May 1968. The purpose of the survey was to identify any radiation hazards which might exist and recommend methods for their control. The survey included evaluation of external gamma radiation and airborne and waterborne radioactivity.

The results of the survey indicate that the external radiation levels on the tailings area exceed recommended exposure limits for individuals in the general population. Therefore, the area should not be released for public use in its present state. Action which would permit the release of the area would involve covering the tailings with uncontaminated soil to an extent that would diminish the external radiation to an acceptable level and to stabilize the covering against wind erosion.

Radiation levels in air and water do not exceed recommended exposure limits. However, to minimize the possibility of increased activity from weather conditions different from those existing during the survey, it is recommended that the tailings be stabilized against wind erosion or periodic monitoring will be necessary.

At the request of the Director, Division of Indian Health, the Bureau of Radiological Health conducted a study of the A-Z Minerals Corporation's tailings pile located at Mexican Hat, Utah. It was requested that the study identify any radiation hazards which might exist and recommend methods for their control. In response to this request, the tailings area was surveyed in May 1968.

The mill was originally operated by the Texas-Zinc Minerals Corporation and transferred to A-Z Minerals Corporation in 1963. Uranium production was discontinued in February 1965. The mill property is located approximately 1 mile southwest of the community of Mexican Hat, Utah (figure 1). The site is situated on the Navajo Indian Reservation and is leased to the company by the Navajo Tribal Council.

The San Juan River flows from east to west approximately 1 mile north of the mill property. Gypsum Creek (actually a dry wash) is located approximately ½ mile east of the tailings pile. The surrounding land is unpopulated and at present is not used for either farming or grazing.

The mill property is divided into three main areas: the tailings area, the mill area, and the housing area (figure 2). The tailings area covers approximately 35 acres and contains an estimated 2.2×10^6 tons of tailings material.[2] The radium content, estimated from the average uranium assay of mill feed material (assuming radium in equilibrium with uranium), is approximately 1,000 pCi/g, or 2,000 total curies.[2] The mill area includes the mill buildings and former stock pile areas. It is located adjacent to and west of the tailings area. The housing area is located approximately 0.5 miles southwest of the tailings area and consists of approximately 60 dwellings, a school, and various community buildings.

Sampling procedures

External radiation measurements were obtained using portable Geiger survey meters. Gamma (closed window) readings were obtained at the surface and at the 3-foot level. A grid was es-

[1] Mr. Snelling is with the Technical Services Program, Southwestern Radiological Health Laboratory, Bureau of Radiological Health, Las Vegas, Nev.

[2] Division of Raw Materials, USAEC, 1967.

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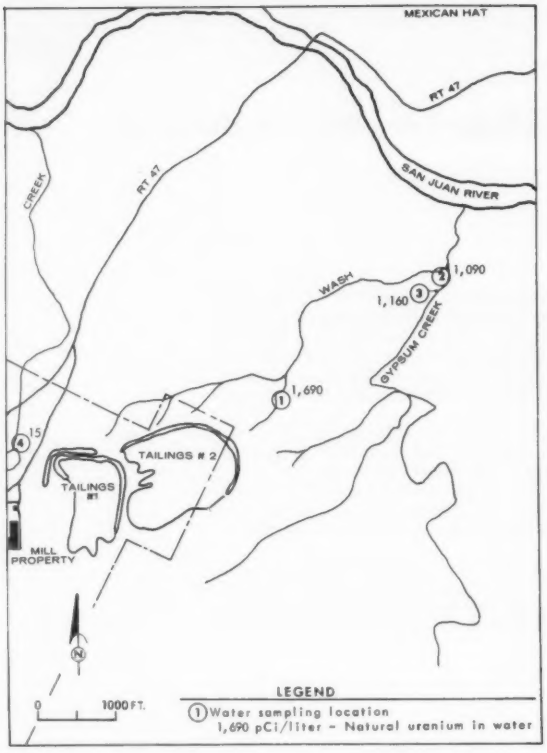

Figure 1. Water sampling locations with natural uranium concentrations Mexican Hat uranium mill, Utah

tablished over the area of interest and readings obtained at 240-foot intervals. A total of 212 pairs of readings was recorded.

In addition to the portable instrument readings, thermoluminescent dosimeters (TLD) were placed in triplicate at the 3-foot level at each of the air sampling stations (figure 3). The TLD's respond only to gamma radiation.

Long-lived airborne particulate radioactivity was sampled by drawing air through a 0.8μ pore size filter. A high-volume air sampler was used at a flow rate of about 5-cubic feet per minute. Nine sampling stations were established on and about the tailings area (figure 2). Continuous 24-hour samples were obtained at each station on each of 11 consecutive days. Each individual filter was analyzed for gross alpha radioactivity. In addition, all filters from each station were composited for analysis of natural uranium, radium-226, and thorium-230.

Eight locations were selected for the collection of radon gas samples (figure 3). The sampling technique consisted of pulling filtered air through a small low-volume pump (10 ml/min) into a 30-liter Mylar bag (*1*). A continuous 24-hour sample was collected at each location. The collected samples were picked up by Public Health Service aircraft and flown to Las Vegas for immediate analysis for radon-222.

The short-lived particulate radon daughter products were sampled by pulling air through a glass-fiber filter using a high-volume air sampler. Each sample was collected for 15 minutes at a flow rate of about 10 cfm. Samples were collected at the radon sampling locations (figure 4). The filters were counted for alpha radioactivity immediately in the field utilizing a specially designed alpha scintillation counter. Activity was calculated using the Tsivoglou technique (*2*).

Wind direction and speed were estimated by visual observation. The predominant wind direction at the mill site for the 11-day sampling period was from the southwest. According to A-Z Minerals Corporation personnel, this is the predominant wind direction throughout the year. During the first 7 days of sampling, the wind was light (0–5 mph). During the last 4 days, however, winds estimated at 10–20 mph were encountered during the daylight hours. In general, lapse conditions were encountered during daylight hours with light inversions at night.

Two springs were located in the vicinity of the tailings area at which ground water samples were taken (figure 1). The first was located about 1,000 feet northeast of the tailings area in the wash running from the tailings area (station 1). The wash was dry between the spring and Gypsum Creek. The second spring was just south of the junction of Gypsum Creek and the wash from the tailings area (stations 2 and 3). Gypsum Creek was dry upstream from the spring. In addition to the ground water samples, the drainage from the acid blowdown operation at the mill was sampled (station 4). One-gallon grab samples were obtained at each location, some stations being sampled on several occasions. The samples were analyzed for radium-226 and natural uranium activity.

Samples of tailings material were collected at the locations indicated in figure 4. Approximately 400 cm^3 of surface material was taken from each

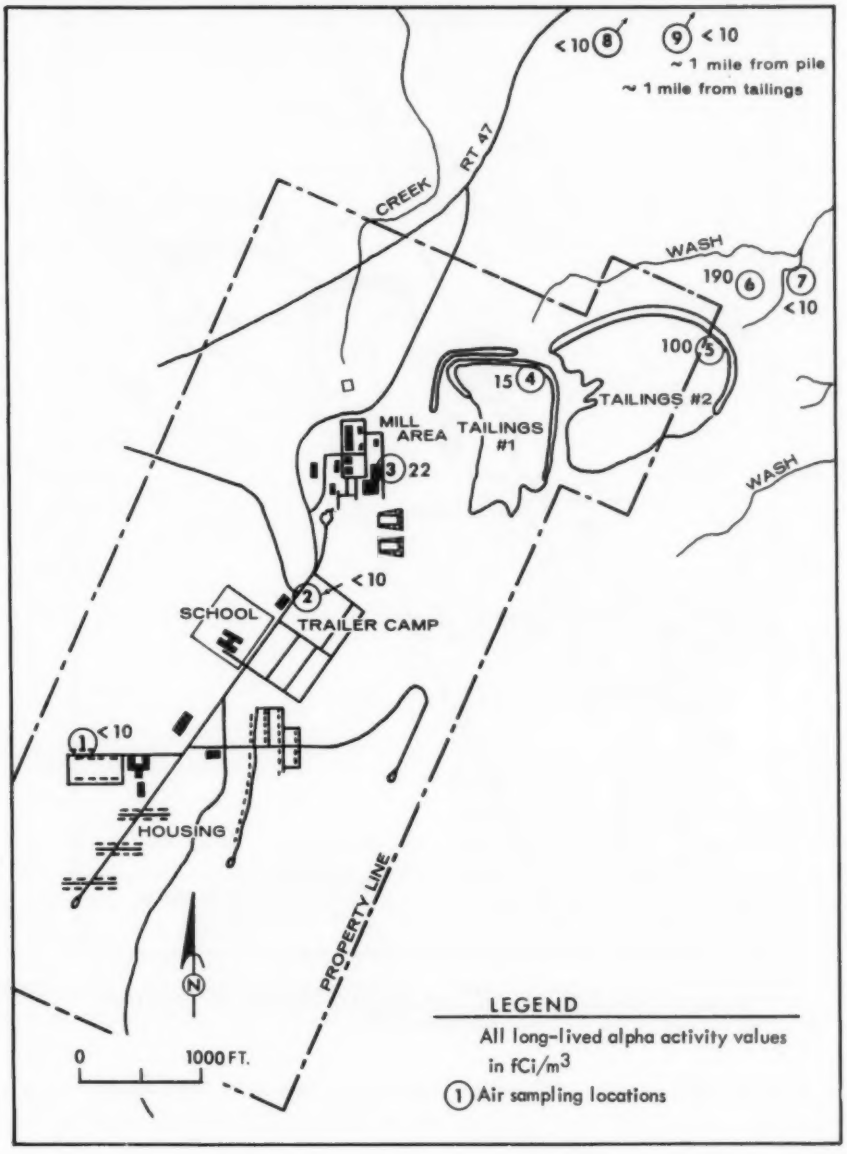

Figure 2. Air sampling locations, Mexican Hat uranium mill

location. The samples were analyzed for thorium-230, radium-226, and natural uranium.

All laboratory analyses were performed at the Southwestern Radiological Health Laboratory. After sample preparation, gross alpha and gross beta radioactivity determinations were performed by counting directly in an end-window gas flow, proportional counter. Radon analysis was per-

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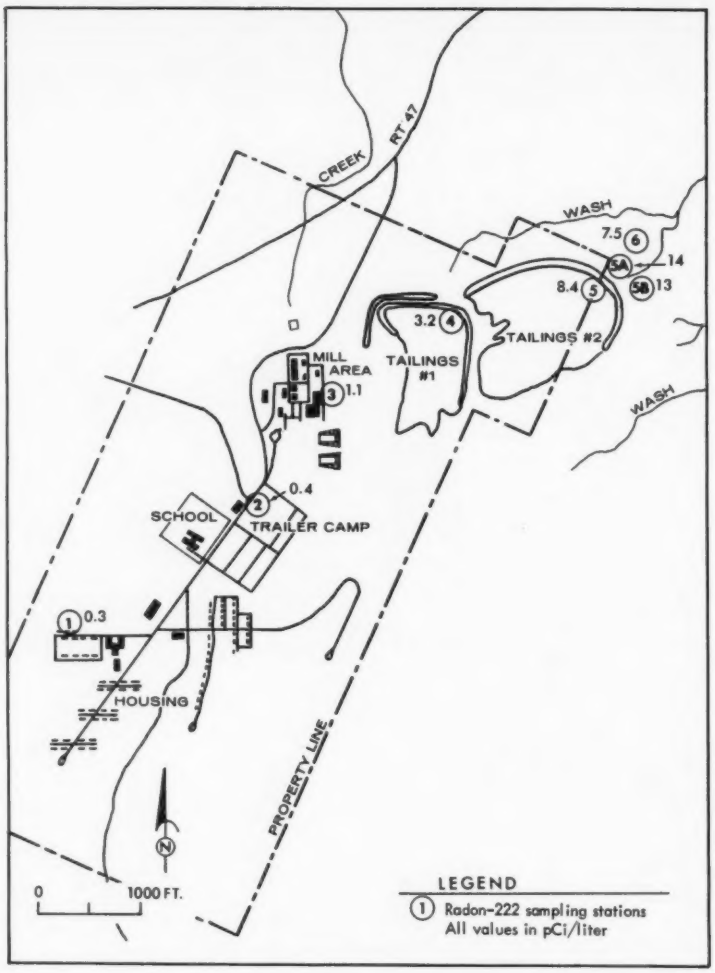

Figure 3. Atmospheric radon-222 sampling stations, Mexican Hat uranium mill

formed by cryogenic separation and subsequent alpha scintillation counting. Radium-226 was determined after sample preparation by the radon emanation technique. Uranium analysis was accomplished by the fluorometric technique, and thorium-230 analysis was performed by means of solvent extraction and subsequent alpha-particle counting.

Results

Table 1 summarizes the results of the portable instrument survey on each of the unique areas. Figures 5 and 6 indicate the location and magnitude of the individual measurements at the surface level and at 3 feet. The natural background radiation was found to average 0.03 mR/h both at

Table 1. External radiation measurements, Mexican Hat uranium mill, May 1968

Area	Number of measurements	Radiation levels at surface (mR/h)		Radiation levels at 3 feet (mR/h)	
		Average	Range	Average	Range
Background	10	0.03	0.02−0.06	0.03	0.02−0.06
Housing	74	.04	.01− .2	.03	.01− .1
Mill	31	.17	.02− .5	.15	.02− .5
Tailings	97	.5	.01−3.0	.5	.02−3.0

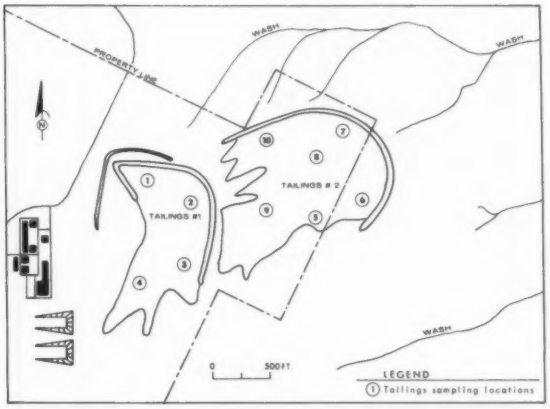

Figure 4. Tailings material sampling locations Mexican Hat uranium mill

Table 2. Comparison of TLD and Geiger meter data May 1968

Station	Exposure at 3 feet (mR/h)	
	TLD[a]	Geiger meter[b]
1	0.01	0.06
2	.03	.05
3	.05	.05
4	.55	.50
5	.42	.40
6	.04	.06
7	.02	.03
8	.01	.02
9	.02	.03
Background 1	.01	.02
Background 2	.01	.03

[a] Eleven day exposure period.
[b] Single survey meter reading.

the surface and the 3-foot level. The radiation exposure at the 3-foot level on the tailings area ranged from a minimum of 0.02 mR/h on the extremities of the pile to a maximum of 3 mR/h on the pile. The average was 0.5 mR/h. No significant difference was noted between the surface and 3-foot level. The average exposure at the 3-foot level within the mill area was 0.15 mR/h with a range of 0.02 to 0.5 mR/h. Again, there was no significant difference between readings at the surface and at 3 feet. The average exposure at the 3-foot level in the housing area was found to average 0.03 mR/h. This exposure represents background radiation. A comparison of the results obtained from the TLD's and the portable instrument measurements is shown in table 2. The results are in good agreement.

Table 3. Gross alpha radioactivity in airborne particulates, Mexican Hat uranium mill

Station number	Gross alpha radioactivity (fCi/m³)											
	May 1968				June 1968							
	28	29	30	31	1	2	3	4	5	6	7	Average
1	12	[a] <10	<10	<10	<10	<10	<10	<10	10	<10	<10	<10
2	19	<10	<10	<10	<10	<10	<10	27	<10	<10	50	<10
3	28	11	15	19	14	22	10	57	40	40	18	22
4	NS	<10	9	41	12	<10	15	<10	10	20	61	15
5	270	NS	NS	78	100	46	NS	270	NS	51	66	100
6	<10	15	84	240	36	24	48	380	510	250	370	190
7	<10	<10	<10	24	<10	NS	NS	NS	NS	NS	NS	<10
8	<10	<10	<10	<10	<10	<10	<10	<10	<10	<10	<10	<10
9	<10	<10	<10	<10	<10	<10	<10	<10	12	11	12	<10

[a] Assumed equal to zero for averaging.
NS, no sample.

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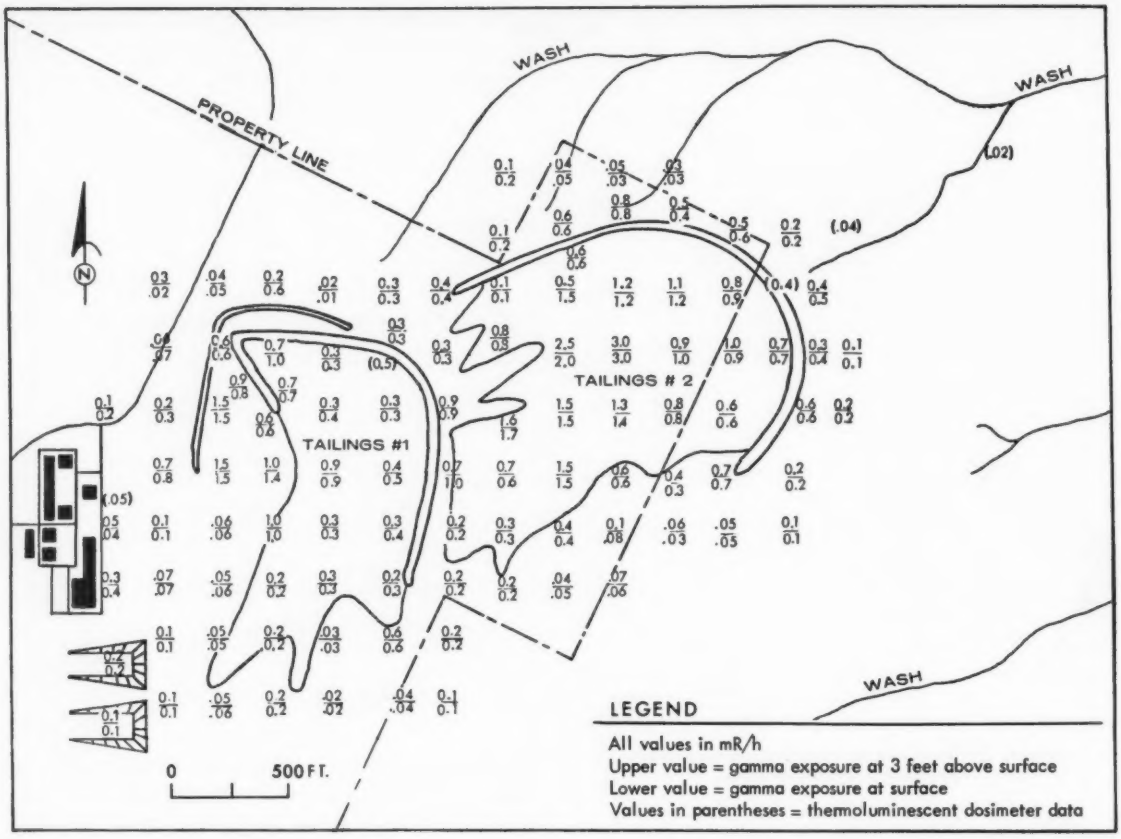

Figure 5. External radiation measurements, Mexican Hat uranium mill tailings piles

The gross alpha radioactivity found on individual filters is shown in table 3. The background station (station 1) averaged less than 10 fCi/m³ for 11 days of sampling. Stations 2, 7, 8, and 9 also averaged less than 10 fCi/m³ indicating background levels of radioactivity. Stations 3 and 4 averaged slightly above background (station 3, 22 fCi/m³; station 4, 15 fCi/m³). Stations 5 and 6 were significantly above background levels. Station 5, located on the northeast edge of the tailings pile, averaged 100 fCi/m³ with a range of 46 to 270 fCi/m³ for 7 days of sampling. Station 6, located approximately 500 feet northeast from the tailings, averaged 190 fCi/m³ with a range from less than 10 fCi/m³ to 510 fCi/m³ for 11 days of sampling. The average station results are shown in figure 9. It is apparent that the activity concentrations are higher than background in the vicinity of the tailings pile and that the values downwind from the tailings are higher than those upwind.

Table 4 shows the average gross alpha radioactivity along with radium-226, thorium-230, and natural uranium activity for each station. It should be noted that stations 3–6 (those indicating positive gross alpha radioactivity results) display radium concentrations averaging about a factor of ten higher than the other stations. These same stations show significant concentrations of thorium-230, whereas all other stations do not. Although the natural uranium concentrations average slightly higher for stations 3–6, the increase does not appear to be significant with the exception of station 3.

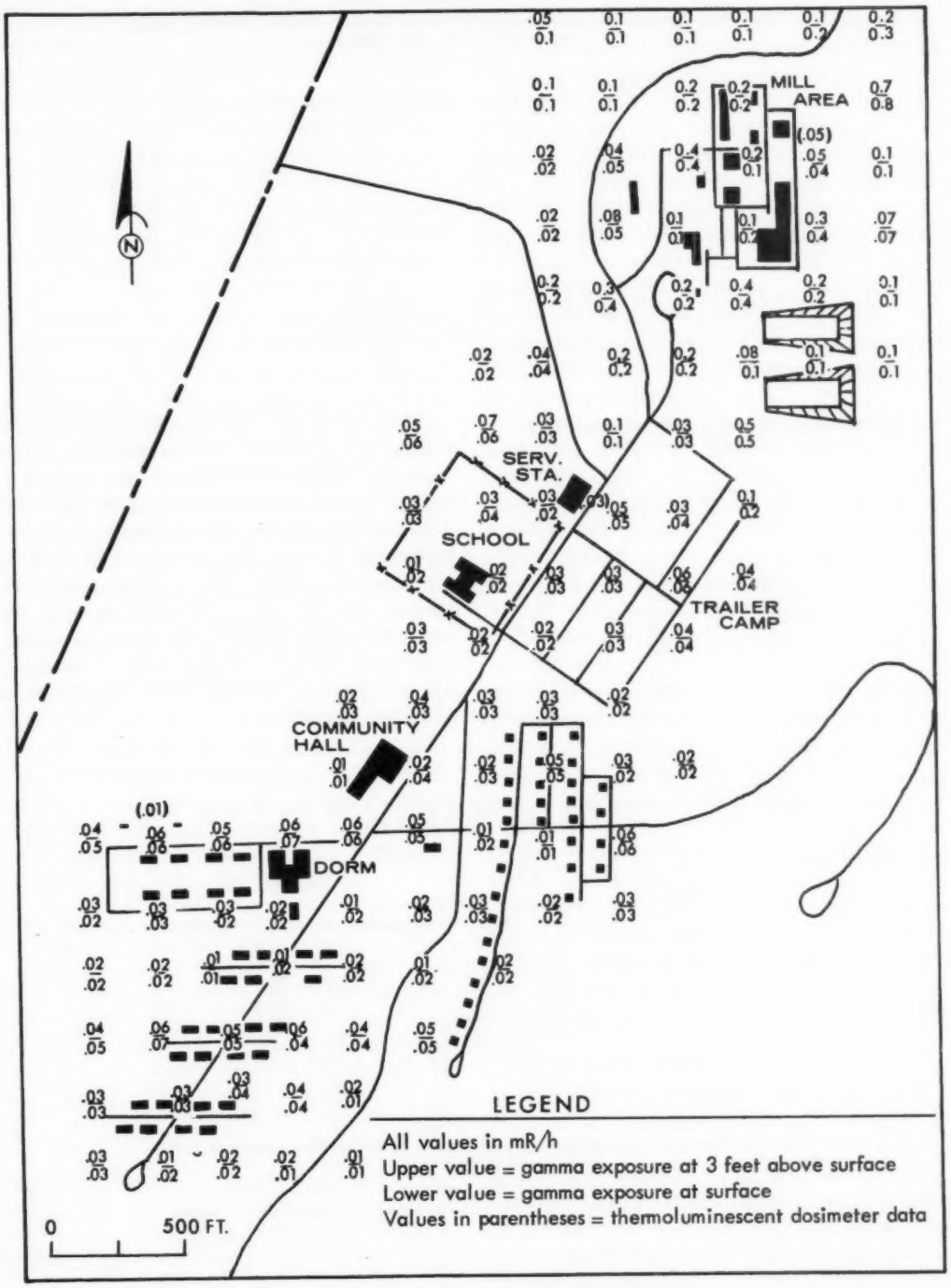

Figure 6. External radiation measurements, Mexican Hat uranium mill

Table 4. Long-lived airborne particulate radioactivity, Mexican Hat uranium mill, May 1968

Station number	Days sampled	Gross alpha radioactivity (fCi/m³)	Radium-226 [a] (fCi/m³)	Thorium-230 [b] (fCi/m³)	Natural uranium [c] (ng/m³)
1	11	<10	0.2	ND	0.2
2	11	<10	.2	ND	1
3	11	22	2	7	10
4	10	15	1	3	2
5	7	100	5	7	3
6	11	190	7	14	4
7	5	<10	.7	ND	.6
8	11	<10	.2	ND	1
9	11	<10	.3	ND	1

[a] Minimum detectable activity for radium-226 — 0.1 pCi/sample.
[b] Minimum detectable activity for thorium-230 — 5.0 pCi/sample.
[c] Minimum detectable activity for natural uranium — 0.1 μg/sample.
ND, nondetectable on composited sample.

Table 5. Atmospheric radon-222, Mexican Hat uranium mill, May–June 1968

Station number	Midpoint of sampling		Radon-222 (pCi/liter)
	Date	Hour	
1	5/27	2030	0.3
2	5/28	2300	.4
3	5/29	2100	1.6
3	6/6	0030	.6
4	5/30	2330	5.7
4	6/4	2400	.7
5	5/31	2100	8.4
5A	6/3	2300	14.0
5B	6/2	2300	13.0
6	6/1	1930	7.5

Table 5 summarizes the radon-222 activity concentrations found at various sampling times and locations. Figure 3 represents the average radon concentrations at each sampling location. Station 1 (a background station) showed a radon concentration of 0.3 pCi/liter. Station 2 (0.4 pCi/liter) can also be considered background level. Those stations on and about the tailings area showed levels of radon significantly above background. Individual samples ranged from a minimum of 0.6 pCi/liter at station 3 to a maximum of 14 pCi/liter at station 5A.

Table 6. Particulate radon daughter products in air, Mexican Hat uranium mill, June 1968

Station number	Collection		Concentration (pCi/m³)		
	Date	Time	Polonium-218	Lead-214	Bismuth-214
5B	6/2	2000	120	90	80
5B	6/3	1100	210	50	10
5A	6/3	1400	10	30	30
5	6/3	1900	0	10	100
5	6/4	1045	0	120	160
4	6/4	1430	0	3	60
3	6/5	1400	220	40	0

Table 7. Long-lived radioactivity in water, Mexican Hat uranium mill, May 1968

Station number	Location	Number of samples	Radium-226 (pCi/liter)	Natural uranium [a] (μg/liter)
1	Wash (spring)	3	1.5	1,690
2	Gypsum Creek	2	.2	1,090
3	Gypsum Creek (spring)	1	.2	1,160
4	Acid blowdown	2	.5	15
Background	Monument Valley	5	.5	2

[a] 1 μg/liter of natural uranium equals 0.7 pCi/liter of alpha radioactivity.

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Table 6 summarizes the results of the radon progeny evaluation. It should be noted that the levels of activity are several orders of magnitude lower than the radon concentrations given in table 5. The apparent disequilibrium between polonium-218 and its daughter products may be due in part to poor counting statistics at the low levels of activity counted. It must be assumed, therefore, that a very low state of equilibrium existed between the radon and its progeny during the survey period.

Table 7 shows the average radium-226 and natural uranium in water results for each sampling location. Because there were no suitable ground water sources in the immediate area, ground water samples from the Monument Valley area (approximately 22 miles distant) are used as background indicators. The uranium levels in the ground water samples were significantly above background, averaging between 1,090 μg/liter and 1,690 μg/liter. The acid blowdown water showed a natural uranium concentration of 15 μg/liter which is only slightly above background. Figure 1 shows the average natural uranium concentration for each location.

Table 8 shows the results of the thorium-230, radium-226, and natural uranium analysis on tailings material. The radium content averaged 370 pCi/g. This average is somewhat below that predicted from the average assay of mill feed material (1,000 pCi/g). Thorium-230 averaged 1,960 pCi/g with individual values ranging from a minimum 220 pCi/g to a maximum of 3,700 pCi/g.

Table 8. Tailings analysis, Mexican Hat uranium mill, May 1968

Area	Number	Uranium (μg/g)	Thorium-230 (pCi/g)	Radium-226 (pCi/g)
Tailings #1	1	63	2,400	340
	2	350	3,000	130
	3	69	3,700	31
	4	27	3,200	27
Tailings #2	5	40	2,800	860
	6	70	220	580
	7	73	220	390
	8	470	2,400	550
	9	120	480	320
	10	180	1,200	470
Average		150	1,960	370

Discussion

The Code of Federal Regulations, Title 10, Part 20 (10 CFR 20), indicates a maximum allowable whole body dose to an individual in an unrestricted area of 0.5 rem in any period of 1 calendar year. The regulations further indicate a maximum allowable dose for continuous exposure to individuals in an unrestricted area of 2 millirems in any 1 hour or 100 millirems in 7 consecutive days.

The Radiation Protection Guide (RPG) adopted by the Federal Radiation Council recommends that the yearly radiation exposure to the whole body of individuals in the population (exclusive of natural background and the deliberate exposure of patients by practitioners of the healing arts) should not exceed 0.5 rem (*3*). The Council further recommends, however, that every reasonable effort should be made to keep exposures as far below this level as possible.

To evaluate the potential long-range hazard associated with the external radiation emitted from the tailings material, the average yearly whole-body exposure for an individual having 24-hour-per-day residence on the area was calculated and compared to the recommended RPG of 0.5 rem/a. The net gamma radiation (i.e., minus background) at 3 feet was used for this computation.

The average exposure at the 3-foot level on the tailings area was found to be 0.5 mR/h (background was assumed negligible). The whole-body average yearly dose for an individual having 24-hour-per-day residence on the tailings area, on the basis of this exposure rate, is approximately 4.4 rem/a. This is nine times the recommended RPG of 0.5 rem/a. These findings serve to preclude the release of the tailings area for public use. Action which would permit release of the area would involve covering the tailings with uncontaminated soil to an extent that would diminish the external radiation to an acceptable level and to stabilize the covering against wind erosion. Otherwise, the area should be fenced and controlled as a radiation area.

The average radiation exposure at the 3-foot level around the mill area was 0.15 mR/h. Subtracting an average background of 0.03 mR/h yields an average net exposure of 0.12 mR/h. The yearly average dose to an individual having 24-hour-per-day residence in this area, on the basis of the average value of 0.12 mR/h, is 1 rem/a. This dose is twice the RPG of 0.5 rem/a but represents the highly unlikely condition of continuous occupancy. With reasonable occupancy assumptions it would not be expected that a yearly average dose in excess of 0.5 rem would occur.

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The radiation exposure at 3 feet in the housing area was found to consist entirely of background radiation.

In order to assess the potential long-range health hazard associated with the inhalation of airborne radionuclides, the average activity concentrations observed for each sampling station are compared to the recommended radiation concentration guides as indicated in 10 CFR 20. Because the guides are applicable to yearly average concentrations, this comparison assumes that the meteorological conditions imposed during the sampling period are representative of the yearly average conditions.

The radiation concentration guide for unidentified natural radionuclides in air, specified in 10 CFR 20, is 80 fCi/m³ (when protactinium-231 is assumed absent). If gross alpha radioactivity is above this guide, identification of the specific radionuclides contributing to the total alpha radioactivity is necessary in order to evaluate the potential long-range hazard associated with the inhalation of the airborne particles.

The nuclides in the natural uranium decay chain most significant as potential health hazards are the alpha-particle emitters. These nuclides and their appropriate guides are shown below:

Thorium-230 (soluble)	80 fCi/m³
Radium-226 (insoluble)	2 pCi/m³
Natural uranium (insoluble)	2 pCi/m³
Lead-210 (soluble)	4 pCi/m³
Polonium-210 (insoluble)	7 pCi/m³

It is apparent from comparison of these values with table 4 that none of those stations showing elevated gross alpha concentrations exceeds the guides for individual radionuclides. The following conclusions can be made; the tailings pile is resulting in elevated concentrations of radioactivity in air in the downwind direction (at the present time, however, these concentrations cannot be shown to exceed the recommended guides for individual isotopes), and it is impossible to predict whether the condition will be worse in the future. At present, the pile is still quite moist and displays a clay-like consistency. As the pile dries out, the condition could very well become worse. On this basis two courses of action are possible. The tailings area may be stabilized against wind erosion. This action would preclude further carriage of tailings material into the surrounding environment in the future and would eliminate the potential for any long-term hazard to anyone inhabiting the area downwind from the mill property. If the tailings area is not stabilized, periodic monitoring will be necessary to insure that the recommended concentration guides are not exceeded.

The Code of Federal Regulations, Title 10, Part 20, indicates a radiation concentration guide for radon-222 in air in unrestricted areas of 3 pCi/liter above natural background. The daughters of radon-222 are assumed to be present to the extent that they exist in unfiltered air (this is interpreted as meaning equilibrium conditions). To evaluate the potential long-range hazard associated with the inhalation of radon and its progeny, the atmospheric radon concentrations are compared to this guide. Average radon concentrations in the tailings area ranged from a minimum of 3.2 pCi/liter at station 4 to a maximum of 14 pCi/liter at station 5A. These values are equal to or greater than the concentration guide of 3 pCi/liter. It was found, however, that a very low percent equilibrium existed between radon and radon daughters. Since approximately 95 percent of the dose to the lungs due to inhalation of radon and its daughters is due to the particulate daughters, the elevated radon levels cannot be considered as representing violation of the guide. It should be pointed out, however, that if enclosed structures were to be built on the tailings material a substantial increase in both radon and daughters could be expected within the structure; and the radiation concentration guide may be exceeded. Therefore, it is recommended that no enclosed structures be built on the tailings material. Station 3 indicated an average radon concentration of 1.1 pCi/liter. Subtracting natural background gives a net average concentration of 0.8 pCi/liter. This is well below the appropriate concentration guide. The housing area (stations 1, 2) showed essentially background levels of radon-222.

Ground water samples obtained in the vicinity of the tailings area showed natural uranium concentrations in excess of 1 nCi/liter. These values are significantly above background. The Code of Federal Regulations, Title 10, Part 20, indicates a radiation concentration guide for natural

uranium of 20 nCi/liter. It is apparent that the concentrations observed are a factor of ten below the recommended guide.

The concentration guide for soluble radium-226 is 30 pCi/liter. None of the samples collected exceeded this guide. It must be concluded, therefore, that although ground water samples in the vicinity of the tailings showed uranium concentrations significantly above background, they are still well below the recommended concentration guides.

Summary

External gamma radiation on the tailings area averaged 0.5 mR/h at 3 feet above the surface. By calculation this comes to a yearly dose of 4.4 rem/a for continuous exposure and exceeds by a factor of nine the recommended Radiation Protection Guide (RPG) of 0.5 rem/a for continuous whole body exposure to individuals in the general population. These findings serve to preclude the release of the tailings area for public use. Action which would permit release of said area would be to cover the tailings with uncontaminated soil to an extent that would diminish the external radiation to an acceptable level and to stabilize the covering against wind erosion. Otherwise, the area should be fenced and controlled as a radiation area.

The tailings pile extends approximately 300 to 500 feet beyond the present property line in the easterly direction. The size of the pile precludes the possibility of returning this material to the property area and therefore this area should be fenced and controlled as a radiation area.

The mill area exhibited external gamma radiation levels averaging 0.12 mR/h above background at 3 feet above the surface. By calculation this comes to a yearly average dose of 1.0 rem for continuous exposure which is two times the appropriate RPG of 0.5 rem/a. With reasonable occupancy assumptions, however, the yearly average dose would probably not exceed 0.5 rem.

The housing area displayed background levels of external radiation.

Airborne transport of tailings material in the northeast direction from the pile is demonstrated by higher than background levels of radioactivity in air. None of the stations showing elevated gross alpha concentrations exceeded the concentration guides for radium-226, thorium-230, or natural uranium, however. Since it is impossible to predict whether the condition will become worse in the future with the drying out of the tailings pile, it is recommended that the area be stabilized against wind erosion. If the area is not stabilized, periodic monitoring will be necessary to insure that the appropriate concentration guides are not exceeded.

The gross alpha radioactivity, radium-226, thorium-230, and natural uranium concentrations in the mill and housing areas were well below the applicable radiation guides.

Average radon gas concentrations on and about the tailings area ranged from 3.2 pCi/liter to 14 pCi/liter. These levels exceed the concentration guide of 3 pCi/liter. It should be pointed out, however, that this guide is applicable to radon-222 and its daughters "as they occur in unfiltered air." This may be interpreted as meaning radon in equilibrium with its daughter products in air. Since it was demonstrated that equilibrium did not exist between radon and its progeny, at least at the time of daughter product sampling, it does not appear that the concentration guide has been exceeded. Levels of radon-222 in the mill area and housing area are substantially below the recommended guide.

It should be emphasized that if enclosed structures were to be built on the tailings material, a substantial increase in radon concentration would occur within the structure and the concentration guide may be exceeded.

Ground water samples obtained in the vicinity of the tailings area showed natural uranium concentrations in excess of 1 nCi/liter. These values represent concentrations substantially above background but are well below the appropriate concentration guide.

Recommendations

On the basis of the data gathered in this survey, it is recommended that the mill tailings be properly stabilized against wind erosion. This action would preclude further transport of tailings material into the surrounding environment and would minimize the potential long-term hazard to anyone inhabiting the area downwind from the mill property. If the tailings area is not stabilized, periodic surveillance will be necessary to insure that significant wind carriage does not occur.

As a result of the external radiation levels on

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the tailings area itself, this area should not be released for public use in its present state. Action which would permit release of the area would be to cover the tailings with uncontaminated soil to an extent that would diminish the external radiation to an acceptable level and to stabilize the covering against wind erosion.

Acknowledgments

Sincere appreciation is expressed to the A-Z Minerals Corporation for the cooperation and assistance provided during the survey.

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SECTION I. MILK AND FOOD

Milk Surveillance, September 1970

Although milk is only one of the sources of dietary intake of environmental radioactivity, it is the food item that is most useful as an indicator of the general population's intake of radionuclide contaminants resulting from environmental releases. Fresh milk is consumed by a large segment of the population and contains several of the biologically important radionuclides that may be released to the environment from nuclear activities. In addition, milk is produced and consumed on a regular basis, is convenient to handle and analyze, and samples representative of general population consumption can be readily obtained. Therefore, milk sampling networks have been found to be an effective mechanism for obtaining information on current radionuclide concentrations and long-term trends. From such information, public health agencies can determine the need for further investigation or corrective public health action.

The Pasteurized Milk Network (PMN) sponsored by the Radiation Office, Environmental Protection Agency, and the Office of Compliance, Food and Drug Administration, U.S. Public Health Service, consists of 63 sampling stations; 61 located in the United States, one in Puerto Rico, and one in the Canal Zone. Many of the State health departments also conduct local milk surveillance programs which provide more comprehensive coverage within the individual State. Data from 15 of these State networks are reported routinely in *Radiological Health Data and Reports*. Additional networks for the routine surveillance of radioactivity in milk in the Western Hemisphere and their sponsoring organizations are:

Pan American Milk Sampling Program (Pan American Health Organization and U.S. Environmental Protection Agency)—5 sampling stations

Canadian Milk Network (Radiation Protection Division, Canadian Department of National Health and Welfare)—16 sampling stations

The sampling locations that make up the networks presently reporting in *Radiological Health Data and Reports* are shown in figure 1. Based on the similar purpose for these sampling activities, the present format integrates the complementary data that are routinely obtained by these several milk networks.

Radionuclide and element coverage

Considerable experience has established that relatively few of the many radionuclides that are formed as a result of nuclear fission become incorporated in milk (*1*). Most of the possible radiocontaminants are eliminated by the selective metabolism of the cow, which restricts gastrointestinal uptake and secretion into the milk. The five fission-product radionuclides which commonly occur in milk are strontium-89, strontium-90, iodine-131, cesium-137, and barium-140. A sixth radionuclide, potassium-40, occurs naturally in 0.0118 percent (*2*) abundance of the element potassium, resulting in a specific activity for potassium-40 of 830 pCi/g total potassium.

Two stable elements which are found in milk, calcium and potassium, have been used as a means for assessing the biological behavior of metaboli-

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Figure 1. Milk sampling networks in the Western Hemisphere

cally similar radionuclides (radiostrontium and radiocesium, respectively). The contents of both calcium and potassium in milk have been measured extensively and are relatively constant. Appropriate values and their variations, expressed in terms of 2-standard deviations (2σ), for these elements are 1.16 ± 0.08 g/liter for calcium and 1.51 ± 0.21 g/liter for potassium. These figures are averages of data from the PMN for May 1963–March 1966 (*3*) and were determined for use in general radiological health calculations or discussions.

Accuracy of data from various milk networks

In order to combine data from the international, national, and State networks considered in this report, it was first necessary to determine the accuracy with which each laboratory is making its determinations and the agreement of the measurements among the laboratories. The Analytical Quality Control Service of the Bureau of Radiological Health conducts periodic studies to assess the accuracy of determinations of radionuclides in milk performed by interested public health radiochemical laboratories. The generalized procedure for making such a study has been outlined previously (*4*).

The most recent study was conducted during May–July 1970, with 28 laboratories participating in an experiment on milk samples containing known concentrations of strontium-89, strontium-90, iodine-131, cesium-137, and barium-140 (*5*). Of the 20 laboratories producing data for the networks reporting in *Radiological Health Data and Reports*, 13 participated in the experiment.

The accuracy results of this experiment are shown in table 1. In general, considerable improvement is needed, especially in the accuracy measurements. These possible differences should be kept in mind when considering the integration of data from the various networks.

Development of a common reporting basis

Since the various networks collect and analyze samples differently, a complete understanding of several parameters is useful for interpreting the data. Therefore, the various milk surveillance networks that report regularly were surveyed for information on analytical methods, sampling and analysis frequencies, and estimated analytical errors associated with the data.

In general, radiostrontium is collected by an ion-exchange technique and determined by beta-particle counting in low-background detectors, and the gamma-ray emitters (potassium-40, iodine-131, cesium-137, and barium-140) are determined by gamma-ray spectroscopy of whole milk. Each laboratory has its own modifications and refinements of these basic methodologies. The methods used

Table 1. Distribution of mean results, quality control experiment

Isotope and known concentration	Number of laboratories in each category			
	Acceptable[a]	Warning level[b]	Unacceptable[c]	Total
Strontium-89: High (258 pCi/liter)	7 (44%)	1 (6%)	8 (50%)	16
Low (15 pCi/liter)	11 (69%)	3 (19%)	2 (12%)	16
Strontium-90: Intermediate (79.4 pCi/liter)	13 (57%)	4 (17%)	6 (26%)	23
Low (32.0 pCi/liter)	5 (25%)	4 (20%)	11 (55%)	20
Iodine-131: High (507 pCi/liter)	18 (67%)	2 (7%)	7 (26%)	27
Low (49 pCi/liter)	16 (64%)	3 (12%)	6 (24%)	25
Cesium-137: High (259 pCi/liter)	20 (74%)	3 (11%)	4 (15%)	27
Low (53 pCi/liter)	17 (66%)	5 (19%)	4 (15%)	26
Barium-140: High (302 pCi/liter)	18 (67%)	2 (7%)	7 (26%)	27
Low (33 pCi/liter)	23 (92%)	0	2 (8%)	25

[a] Measured concentration less than or equal to 2σ of the known concentration.
[b] Measured concentration greater than 2σ and less than or equal to 3σ of the known concentration.
[c] Measured concentration greater than 3σ of the known concentration.

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by each of the networks have been referenced in earlier reports appearing in *Radiological Health Data and Reports.*

A previous article (*6*) summarized the criteria used by the State networks in setting up their milk sampling activities and their sample collection procedures as determined during a 1965 survey. This reference and earlier data article for the particular network of interest may be consulted should events require a more definitive analysis of milk production and milk consumption coverage afforded by a specific network.

Many networks collect and analyze samples on a monthly basis. Some collect samples more frequently but composite the several samples for one analysis, while others carry out their analyses more often than once a month. Many networks are analyzing composite samples on a quarterly basis for certain nuclides. The frequency of collection and analysis varies not only among the networks but also at different stations within some of the networks. In addition, the frequency of collection and analysis is a function of current environmental levels. The number of samples analyzed at a particular sampling station under current conditions is reflected in the data presentation. Current levels for strontium-90 and cesium-137 are relatively stable over short time periods, and sampling frequency is not critical. For the short-lived radionuclides, particularly iodine-131, the frequency of analysis is critical and is generally increased at the first measurement or recognition of a new influx of this radionuclide.

The data presentation also reflects whether raw or pasteurized milk was collected. An analysis of raw and pasteurized milk samples collected during January 1964 to June 1966 indicated that for relatively similar milkshed or sampling areas, the differences in connection of radionuclides in raw and pasteurized milk are not statistically significant (*7*). Particular attention was paid to strontium-90 and cesium-137 in that analysis.

Practical reporting levels were developed by the participating networks, most often based on 2-standard deviation counting errors or 2-standard deviation total analytical errors from replicate analyses (*3*). The practical reporting level reflects analytical factors other than statistical radioactivity counting variations and will be used as a practical basis for reporting data.

The following practical reporting levels have been selected for use by all networks whose practical reporting levels are given as equal or to less than the given value.

Radionuclide	Practical reporting level (pCi/liter)
Strontium-89	5
Strontium-90	2
Iodine-131	10
Cesium-137	10
Barium-140	10

Some of the networks gave practical reporting levels greater than those above. In these cases, the larger value is used so that only data considered by the network as meaningful will be presented. The practical reporting levels apply to the handling of individual sample determinations. The treatment of measurements equal to or below these practical reporting levels for calculation purposes, particularly in calculating monthly averages, is discussed in the data presentation.

Analytical error or precision expressed as pCi/liter or percent in a given concentration range have also been reported by the networks (*3*). The precision errors reported for each of the radionuclides fall in the following ranges:

Radionuclide	Analytical errors of precision (2-standard deviations)
Strontium-89	1–5 pCi/liter for levels <50 pCi/liter; 5–10% for levels ≥ 50 pCi/liter;
Strontium-90	1–2 pCi/liter for levels <20 pCi/liter; 4–10% for levels ≥ 20 pCi/liter;
Iodine-131, Cesium-137, Barium-140	4-10 pCi/liter for levels <100 pCi/liter; 4–10% for levels ≥ 100 pCi/liter.

For iodine-131, cesium-137, and barium-140, there is one exception for these precision error ranges: 25 pCi/liter at levels <100 pCi/liter for Colorado. This is reflected in the practical reporting level for the Colorado milk network.

Federal Radiation Council guidance applicable to milk surveillance

In order to place the U.S. data on radioactivity

Table 2. Concentrations of radionuclides in milk for September 1970 and 12-month period October 1969 through September 1970

Sampling location		Type of sample[a]	Radionuclide concentration (pCi/liter)					
			Strontium–90		Iodine–131		Cesium–137	
			Monthly average[b]	12-month average	Monthly average[b]	12-month average	Monthly average[b]	12-month average
UNITED STATES:								
Ala:	Montgomery[c]	P	NA	6	0	0	12	9
Alaska:	Palmer[c]	P	7	5	0	0	0	6
Ariz:	Phoenix[c]	P	NA	1	0	0	0	0
Ark:	Little Rock[c]	P	12	13	0 (2)	0	15 (2)	17
Calif:	Sacramento[c]	P	NA	1	0	0	0	0
	San Francisco[c]	P	NA	2	0	0	0	0
	Del Norte	P	15	17	0	0	6	19
	Fresno	P	2	2	0	0	0	2
	Humboldt	P	4	5	0	0	0	5
	Los Angeles	P	2	2	0	0	0	2
	Mendocino	P	4	3	0	0	0	3
	Sacramento	P	2	2	0	0	0	2
	San Diego	P	2	2	0	0	0	1
	Santa Clara	P	1	2	0	0	0	1
	Shasta	P	2	3	0	0	0	2
	Sonoma	P	2	3	0	0	0	3
Colo:	Denver[c]	P	NA	5	0	0	0	1
	West	R	(d)		(e)0	(e)0	(e)0	4
	Northeast	R	(d)		(e)0 (2)	(e)0	(e)0 (2)	(e)0
	East	R	(d)		(e)0	(e)0	(e)0	(e)0
	Southeast	R	(d)		NS	(e)0	NS	(e)0
	South central	R	(d)		NS	NS	NS	NS
	Southwest	R	(d)		(e)0	(e)0	(e)0	(e)3
	Northwest	R	(d)		(e)0	(e)0	(e)0	(e)0
Conn:	Hartford[c]	P	NA	7	0	0	13	12
	Central	P	6	7	0	0	15	13
Del:	Wilmington[c]	P	NA	9	0	0	0	6
D.C.:	Washington[c]	P	NA	8	0	0	0	6
Fla:	Tampa[c]	P	5	6	0	0	59	51
	West	R	12	10	0	0	19	22
	North	R	7	11	0	0	0	23
	Northeast	R	6	6	0	0	59	40
	Central	R	NA	7	0	0	20	52
	Tampa Bay area	R	6	6	0	0	57	47
	Southeast	R	6	7	0	0	76	71
Ga:	Atlanta[c]	P	NA	10	0	0	12	16
Hawaii:	Honolulu[c]	P	3	2	0	0	0	0
Idaho:	Idaho Falls[c]	P	6	5	0	0	0	4
Ill:	Chicago[c]	P	7	8	0	0	13	9
Ind:	Indianapolis[c]	P	NA	8	0	0	0	5
	Northeast	P	8	9	0	1	15	14
	Southeast	P	8	11	0	0	10	12
	Central	P	6	9	15	1	10	10
	Southwest	P	10	11	0	0	10	12
	Northwest	P	9	10	0	0	20	15
Iowa:	Des Moines[c]	P	NA	6	0	0	0	2
	Iowa City	P	7	8	0	0	11	10
	Des Moines	P	9	8	0 (5)	0	3	6
	Spencer	P	6	6	0	0	0	5
	Fredericksburg	P	NS	NS	NS	NS	NS	NS
Kans:	Wichita[c]	P	NA	8	0	0	0	1
	Coffeyville	P	7	10	0	0	0	0
	Dodge City	P	9	9	0	0	0	1
	Falls City	P	0	14	0	0	13	2
	Hays	P	12	13	13	1	0	0
	Kansas City	P	15	9	0	0	11	3
	Topeka	P	3	11	0	0	0	2
	Wichita	P	8	10	0	0	0	0
Ky:	Louisville[c]	P	NA	8	0	0	0	4
La:	New Orleans[c]	P	14	15	0	0	19	18
Maine:	Portland[c]	P	NA	10	0	0	13	21
Md:	Baltimore[c]	P	NA	8	0	0	12	8
Mass:	Boston[c]	P	9	10	0	0	23	21
Mich:	Detroit[c]	P	NA	8	0	0	13	10
	Grand Rapids[c]	P	NA	9	0	0	16	12
	Bay City	P	NA	7	(e)0	(e)0	11	10
	Charlevoix	P	10	12	(e)0 (4)	(e)0	15 (4)	16
	Detroit	P	7	8	(e)0	(e)0	0	7
	Grand Rapids	P	9	9	(e)0	(e)0	0	9
	Lansing	P	NA	9	(e)0 (2)	(e)0	10 (2)	15
	Marquette	P	NA	14	(e)0	(e)0	19	22
	Monroe	P	5	6	(e)0 (2)	(e)0	0 (2)	4
	South Haven	P	8	8	(e)0 (4)	(e)0	4 (4)	7
Minn:	Minneapolis[c]	P	NA	10	0	0	15	8
	Bemidji	P	10	10	0	0	26	21
	Mankato	P	7	4	0	0	0	11
	Rochester	P	7	6	0	0	0	11
	Duluth	P	13	14	0	0	20	26
	Worthington	P	3	4	0	0	0	0
	Minneapolis	P	10	9	0	0	14	14
	Fergus Falls	P	7	5	0	0	16	13
	Little Falls	P	15	10	0	0	27	32

See footnotes at end of table.

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Table 2. Concentrations of radionuclides in milk for September 1970 and 12-month period October 1969 through September 1970—Continued

Sampling location		Type of sample[a]	Radionuclide concentration (pCi/liter)					
			Strontium-90		Iodine-131		Cesium-137	
			Monthly average[b]	12-month average	Monthly average[b]	12-month average	Monthly average[b]	12-month average
UNITED STATES—Continued								
Miss:	Jackson[c]	P	NA	12	0	0	0	11
Mo:	Kansas City[c]	P	NA	8	0	0	0	2
	St. Louis[c]	P	NA	8	0	0	0	3
Mont:	Helena[c]	P	NA	5	0	0	23	8
Nebr:	Omaha[c]	P	NA	6	0	0	0	1
Nev:	Las Vegas[c]	P	NA	2	0	0	0	0
N.H:	Manchester[c]	P	NA	8	0	0	23	18
N.J:	Trenton[c]	P	NA	8	0	0	11	9
N.Mex:	Albuquerque[c]	P	NA	3	0	0	0	0
N.Y:	Buffalo[c]	P	8	7	0	0	0	8
	New York City[c]	P	NA	9	0	0	24	16
	Syracuse[c]	P	NA	7	0	0	0	10
	Albany	P	7	5	0 (5)	0	0	0
	Buffalo	P	(e)0	(e)0	0	0	0	0
	Massena	P	11	7	0 (2)	0	20	0
	Newburg	P	6	7	0 (2)	0	0	0
	New York City	P	14	8	0	0	0	0
	Syracuse	P	(e)0	3	0	0	0	0
N.C:	Charlotte[c]	P	NA	11	0	0	22	14
N.Dak:	Minot[c]	P	NA	9	0	0	0	14
Ohio:	Cincinnati[c]	P	NA	8	0	0	0	8
	Cleveland[c]	P	NA	9	0	0	13	8
Okla:	Oklahoma City[c]	P	NA	6	0	0	0	5
	Oklahoma City	P	NS		NS		NS	
	Enid	P	NS		NS		NS	
	Tulsa	P	NS		NS		NS	
	Lawton	P	NS		NS		NS	
	Ardmore	P	NS		NS		NS	
Oreg:	Portland[c]	P	3	6	0	0	0	5
	Baker	P	NA	3	ND	0	18	3
	Coos Bay	P	NA	5	ND	0	24	14
	Eugene	P	NA	3	ND	0	16	4
	Medford	P	NA	2	ND	0	24	10
	Portland composite	P	NA	4	ND	0	19	8
	Portland local	P	NA	4	ND	0	11	6
	Redmond	P	NA	2	ND	0	0	3
	Tillamook	R	NA	5	ND	0	38	16
Pa:	Philadelphia[c]	P	NA	9	0	0	0	5
	Pittsburgh[c]	P	NA	10	0	0	16	11
	Dauphin	P	9	7	0	3	11	14
	Erie	P	0	11	0	3	26	29
	Philadelphia	P	0	9	0	3	20	15
	Pittsburgh	P	10	13	0	2	14	21
R.I:	Providence[c]	P	NA	8	0	0	21	16
S.C:	Charleston[c]	P	7	10	0	1	16	20
S.Dak:	Rapid City[c]	P	6	7	0	0	0	4
Tenn:	Chattanooga[c]	P	NA	9	0	0	0	11
	Memphis[c]	P	NA	8	0	0	14	7
	Chattanooga	P	10	10	0	0	11	13
	Clinton	P	10	12	0	0	0	13
	Fayetteville	P	11	12	0	0	0	10
	Knoxville	P	9	8	0 (2)	0	9 (2)	7
	Nashville	P	5	7	0 (2)	0	0 (2)	6
Tex:	Austin[c]	P	NA	3	0	0	0	3
	Dallas[c]	P	NA	6	0	0	0	7
	Amarillo	R	5	4	0	0	0	2
	Corpus Christi	R	NS	4	NS	0	NS	2
	El Paso	R	NS	3	NS	0	NS	0
	Fort Worth	R	NS	5	NS	0	NS	2
	Harlingen	R	0	3	0	0	0	0
	Houston	R	NS	8	NS	0	NS	11
	Lubbock	R	NS	4	NS	0	NS	0
	Midland	R	NS	3	NS	0	0	0
	San Antonio	R	3	4	0	0	NS	0
	Texarkana	R	NS	8	NS	0	NS	5
	Tyler	R	NS	15	NS	0	NS	17
	Uvalde	R	NS	2	NS	0	NS	0
	Wichita Falls	R	NS	9	NS	0	NS	5
Utah:	Salt Lake City[c]	P	4	5	0 (2)	0	13 (2)	5
Vt:	Burlington[c]	P	NA	8	0	0	19	13
Va:	Norfolk[c]	P	8	9	0	0	0	9
Wash:	Seattle[c]	P	NA	5	0	0	0	2
	Spokane[c]	P	NA	6	0	0	0	2
	Benton County	R	0	0	0	0	0	5
	Franklin County	R	NS	2	NS	0	NS	3
	Sandpoint, Idaho	R	11	10	0	0	23	22
	Skagit County	R	6	6	0	0	0	8
W.Va:	Charleston[c]	P	NA	8	0	0	12	5
Wisc:	Milwaukee[c]	P	NA	6	0	0	13	11
Wyo:	Laramie[c]	P	NA	5	0	0	0	1

See footnotes at end of table.

Table 2. Concentrations of radionuclides in milk for September 1970 and 12-month period October 1969 through September 1970—Continued

Sampling location	Type of sample[a]	Radionuclide concentration (pCi/liter)					
		Strontium–90		Iodine–131		Cesium–137	
		Monthly average[b]	12-month average	Monthly average[b]	12-month average	Monthly average[b]	12-month average
CANADA:							
Alberta: Calgary	P	8	8	[d]		9	17
Edmonton	P	11	8	[d]		22	17
British Columbia: Vancouver	P	9	10	[d]		30	27
Manitoba: Winnipeg	P	8	7	[d]		27	25
New Brunswick: Fredericton	P	15	13	[d]		24	21
Newfoundland: St. Johns	P	30	18	[d]		39	36
Nova Scotia: Halifax	P	12	10	[d]		24	20
Ontario: Thunder Bay	P	12	12	[d]		30	28
Ottawa	P	6	6	[d]		16	14
Sault Ste. Marie	P	16	14	[d]		39	31
Toronto	P	4	4	[d]		8	10
Windsor	P	5	5	[d]		9	10
Quebec: Montreal	P	8	7	[d]		13	18
Quebec	P	8	10	[d]		20	26
Saskatchewan: Regina	P	8	7	[d]		11	14
Saskatoon	P	9	9	[d]		21	13
CENTRAL AND SOUTH AMERICA:							
Colombia: Bogota	P	0	0	0	0	0	0
Chile: Santiago	P	0	0	0	0	0	0
Ecuador: Guayaquil	P	0	0	0	0	0	0
Jamaica: Montego Bay	P	NS	4	NS	0	NS	78
Venezuela: Caracas	P	3	1	0	0	0	0
Canal Zone: Cristobal[e]	P	NA	1	0	0	15	10
Puerto Rico: San Juan[e]	P	0	3	0	0	14	7
PMN Network average[f]		7	7	0	0	8	9

[a] P, pasteurized milk.
R, raw milk.
[b] When an individual sampling result was equal to or less than the practical reporting level, a value of "0" was used for averaging. Monthly averages less than the practical reporting level reflect the fact that some but not all of the individual samples making up the average contained levels greater than the practical reporting level. When more than one analysis was made in a monthly period, the number of samples in the monthly average is given in parentheses.
[c] PHS Pasteurized Milk Network station. All other sampling locations are part of the State or National network.
[d] Radionuclide analysis not routinely performed.
[e] The practical reporting levels for these networks differ from the general ones given in the text. Sampling results for the networks were equal to or less than the following practical reporting levels:
Iodine-131: Colorado–25 pCi/liter; Michigan–14 pCi/liter; Oregon–15 pCi/liter
Cesium-137: Colorado–25 pCi/liter; New York–20 pCi/liter; Oregon–15 pCi/liter
Strontium-90: New York-3 pCi/liter
[f] This entry gives the average radionuclides concentrations for the PHS Pasteurized Milk Network stations denoted by footnote [e].
NA, no analysis.
NS, no sample collected.

in milk presented in *Radiological Health Data and Reports* in perspective, a summary of the guidance provided by the Federal Radiation Council for specific environmental conditions was presented in the December 1970 issue of *Radiological Health Data and Reports.*

Data reporting format

Table 2 presents the integrated results of the international, national, and State networks discussed earlier. Column 1 lists all the stations which are routinely reported to *Radiological Health Data and Reports.* The relationship between the PMN stations and the State stations is shown in figure 2. The first column under each of the radionuclides reported gives the monthly average for the station and the number of samples analyzed in that month in parentheses. When an individual sampling result is equal to or below the practical reporting level for the radionuclide, a value of zero is used for averaging. Monthly averages are calculated using the above convention. Averages which are equal to or less than the practical reporting levels reflect the presence of radioactivity in some of the individual samples greater than the practical reporting level.

The second column under each of the radionuclides reported gives the 12-month average for the station as calculated from the preceding 12 monthly averages, giving each monthly average equal weight. Since the daily intake of radio-

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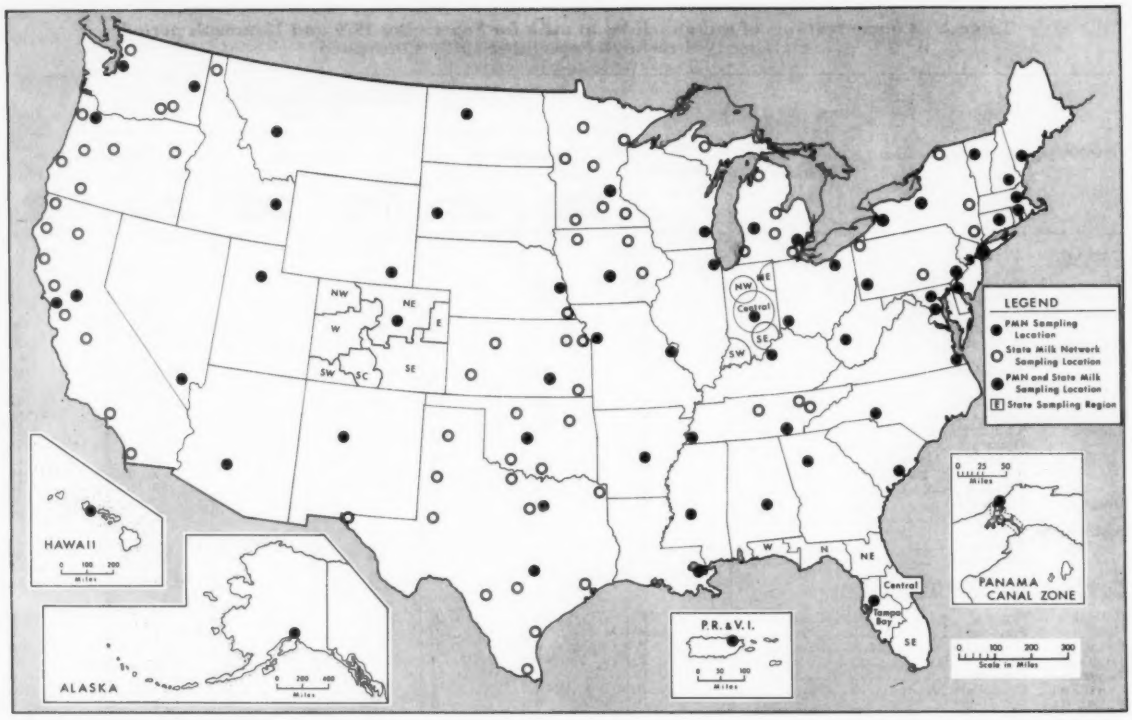

Figure 2. State and PMN milk sampling locations in the United States

activity by exposed populations groups, averaged over a year, constitutes an appropriate criterion for the case where the FRC radiation protection guides apply, the 12-month average serves as a basis for comparison.

Discussion of current data

In table 2, surveillance results are given for strontium-90, iodine-131, and cesium-137 for September 1970 and the 12-month period, October 1969 to September 1970. Except where noted, the monthly average represents a single sample for the sampling station. Strontium-89 and barium-140 data have been omitted from table 2 since levels at the great majority of the stations for September 1970 were below the respective practical reporting levels. Table 3 gives monthly averages for those stations at which strontium-89 or barium-140 were detected.

Iodine-131 results are included in the table, even though they were generally below practical reporting levels. Because of the lower values reflected by the radiation protection guidance provided by the Federal Radiation Council, levels in milk for this radionuclide are of particular public health interest. In general, the practical reporting level for iodine-131 is numerically equal to the upper value of Range I (10 pCi/liter) of the FRC radiation protection guide.

Strontium-90 monthly averages ranged from 0 to 15 pCi/liter in the United States for September 1970 and the highest 12-month average was 17

Table 3. Strontium-89 and barium-140 in milk September 1970

Sampling location		Radionuclide concentration (pCi/liter)	
		Strontium-89	Barium-140
Kans:	Coffeyville (State)	6	
	Dodge City (State)		11
	Falls City (State)		13
	Kansas City (State)		11
	Topeka (State)	10	
Chile:	Santiago	20	
Ecuador:	Guayaquil	10	

pCi/liter (Del Norte, Calif.) representing 8.5 percent of the Federal Radiation Council radiation protection guide. Cesium-137 monthly averages ranged from 0 to 76 pCi/liter in the United States for September 1970, and the highest 12-month average was 71 pCi/liter (Southeast Fla.), representing 2.0 percent of the value derived from Federal Radiation Council recommendations. Of particular interest are the consistently higher cesium-137 levels that have been observed in Florida (*8*) and Jamaica. Iodine-131 results for individual samples were all below the practical reporting level, with the exception of Hays, Kans. (State), 13 pCi/liter, and central Indiana (State), 15 pCi/liter, for September 1970.

The Canadian Department of National Health and Welfare have made recent changes in the frequency of analyses of samples from the network. Samples are collected on a weekly basis and shipped to Ottawa where they are further composited to make a monthly sample for each station. The monthly sample is gamma-scanned and analyzed for cesium-137 and potassium. Aliquots from the monthly samples are composited to make a quarterly sample for each station. The quarterly sample is analyzed for strontium-90 and calcium.

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Bureau of Radiological Health
Division of Environmental Sanitation
California State Department of Health

Radiation Protection Division
Canadian Department of National Health and Welfare

Radiological Health Section
Division of Air, Occupational and Radiation Hygiene
Colorado State Department of Health

Radiological Health Services
Division of Medical Services
Connecticut State Department of Health

Radiological and Occupational Health Section
Department of Health and Rehabilitative Services
State of Florida

Bureau of Environmental Sanitation
Division of Sanitary Engineering
Indiana State Board of Health

Division of Radiological Health
Environmental Engineering Services
Iowa State Department of Health

Radiation Control Section
Environmental Health Division
Kansas State Department of Health

Radiological Health Service
Division of Occupational Health
Michigan Department of Health

Radiation Control Section
Division of Environmental Health
State of Minnesota Department of Health

Bureau of Nuclear Engineering
New York State Department of Environmental Conservation

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Division of Occupational and Radiological Health
Environmental Health Services
Oklahoma State Department of Health

Radiological Health Services
Division of Preventable Diseases
Tennessee Department of Public Health

Environmental Radiation Surveillance Program
Division of Sanitation and Engineering
Oregon State Board of Health

Division of Occupational Health
Environmental Health Services
Texas State Department of Health

Radiological Health Section
Bureau of Environmental Health
Pennsylvania Department of Public Health

Office of Air Quality Control
Division of Technical Services
Washington State Department of Health

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Food and Diet Surveillance

Efforts are being made by various Federal and State agencies to estimate the dietary intake of selected radionuclides on a continuing basis. These estimates, along with the guidance developed by the Federal Radiation Council, provide a basis for evaluating the significance of radioactivity in foods and diet.

Networks presently in operation and reported routinely include those listed below. These networks provide data useful for developing estimates of nationwide dietary intakes of radionuclides. Programs reported in *Radiological Health Data and Reports* are shown below.

Program	Period reported	Issue
California Diet Study	January–June 1970	November 1970
Carbon-14 and Tritium in Total Diet and Milk	January 1969–June 1970	January 1971
Connecticut Standard Diet	July–December 1969	December 1970
Institutional Diet Samples	January–March 1970	November 1970
Strontium-90 in Tri-City Diets	January–December 1969	June 1970

1. Radionuclides in Institutional Diet Samples April-June 1970

Bureau of Radiological Health and
Food and Drug Administration

The determination of radionuclide concentrations in the diet constitutes an important element of an integrated program of environmental radiological surveillance and assessment. In recognition of the potential significance of the diet in contributing to total environmental radiation exposures, the Public Health Service initiated its Institutional Diet Sampling Program in 1961. This program is administered by the Bureau of Radiological Health with the assistance of the Office of Compliance, Food and Drug Administration (*1*).

The program was designed to provide estimates of the dietary intake of radionuclides in a selected population group ranging from children to young adults of school age. At the present time 19 basic institutions and eight auxiliary institutions, distributed geographically as shown in figure 1, are being sampled. Previous results showed that the daily dietary intake of teenage girls and children from 9 to 12 years of age were comparable, while teenage boys consumed 20 percent more food per day (*1*, *2*). Consequently, estimates for teenage boys and/or girls can be calculated on the basis of the dietary intakes of children.

In general, the sampling procedure is the same at each institution. Each sample represents the edible portion of the diet for a full 7-day week (21 meals plus soft drinks, candy bars, or other in-between snacks) obtained by duplicating the meals of a different individual each day. Drinking water, not included in the samples, is also sampled periodically. Each daily sample is kept frozen until the end of the collection period and is then packed in dry ice and shipped by air express to either the Southwestern Radiological Health Laboratory, Las Vegas, Nev; the Southeastern Radiological Health Laboratory, Montgomery, Ala; or the Northeastern Radiological Health Laboratory, Winchester, Mass. A detailed description of sampling and analytical procedures was presented earlier (*3*).

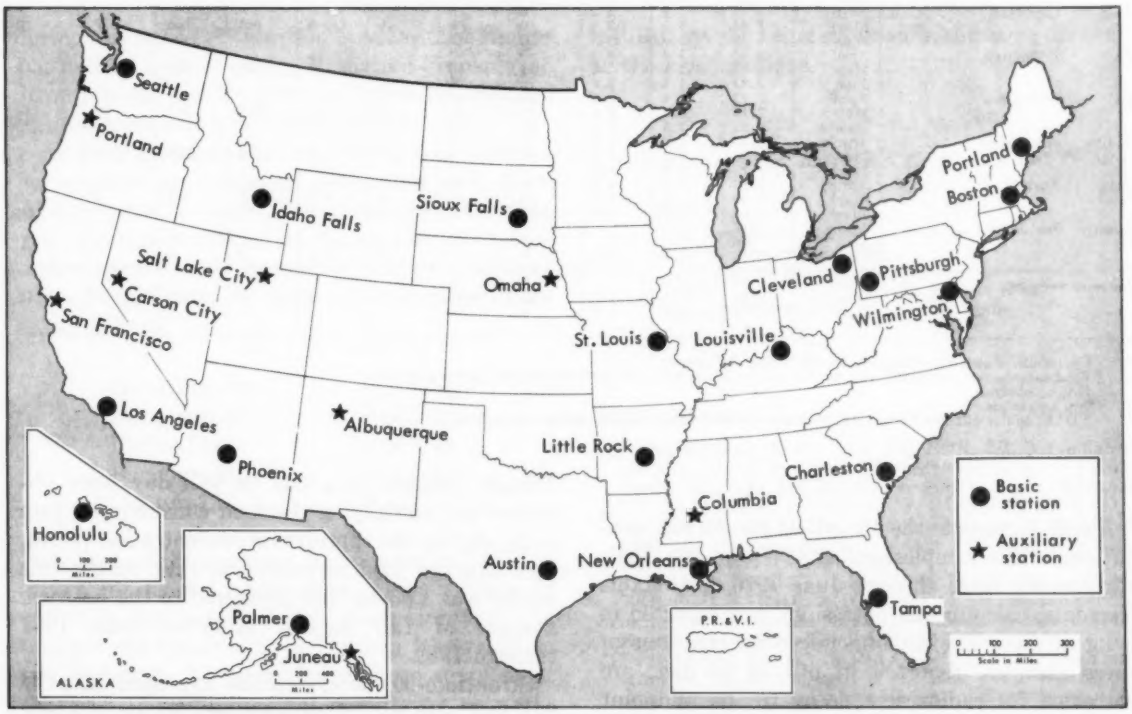

Figure 1. Institutional diet sampling locations as of June 1970

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Table 1. Concentration and intake of stable elements and radionuclides in institutional total diets of children (basic stations), April–June 1970

Location of Institution	Month (1970)	Total weight (kg/day)	Potassium		Cesium–137		Calcium[a]		Strontium–89[a]		Strontium–90[a]	
			(g/kg)	(g/day)	(pCi/kg)	(pCi/day)	(g/kg)	(g/day)	(pCi/kg)	(pCi/day)	(pCi/kg)	(pCi/day)
Alaska: Palmer	April	NS										
	May	NS										
	June	1.05	1.5	1.5	22	23	0.7	0.7	0	0	9	9
Ariz: Phoenix	April	NS										
	May	NS										
	June	1.5	1.4	2.1	0	0	.7	1.1	0	0	9	14
Ark: Little Rock	April[b]	1.12	1.5	1.7	12	13						
	May	1.14	1.5	1.8	22	25	.6	0.7	NA	NA	9	10
	June[b]	1.14	1.7	1.9	14	16						
Calif: Los Angeles	April	NS										
	May	NS										
	June[b]	1.37	1.2	1.6	0	0	.5	0.7	0	0	5	6
Del: Wilmington	April	1.50	1.5	3.1	0	0						
	May	2.28	1.5	3.4	0	0	.6	1.2	NA	NA	6	12
	June	1.86	1.5	2.8	15	28						
Fla: Tampa	April	1.43	1.5	2.0	18	26						
	May[b]	1.39	1.5	2.1	20	28	.7	1.1	NA	NA	8	12
	June	1.83	1.7	3.0	37	68						
Hawaii: Honolulu	April	NS										
	May	NS										
	June[b]	2.12	1.5	3.1	0	0	.5	1.1	0	0	0	0
Idaho: Idaho Falls	April[b]	1.89	1.6	3.0	13	25						
	May[b]	1.53	1.5	2.3	0	0	1.2	2.3	0	0	9	17
	June[b]	1.97	1.7	3.3	12	24						
Ky: Louisville	April	2.52	1.6	4.1	15	38						
	May	2.60	1.9	4.9	18	47	.6	1.5	NA	NA	7	19
	June	2.65	1.9	4.9	18	48						
La: New Orleans	April[b]	1.89	1.6	3.1	0	0						
	May[b]	1.67	1.6	2.7	12	20	.6	1.0	NA	NA	10	17
	June[b]	1.37	1.1	1.6	0	0						
Maine: Portland	April	1.81	1.9	3.4	15	27						
	May	1.81	1.6	2.9	13	24	.9	1.6	NA	NA	11	19
	June	1.75	1.6	2.8	14	25						
Mass: Boston	April	2.22	1.4	3.1	12	27						
	May	2.02	1.8	3.6	18	36	.6	1.2	0	0	7	15
	June	2.08	1.4	2.9	0	0						
Mo: St. Louis	April[b]	2.37	1.7	4.0	14	33						
	May[b]	2.11	1.9	4.0	0	0	1.2	2.7	NA	NA	7	15
	June[b]	2.27	1.9	4.3	13	30						
Ohio: Cleveland	April	1.24	1.8	2.2	0	0						
	May[b]	1.45	1.6	2.3	0	0	.6	0.8	NA	NA	7	10
	June	1.35	1.7	2.3	0	0						
Pa: Pittsburgh	April	2.35	1.4	3.3	0	0						
	May	2.27	1.3	3.0	14	32	.5	1.2	NA	NA	6	13
	June	2.25	1.3	2.9	12	27						
S.C: Charleston	April[b]	2.12	1.4	2.9	20	42						
	May	2.45	1.4	3.5	12	29	.7	1.5	NA	NA	10	23
	June	2.22	1.3	3.0	20	44						
S. Dak: Sioux Falls	April[b]	1.35	1.9	2.6	0	0						
	May[b]	1.20	1.5	1.8	0	0	.8	1.0	NA	NA	5	7
	June[b]	1.20	1.6	1.9	0	0						
Tex: Austin	April[b]	2.26	1.5	3.5	0	0						
	May[b]	1.80	1.6	2.9	16	29	.8	1.5	NA	NA	3	6
	June[b]	1.75	1.6	2.9	0	0						
Wash: Seattle	April	NS										
	May	NS										
	June[b]	2.00	1.3	2.6	0	0	.5	1.1	0	0	7	13
Network average	April	1.86	1.6	3.0	9	17						
	May	1.84	1.6	2.9	10	19	0.7	1.3	0	0	7	12
	June	1.78	1.5	2.7	9	18						

[a] Composite of monthly samples collected during the quarter.
[b] Food samples were collected from two or more children who were not between the ages of 9 and 12.
NA, no analysis.
NS, no sample.
NOTE: Iodine-131 and barium-140 were not detected at any station during this period.

Results

Table 1 presents the analytical results for institutional diet samples collected from the basic stations for April through June 1970. The stable elements, calcium and potassium, are reported in g/kg of diet. The radionuclide concentrations of these samples, reported in pCi/kg of diet, are corrected for radioactive decay to the midpoint of the sample collection period, where applicable. Dietary intakes in g/day or pCi/day were obtained by multiplying the food consumption rate in kg/day by the appropriate concentration values. The average food consumption rate during this period was 1.83 kg/day compared to the network average of 1.87 kg/day observed from 1961 through 1969.

Strontium-90 dietary intake averaged 12 pCi/day and cesium-137 intake averaged 18 pCi/day

Table 2. Concentration and intake of stable elements and radionuclides in institutional total diets of children (auxiliary stations), April–June 1970

Location of Institution	Month (1970)	Total weight (kg/day)	Potassium		Cesium-137		Calcium[a]		Strontium-89[a]		Strontium-90[a]	
			(g/kg)	(g/day)	(pCi/kg)	(pCi/day)	(g/kg)	(g/day)	(pCi/kg)	(pCi/day)	(pCi/kg)	(pCi/day)
Alaska: Juneau	April	NS										
	May	NS										
	June	1.48	1.4	2.1	0	0	0.5	0.8	0	0	4	6
Calif: San Francisco	April	NS										
	May	NS										
	June	NS										
Miss: Columbia	April	1.91	1.6	3.0	16	31						
	May	2.08	1.5	3.0	18	37	1.1	2.1	NA	NA	10	20
	June	1.83	1.4	2.5	16	29						
Nebr: Omaha	April	2.35	1.8	4.2	12	28						
	May	1.99	1.6	3.2	0	0	.8	1.8	NA	NA	6	11
	June	1.89	1.6	3.0	0	0						
Nev: Carson City	April	NS										
	May	NS										
	June	1.32	1.4	1.8	0	0	.9	1.2	0	0	7	9
N. Mex: Albuquerque	April	2.13	1.4	3.0	0	0						
	May	2.15	1.5	3.2	0	0	.9	2.0	NA	NA	3	6
	June	2.50	1.7	4.3	0	0						
Oreg: Portland	April	NS										
	May	NS										
	June	1.56	1.9	3.0	0	0	.5	.8	7	11	0	0
Utah: Salt Lake City	April	NS										
	May	2.29	1.2	2.7	0	0	.5	1.2	0	0	0	0
	June	NS										
Network average	April	2.13	1.6	3.4	9	20						
	May	2.13	1.5	3.0	5	9	0.7	1.4	1	2	4	7
	June	1.76	1.6	2.8	3	5						

[a] Composite of monthly samples collected during the quarter.
NA, no analysis.
NS, no sample.
NOTE: Iodine-131 and barium-140 were not detected at these stations during this period.

during this period. These results fall within Range I as defined by the Federal Radiation Council (4). Strontium-89, barium-140, and iodine-131 concentrations were below detectable levels.

All concentrations that are less than or equal to the appropriate minimum detectable level have been reported as zero. The minimum detectable concentration is defined as the measured concentration equal to the 2-standard deviation analytical error. Accordingly, the minimum detectable limits are as follows:

Strontium-89	5 pCi/kg
Strontium-90	2 pCi/kg
Iodine-131	10 pCi/kg
Barium-140	10 pCi/kg
Cesium-137	10 pCi/kg

Data from eight auxiliary stations are included in table 2 for general information. These stations do not meet the criterion that the majority of the samples are collected from children who range in age from 9 to 12 years. In order to supplement the existing environmental monitoring networks of the Bureau of Radiological Health, these eight institutions are being sampled in the same manner as the basic stations.

Recent coverage in *Radiological Health Data and Reports*:

Period	Issue
July–September 1969	June 1970
October–December 1969 and Annual Summary, 1969	August 1970
January–March 1970	November 1970

REFERENCES

(1) ANDERSON, E. C., D. J. NELSON, JR. Surveillance for radiological contamination in foods. Amer J Public Health 52:1391–1400 (September 1962).

(2) PUBLIC HEALTH SERVICE, DIVISION OF RADIOLOGICAL HEALTH. Radionuclides in institutional total diet samples, January–March 1965. Radiol Health Data 6:548–554 (October 1965).

(3) PUBLIC HEALTH SERVICE, NATIONAL CENTER FOR RADIOLOGICAL HEALTH. Radionuclides in institutional total diet samples, January–March 1968. Radiol Health Data Rep 9:557–560 (October 1968).

(4) FEDERAL RADIATION COUNCIL. Background material for the development of radiation protection standards, Report No. 2. Superintendent of Documents, U.S. Government Printing Office, Washington, D.C. 20402 (September 1961).

XUM

Carbon-14 and Tritium in Total Diet and Milk January 1969-June 1970

Bureau of Radiological Health
U.S. Public Health Service

The Bureau of Radiological Health initiated a program in July 1965 to determine the levels of carbon-14 and tritium in the total diet and milk in the United States. Initially, monthly samples from each of the PHS Institutional Total Diet Sampling Network (ITDSN) and Pasteurized Milk Network (PMN) stations were composited and analyzed according to six selected regions: Northeast, South, Delta, Central, Southwest, and Northwest. Figure 1 shows the ITDSN and PMN sampling stations in each of the designated regions.

In January 1966, the program was modified to include selected stations in each of the previously mentioned regions plus Alaska and Hawaii. The nine geographically distributed sampling stations are: Palmer, Alaska; Honolulu, Hawaii; Idaho Falls, Idaho; Chicago, Ill; New Orleans, La; Boston, Mass; Portland, Oreg; and Charleston, S.C. Los Angeles, Calif. is not a PMN station, but a special milk sample is collected for purposes of comparison with the routine ITDSN sample.

A 1-liter milk sample and a 2-kilogram food sample are sent to the PHS Northeastern Radiological Health Laboratory for analysis. The milk samples analyzed represent a composite of the samples collected during the last 2 weeks of the month, and the total diet samples analyzed represent the samples collected for that month.

The carbon-14 analyses are performed semiannually and the tritium analyses were performed monthly through June 1969. Tritium analyses were discontinued then because the results were only slightly above the limit of detectability and were less than 1 percent of the maximum permissible concentration for milk and food consumed by the general population.

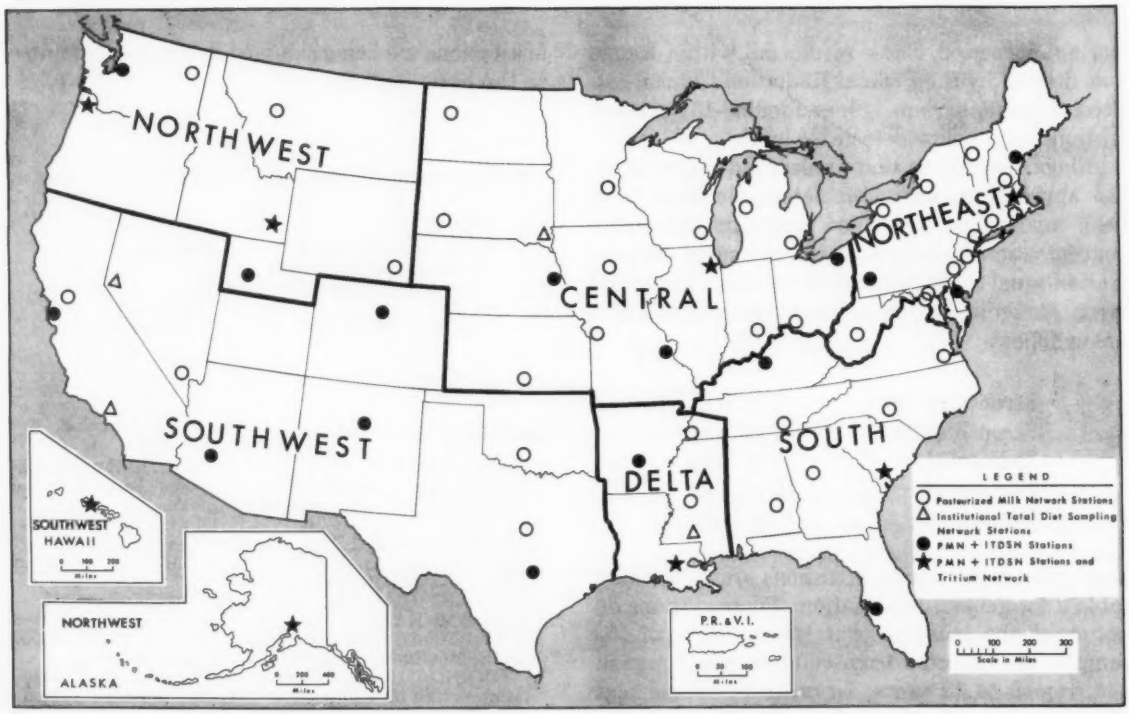

Figure 1. ITDSN and PMN sampling stations

Table 1. Carbon-14 in total diet and milk, January 1969–June 1970

Location	Date collected	Concentration (pCi/kg ±2σ)	
		Total diet	Milk
Alaska: Palmer	April 1969	1,040±50	560±26
	October 1969	880±46	500±20
	March 1970	910±42	470±22
Calif: Los Angeles	April 1969	980±35	510±18
	October 1969	1,000±33	490±21
	March 1970	860±31	
	April 1970		480±22
Hawaii: Honolulu	April 1969	990±37	460±19
	October 1969	950±44	470±13
	April 1970	NS	NS
Idaho: Idaho Falls	April 1969	1,080±31	500±19
	October 1969	980±49	490±19
	April 1970	1,040±33	510±17
Ill: Chicago	April 1969	NS	470±14
	October 1969	NS	460±24
	April 1970	NS	460±22
La: New Orleans	April 1969	1,020±39	510±14
	October 1969	1,040±34	490±16
	April 1970	780±41	
	May 1970		450±22
Mass: Boston	April 1969	1,070±53	480±21
	October 1969	1,040±33	480±17
	April 1970	1,000±31	460±22
Oreg: Portland	April 1969	1,160±49	480±18
	October 1969	1,140±34	490±13
	March 1970	830±31	
	April 1970		520±18
S. C: Charleston	April 1969	850±40	490±52
	October 1969	1,000±43	510±13
	April 1970	840±30	
	May 1970		410±22

NS, no sample.

Analytical procedures

The carbon-14 content is determined by combustion of the milk or food sample and collection and purification of the evolved carbon dioxide. The carbon dioxide is then converted to methane and counted in a gas-proportional counter (*1*). Water simultaneously produced by the combustion is used for the tritium determination. When only the tritium analysis is required, the sample is distilled to dryness and the collected water is used for the determination. In some cases, the sample is electrolytically enriched (*1*) prior to liquid scintillation counting.

Results and discussion

The carbon-14 concentration in food and milk was measured as dpm of carbon-14 per gram of carbon and converted to pCi/kg by using the percent carbon in the sample. The carbon content in food ranged from 9.9 percent to 13.6 percent with an average of 11.9 percent, and the carbon content in milk ranged from 3.3 percent to 7.3 percent with an average of 5.8 percent.

For the tritium measurements, the water content per kilogram of food ranged from 71.1 percent to 95.6 percent with an average of 91.0 percent. The water content of milk is assumed to be 97 percent. The average carbon and water content were used for the calculation of activity per kilogram.

Tables 1 and 2 show the carbon-14 and tritium concentrations in total diet and milk for January 1969–June 1970. Table 3 shows the mean carbon-14 concentrations for all stations for the periods indicated. The data show a gradual downward trend. The mean tritium concentrations in table 4 show a decrease of about a factor of 2 since July 1965. The values for July 1965–December 1968 have been previously published (*2*).

Other coverage in *Radiological Health Data and Reports:*

Period	Issue
July 1965–December 1968	November 1969

REFERENCES

(*1*) DROBINSKI, J. C. JR., D. P. LAGATTA, A. S. GOLDIN, and J. G. TERRILL. Analyses of environmental samples for carbon-14 and tritium. Health Phys 11:385–395 (1965).

(*2*) BUREAU OF RADIOLOGICAL HEALTH. Carbon-14, tritium, and plutonium-239 in milk and total diet, July 1965 to December 1968. Radiol Health Data Rep 10:500–505 (November 1969).

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Table 2. Tritium content of total diet and milk, January–June 1969

Location	Date collected (1969)	Concentration (nCi/kg ±2σ)	
		Total diet	Milk
Alaska: Palmer	January	1.9± .7	1.0± .7
	February	1.3± .8	[a] .7±1.2
	March	[b] .9± .2	[b] .9± .4
	April	[a] .4± .6	[a] 1.3±1.3
	May	1.0± .5	[a] .3± .5
	June	1.0± .3	1.3± .3
Calif: Los Angeles	January	NS	.9± .7
	February	[b] .4± .3	[a,b] 0± .3
	March	[a] .5± .7	1.2± .7
	April	[a] 0± .7	[a] .3± .7
	May	[a] 1.0±1.5	NS
	June	[a] .2± .3	[a] .3± .4
Hawaii: Honolulu	January	1.8± .7	[a] .6± .9
	February	1.1± .8	1.2± .9
	March	[a] .3± .8	[a] .5± .9
	April	[b] .5± .2	[a,b] .1± .2
	May	.9± .7	[a] .6±1.3
	June	[a] 0± .5	[a] 0± .6
Idaho: Idaho Falls	January	1.5± .6	1.9± .7
	February	.8± .6	1.4±1.2
	March	1.6± .8	3.4± .9
	April	[a] .5± .8	[a] 1.2±1.3
	May	.6± .5	[a] 1.0±1.6
	June	[a] .3± .7	1.1± .6
Ill: Chicago	January	NS	[a] .8± .8
	February	NS	[a] .4± .8
	March	NS	1.1± .9
	April	NS	1.3± .7
	May	NS	1.6± .8
	June	NS	[a] 1.1±1.3
La: New Orleans	January	1.2±1.0	[a] .4± .9
	February	1.0± .6	[a] .8± .9
	March	NS	[a] .2± .6
	April	[a] .2± .7	1.3± .7
	May	[a,b] .1± .2	[b] .4± .3
	June	NS	NS
Mass: Boston	January	1.3± .6	[a] .6± .8
	February	2.6± .7	1.6± .8
	March	[a] .8± .8	1.4± .9
	April	1.9± .9	1.1±1.0
	May	1.5± .7	1.1± .7
	June	1.0± .6	1.5±1.4
Oreg: Portland	January	[b] 1.0± .3	[a,b] .2± .3
	February	[a] .5± .8	[a] .6± .9
	March	[a] .5± .7	1.3± .7
	April	[a] .2± .6	[a] .1± .8
	May	[a] .1± .5	[a] .3± .6
	June	[a] .1± .3	[a] .4± .6
S. C: Charleston	January	[a] .9±1.0	1.3± .9
	February	.7± .6	[a] .7± .9
	March	[a] .2± .5	.6± .5
	April	[a] .1± .6	.7± .6
	May	[a] 0± .5	[a] .2± .8
	June	NS	NS

[a] Values are not statistically significant at the 95-percent confidence level.
[b] Electrolytically enriched.
NS, no sample.

Table 3. Summary of the mean carbon-14 concentrations in total diet and milk for all stations

Date collected	Concentration (pCi/kg)	
	Total diet	Milk
July–December 1965	1,120	570
April 1966	1,100	610
October 1966	1,200	555
April 1967	1,290	680
October 1967	1,170	570
April 1968	1,160	550
October 1968	1,030	500
April 1969	1,020	500
October 1969	1,000	490
March–April 1970	890	470

Table 4. Summary of the mean tritium concentrations in total diet and milk for all stations

Date collected	Concentration[a] (nCi/kg)	
	Total diet	Milk
July–December 1965	1.49	1.28
1966	1.17	1.02
1967	.41	.41
1968	.52	.32
January–June 1969	.63	.60

[a] Zero was used for these samples with concentrations below the 2-sigma error for averaging purposes.

SECTION II. WATER

The Public Health Service, the Federal Water Quality Administration and other Federal, State, and local agencies operate extensive water quality sampling and analysis programs for surface, ground, and treated water. Most of these programs include determinations of gross beta and gross alpha radioactivity and specific radionuclides.

Although the determination of the total radionuclide intake from all sources is of primary importance, a measure of the public health importance of radioactivity levels in water can be obtained by comparison of the observed values with the Public Health Service Drinking Water Standards (*1*). These standards, based on consideration of Federal Radiation Council (FRC) recommendations (*2–4*), set the limits for approval of a drinking water supply containing radium-226 and strontium-90 at 3 pCi/liter and 10 pCi/liter, respectively. Limits may be set higher if the total intake of radioactivity from all sources remains within the guides recommended by FRC for control action. In the known absence[1] of strontium-90 and alpha-particle emitters, the limit is 1,000 pCi/liter gross beta radioactivity, except when additional analysis indicates that concentrations of radionuclides are not likely to cause exposures greater than the limits indicated by the Radiation Protection Guides. Surveillance data from a number of Federal and State programs are published periodically to show current and long-range trends. Water sampling activities reported in *Radiological Health Data and Reports* are listed below.

[1] Absence is taken to mean a negligibly small fraction of the specific limits of 3 pCi/liter and 10 pCi/liter for unidentified alpha-particle emitters and strontium-90, respectively.

Water sampling program	Period reported	Issue
California	July–December 1968	August 1970
Coast Guard	January 1968–July 1969	February 1970
Interstate Carrier Drinking Water	1967–1969	December 1970
Kansas	January–December 1969	September 1970
Minnesota	January–June 1969	January 1970
North Carolina	January–December 1967	May 1969
New York	January–June 1969	June 1970
Radiostrontium in Tap Water, *HASL*	January–December 1969	July 1970
Tritium in Community Water Supplies	1969	December 1970
Tritium in Surface Waters	January–June 1970	November 1970
Washington	January–June 1969	June 1970

REFERENCES

(*1*) U.S. PUBLIC HEALTH SERVICE. Drinking water standards, revised 1962, PHS Publication No. 956. Superintendent of Documents, U.S. Government Printing Office, Washington, D.C. 20402 (March 1963).

(*2*) FEDERAL RADIATION COUNCIL. Radiation Protection Guidance for Federal Agencies, Memorandum for the President, September 1961. Reprint from the Federal Register of September 26, 1961.

(*3*) FEDERAL RADIATION COUNCIL. Background material for the development of Radiation Protection Standards, Report No. 1. Superintendent of Documents, U.S. Government Printing Office, Washington, D.C. 20402 (May 1960).

(*4*) FEDERAL RADIATION COUNCIL. Background material for the development of Radiation Protection Standards, Report No. 2. Superintendent of Documents, U.S. Government Printing Office, Washington, D.C. 20402 (September 1961).

XUM

SECTION III. AIR AND DEPOSITION

Radioactivity in Airborne Particulates and Precipitation

Continuous surveillance of radioactivity in air and precipitation provides one of the earliest indications of changes in environmental fission product radioactivity. To date, this surveillance has been confined chiefly to gross beta radioanalysis. Although such data are insufficient to assess total human radiation exposure from fallout, they can be used to determine when to modify monitoring in other phases of the environment.

Surveillance data from a number of programs are published monthly and summarized periodically to show current and long-range trends of atmospheric radioactivity in the Western Hemisphere. These include data from activities of the Environmental Protection Agency, the Canadian Department of National Health and Welfare, the Mexican Commission of Nuclear Energy, and the Pan American Health Organization.

In addition to those programs presented in this issue, the following programs were previously covered in *Radiological Health Data and Reports.*

Network	Period	Issue
Fallout in the United States and Other Areas, *HASL*	January–June 1968	October 1969
Plutonium in Airborne Particulates and Precipitation, *PHS*	April–December 1969	October 1970
Surface Air Sampling, 80th Meridian Network	January–December 1966	October 1970
Tritium in Precipitation	July–December 1969	July 1970

1. Radiation Alert Network September 1970

National Air Pollution Control Administration
U.S. Public Health Service

Surveillance of atmospheric radioactivity in the United States is conducted by the Radiation Alert Network (RAN) which regularly gathers samples at 70 locations distributed throughout the country (figure 1). Most of the stations are operated by State Health Department personnel.

The station operators perform "field estimates" on the airborne particulate samples at 5 hours after collection, when most of the radon daughter products have decayed, and at 29 hours after collection, when most of the thoron daughter products have decayed. They also perform field estimates on dried precipitation samples and report all results to appropriate National Air Pollution Control Administration officials by mail or telephone depending on levels found. A compilation of the daily field estimates is available upon request from the Data Acquisition and Analysis Branch, Division of Air Quality and Emission Data, NAPCA, Cincinnati, Ohio. A detailed description of the sampling and analytical procedures was presented in the March 1968 issue of *Radiological Health Data and Reports*.

Table 1 presents the monthly average gross beta radioactivity in surface air particulates and deposition by precipitation, as measured by the field estimate technique during September 1970. Time profiles of gross beta radioactivity in air for eight Radiation Alert Network stations are shown in figure 2.

All field estimates reported were within normal limits for the reporting station.

Figure 1. Radiation Alert Network sampling stations

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Table 1. Gross beta radioactivity in surface air and precipitation, September 1970

Station location		Number of samples	Air surveillance gross beta radioactivity (pCi/m³)			Last profile in RHD&R	Number of samples	Precipitation			
			Maximum	Minimum	Average[a]			Total depth (mm)	Field estimation of deposition		
									Number of samples	Depth (mm)	Total deposition (nCi/m³)
Ala:	Montgomery	17	4	1	2	Oct 70	1	36	1	36	0
Alaska:	Anchorage	[b]				June 70	[c]				
	Attu Island	29	0	0	0	Nov 70	[c]				
	Fairbanks	19	2	0	1	July 70	9	28	9	28	18
	Juneau	6	7	2	5	Aug 70	5	147	5	147	0
	Kodiak	1	0	0	0	Sept 70	[c]				
	Nome	14	2	0	1	Dec 70	[c]				
	Point Barrow	[b]				Dec 70	[c]				
Ariz:	Phoenix	13	7	0	4	Aug 70	[c]				
Ark:	Little Rock	[b]				June 70	[c]				
Calif:	Berkeley	19	3	0	1	Sept 70	[c]				
	Los Angeles	20	4	0	2	Jan 71	[c]				
C.Z:	Ancon	11	0	0	0	Sept 70	[c]				
Colo:	Denver	19	7	0	3	Sept 70	[c]				
Conn:	Hartford	21	2	0	1	July 70	9	71	9	71	0
Del:	Dover	21	2	0	1	May 70	[c]				
D.C:	Washington	25	1	0	0	Dec 70	[c]				
Fla:	Jacksonville	20	1	0	0	June 70	4	66	3	62	10
	Miami	10	0	0	0	July 70	8	196	8	196	0
Ga:	Atlanta	19	2	1	1	Jan 71	[c]				
Guam:	Agana	[b]				May 70	[c]				
Hawaii:	Honolulu	25	1	0	0	Nov 70	5	27	[d]		
Idaho:	Boise	21	5	0	3	Nov 70	2	22	2	22	6
Ill:	Springfield	4	5	1	2	Dec 70	[c]				
Ind:	Indianapolis	21	5	0	1	Jan 71	[c]				
Iowa:	Iowa City	20	4	0	1	Sept 70	5	154	5	154	32
Kans:	Topeka	21	3	0	1	June 70	11	161	11	161	0
Ky:	Frankfort	4	1	0	1	Dec 70	[c]				
La:	New Orleans	18	1	0	0	Aug 70	12	151	[d]		
Maine:	Augusta	20	3	0	1	Jan 71	8	79	8	79	0
Md:	Baltimore	20	1	0	1	July 70	1	3	1	3	0
Mass:	Lawrence	21	2	0	1	Sept 70	5	55	5	55	1
	Winchester	18	2	0	1	Oct 70	5	68	5	68	0
Mich:	Lansing	20	2	0	1	Nov 70	10	112	10	112	11
Minn:	Minneapolis	21	2	0	1	May 70	7	117	7	117	5
Miss:	Jackson	15	2	0	1	Aug 70	2	53	2	53	8
Mo:	Jefferson City	22	3	0	1	Jan 71	12	224	12	224	0
Mont:	Helena	20	3	1	1	Oct 70	1	3	1	3	0
Nebr:	Lincoln	16	5	0	2	Jan 71	4	133	4	133	14
Nev:	Las Vegas	21	2	0	2	July 70	[c]				
N.H:	Concord	[b]				Dec 70	[c]				
N.J:	Trenton	21	2	0	1	Aug 70	6	31	5	28	2
N.Mex:	Santa Fe	17	3	0	1	Oct 70	6	36	6	36	0
N.Y:	Albany	2	0	0	0	Jan 71	2	42	2	42	13
	Buffalo	18	1	0	1	Sept 70	[c]				
	New York City	[b]				Oct 70	[c]				
N.C:	Gastonia	17	15	2	8	Sept 70	1	10	[d]		
N.Dak:	Bismarck	20	5	0	2	Dec 70	4	47	4	47	6
Ohio:	Cincinnati	[b]				May 70	[c]				
	Columbus	5	2	0	1	Aug 70	[c]				
	Painesville	19	4	0	1	July 70	12	184	12	184	27
Okla:	Oklahoma City	20	3	0	1	Nov 70	6	242	6	242	4
	Ponca City	21	6	0	2	July 70	9	137	9	137	0
Oreg:	Portland	21	2	0	1	Apr 70	6	39	6	39	0
Pa:	Harrisburg	14	2	0	1	Apr 70	1	21	1	21	2
P.R:	San Juan	[b]				Aug 70	[c]				
R.I:	Providence	18	2	0	1	Nov 70	3	13	3	13	1
S.C:	Columbia	19	4	0	2	Oct 70	3	69	3	69	0
S.Dak:	Pierre	21	8	1	3	Aug 70	[c]				
Tenn:	Nashville	20	3	0	1	Nov 70	5	50	5	50	2
Tex:	Austin	[b]				May 70	[c]				
	El Paso	[b]				Dec 70	[c]				
Utah:	Salt Lake City	29	3	0	1	Jan 71	5	57	5	57	2
Vt:	Barre	18	2	0	1	June 70	8	74	8	74	11
Va:	Richmond	21	2	0	1	June 70	2	29	2	29	7
Wash:	Seattle	10	0	0	0	June 70	5	43	[d]		
	Spokane	14	3	0	2	May 70	[c]				
W.Va:	Charleston	21	3	0	1	Oct 70	4	66	3	62	4
Wisc:	Madison	19	3	0	1	June 70	9	606	9	606	7
Wyo:	Cheyenne	21	7	0	3	July 70	4	32	4	32	4
Network summary		1,057	16	0	1		227	101	5	94	5

[a] The monthly average is calculated by weighting the field estimates of individual air samples with length of sampling period.
[b] No report received. (Air samples received without field estimate data are not considered by the data program.)
[c] No precipitation sample collected.
[d] This station is part of the plutonium in precipitation network. No grose beta measurements are done.

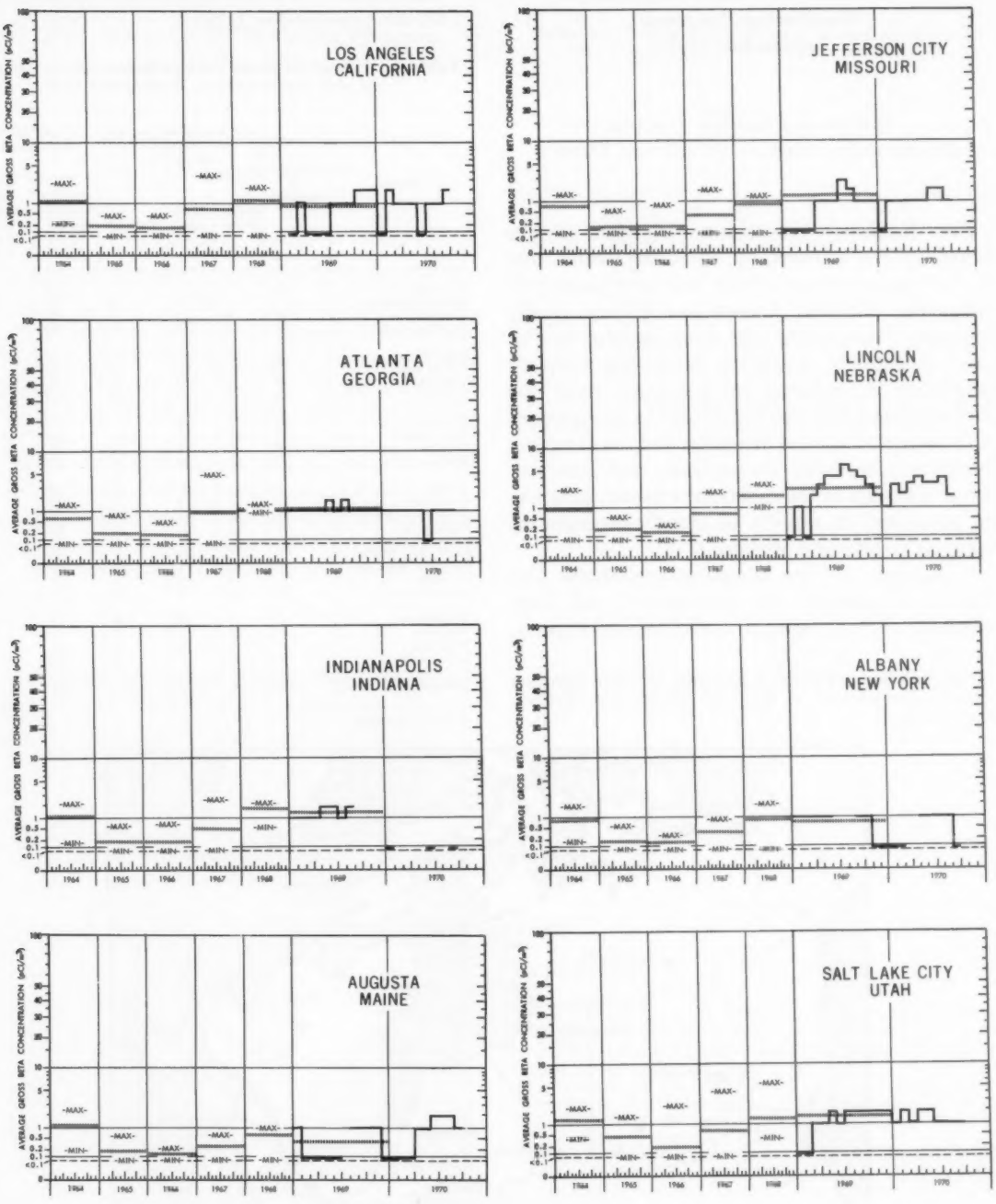

Figure 2. Monthly and yearly profiles of beta radioactivity in air—Radiation Alert Network, 1964-September 1970

2. Canadian Air and Precipitation Monitoring Program[1] September 1970

Radiation Protection Division
Department of National Health and Welfare

The Radiation Protection Division of the Canadian Department of National Health and Welfare monitors surface air and precipitation in connection with its Radioactive Fallout Study Program. Twenty-four collection stations are located at airports (figure 3), where the sampling equipment is operated by personnel from the Meteorological Services Branch of the Department of Transport. Detailed discussions of the sampling procedures, methods of analysis, and interpretation of results of the radioactive fallout program are contained in reports of the Department of National Health and Welfare (*1–5*).

A summary of the sampling procedures and methods of analysis was presented in the May 1969 issue of *Radiological Health Data and Reports.*

[1] Prepared from information and data obtained from the Canadian Department of National Health and Welfare, Ottawa, Canada.

Surface air and precipitation data for September 1970 are presented in table 2.

Table 2. Canadian gross beta radioactivity in surface air and precipitation, September 1970

Station	Number of samples	Air surveillance, gross beta radioactivity (pCi/m³)			Precipitation measurements	
		Maximum	Minimum	Average	Average concentration (pCi/liter)	Total deposition (nCi/m²)
Calgary	0	0	0	0	177	4.1
Coral Harbour	6	.1	.0	.0	48	2.5
Edmonton	7	.2	.1	.2	66	1.9
Ft. Churchill	5	.1	.0	.1	79	2.4
Fredericton	4	.1	.0	.1	NS	NS
Goose Bay	5	.1	.0	.1	15	1.0
Halifax	6	.1	.1	.1	31	3.4
Inuvik	5	.1	.0	.0	412	16.0
Montreal	30	.3	.0	.1	NS	NS
Moosonee	5	.1	.1	.1	27	2.8
Ottawa	7	.2	.1	.1	12	1.6
Quebec	7	.1	.0	.1	27	3.7
Regina	5	.2	.1	.1	33	2.3
Resolute	5	.1	.0	.1	39	.6
St. John's, Nfld.	5	.1	.1	.1	25	2.8
Saskatoon	5	.1	.0	.1	85	.3
Sault Ste. Marie	4	.2	.1	.1	26	5.1
Thunder Bay	7	.2	.1	.1	40	3.5
Toronto	7	.3	.1	.2	25	1.3
Vancouver	5	.1	.0	.1	74	4.0
Whitehorse	7	.1	.0	.0	75	3.4
Windsor	7	.3	.1	.1	14	.8
Winnipeg	5	.1	.1	.1	48	5.3
Yellowknife	7	.1	.0	.0	33	.8
Network summary	156	0.3	0.0	0.1	64	3.2

NS, no sample.

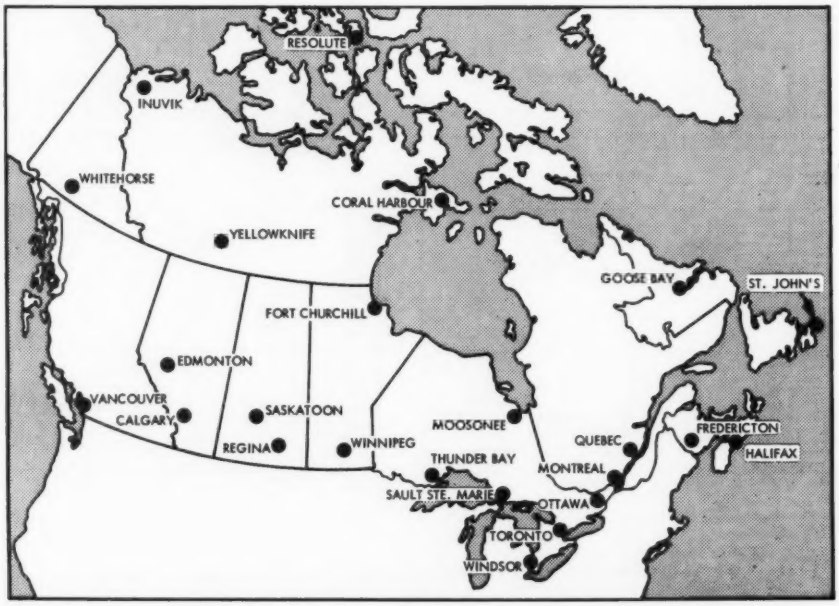

Figure 3. Canadian air and precipitation sampling stations

3. Pan American Air Sampling Program September 1970

Pan American Health Organization and
U.S. Public Health Service

Gross beta radioactivity in air is monitored by countries in the Americas under the auspices of the collaborative program developed by the Pan American Health Organization (PAHO) and the U.S. Public Health Service (PHS) to assist PAHO-member countries in developing radiological health programs.

The air sampling station locations are shown in figure 3. Analytical techniques were described in the March 1968 issue of *Radiological Health Data and Reports*. The September 1970 air monitoring results from the participating countries are given in table 3.

Table 3. Summary of gross beta radioactivity in Pan American surface air, September 1970

Station location		Number of samples	Gross beta radioactivity (pCi/m³)		
			Maximum	Minimum	Average[a]
Argentina:	Buenos Aires	6	0.51	0.11	0.26
Bolivia:	La Paz	17	3.05	.43	1.20
Chile:	Santiago	24	.80	.10	.32
Colombia:	Bogota	22	.26	.01	.10
Ecuador:	Cuenca	NS	NS	NS	NS
	Guayaquil	NS	NS	NS	NS
	Quito	NS	NS	NS	NS
Guyana:	Georgetown	14	.52	.05	.28
Jamaica:	Kingston	NS	NS	NS	NS
Peru:	Lima	25	1.74	.13	.94
Venezuela:	Caracas	11	.58	.04	.23
West Indies:	Trinidad	NS	NS	NS	NS
Pan American summary		119	3.05	0.01	0.52

[a] The monthly average is calculated by weighting the individual samples with length of sampling period. Values less than 0.005 pCi/m³ are reported and used in averaging as 0.00 pCi/m³.
NS, no sample.

Figure 4. Pan American Air Sampling Program

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(1) BIRD, P. M., A. H. BOOTH, and P. G. MAR. Annual report of 1959 on the Radioactive Fallout Study Program, CNHW-RP-3. Department of National Health and Welfare, Ottawa, Canada (May 1960).

(2) BIRD, P. M., A. H. BOOTH and P. G. MAR. Annual report for 1960 on the Radioactive Fallout Study Program, CNHW-RP-3. Department of National Health and Welfare, Ottawa, Canada (December 1961).

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(4) BEALE, J. and J. GORDON. The operation of the Radiation Protection Division Air Monitoring Program, RPD-11. Department of National Health and Welfare, Ottawa, Canada (July 1962).

(5) BOOTH, A. H. The calculation of permissible levels of fallout in air and water and their use in assessing the significance of 1961 levels in Canada, RPD-21. Department of National Health and Welfare, Ottawa, Canada (August 1962).

XUM

Fallout in the United States and Other Areas[1] July-December 1968 and January-December 1969

Health and Safety Laboratory
U.S. Atomic Energy Commission

Monthly fallout deposition rates for strontium-90 are determined by the Health and Safety Laboratory (HASL) for 36 sites in the United States and 103 locations in other countries. HASL data from all of the active United States stations and other selected points in the Western Hemisphere (figure 1) covering the period from July–December 1968, January–December 1969 are summarized in tables 1 through 6. All the stations of the 80th Meridian Network are represented.

[1] The data in this article were taken from "Fallout Program Quarterly Summary Report," HASL-237:A-4 to A-285. (October 1, 1970).

Methods of collection

Two methods of fallout collection are employed by HASL. In the first, precipitation and dry fallout are collected for a period of 1 month in a stainless-steel pot with an exposed area of 0.076 m². At the end of the collection period, the contents are transferred, by careful scrubbing with a rubber spatula, to a polyethylene sample bottle which is then shipped to the laboratory for analysis.

The second method involves the use of a polyethylene funnel, with an exposed area of 0.072 m², attached to an ion exchange column. After a 1-month collection, the inside of the funnel is wiped with a tissue, and the tissue is inserted in

Table 1. Strontium-90 fallout in the United States, HASL, July–December 1968

Sampling location		Type of collection	Deposition (nCi/m²)					
			July	August	September	October	November	December
Ala:	Birmingham	pot	0.24	0.09	0.06	0.02	0.08	0.06
Alaska:	Anchorage	column	.08	.02	.02	.02	[a]	[a]
	Barrow	column	.02	[a]	[a]	.01	[a]	[a]
	Cold Bay	column	.04	.09	.08	.08	.07	.03
	Fairbanks	column	.05	.07	.01	.01	[a]	.01
	Juneau	column	.12	[b]	.23	.08	.05	[a]
	Nome	column	.11	[b]	[b]	[b]	[b]	[b]
Calif:	W. Los Angeles	pot	.02	.01	[a]	.02	[a]	.04
	San Francisco	column	[a]	.02	[c].02	[c].02	[b]	.04
Colo:	Denver	column	.10	.06	.06	.05	.02	[a]
Fla:	Coral Gables	pot	[c].04	[c].04	.06	[b]	[b]	.09
	Miami	column	.12	.07	.06	.02	.01	[a]
Hawaii:	Hilo	column	[c].14	.15	.06	.04	.04	.17
	Honolulu	pot	[c].40	[c].40	[c].16	[c].16	[c].06	[c].06
	Lihue	column	.05	.04	.04	.02	.07	.05
	Mauna Loa	column	.02	.04	.04	.01	.02	[a]
Ill:	Argonne	pot	.07	.11	.09	.04	.04	.04
La:	New Orleans	column	.05	.03	.02	.02	.04	.04
Minn:	International Falls	column	.18	.07	.06	.06	[a]	[a]
Mo:	Columbia	column	.07	.05	.05	.04	.06	.06
Mont:	Helena	column	.06	.12	.08	.02	.02	[a]
N.Y.:	New York City	pot	.07	.08	.04	.06	.09	.06
N. Dak:	Williston	column	.14	.09	.03	.03	.02	[a]
Ohio:	Wooster	pot	.11	.06	.02	.03	.06	.03
Okla:	Tulsa	pot	[b]	.04	.04	.03	.07	[b]
Oreg:	Medford	column	.01	.10	.02	.02	.04	.02
S.C:	Columbia	column	.10	.02	.04	.03	.08	.02
S. Dak:	Vermillion	pot	.32	.09	.10	.09	.02	.05
Tex:	Dallas	column	.04	.03	.04	.02	.05	.02
	El Paso	column	.04	.03	.01	.02	.03	[a]
	Houston	column	.09	.04	.04	.03	[a]	.03
Utah:	Salt Lake City	pot	.03	.33	.08	.05	.12	.04
Va:	Sterling	column	.12	.06	.03	.02	.05	.02
Wash:	Seattle	pot	[c].03	[c].10	[c].10	.09	.06	.10
	Forks	column	.12	.08	.06	.25	.14	.16
Wisc:	Green Bay	column	.11	.08	.11	.03	.02	.02

[a] Zero or trace.
[b] Data not available.
[c] Proportioned from originally consolidated data.

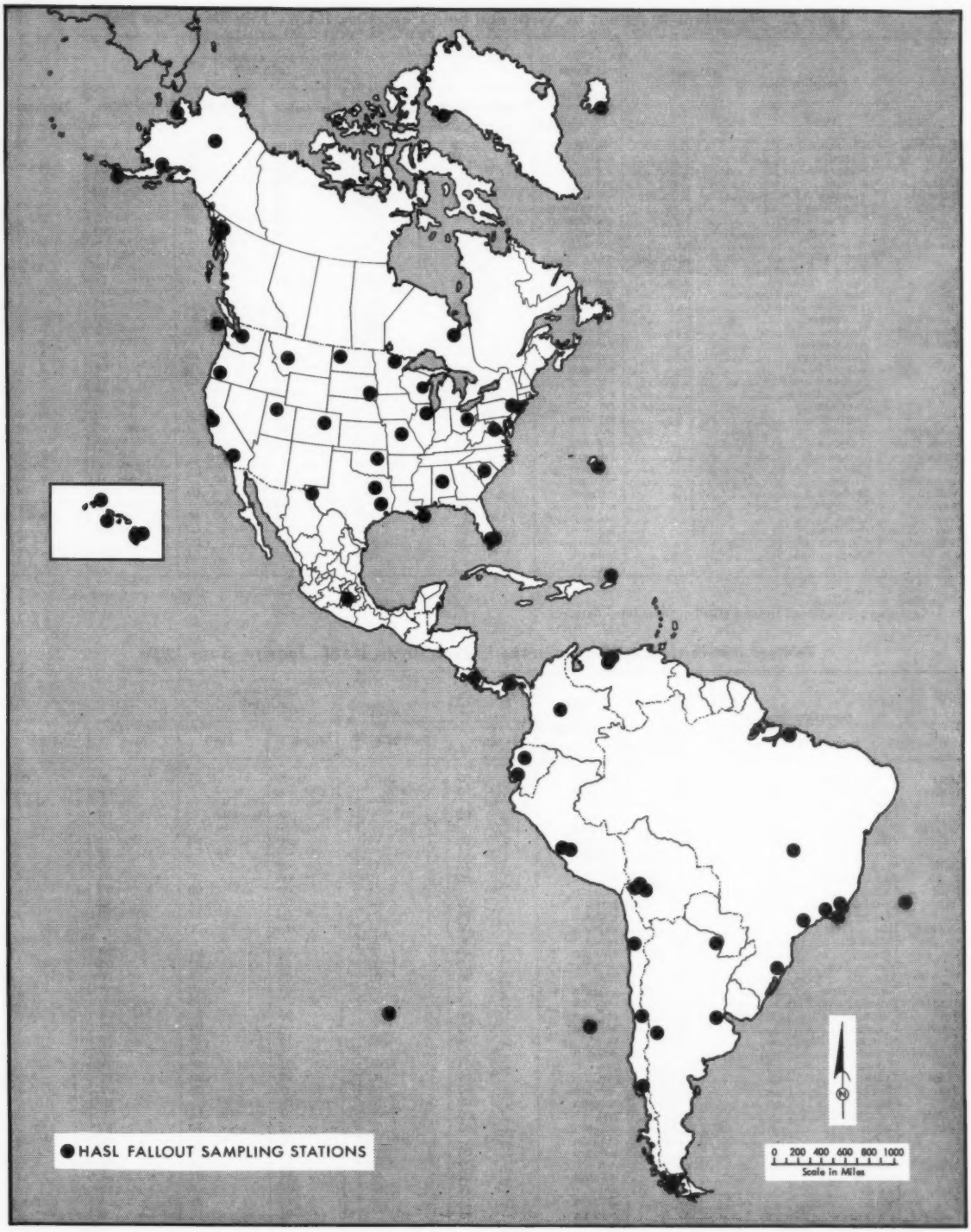

Figure 1. HASL fallout sampling stations in the Western Hemisphere

Table 2. Strontium-90 fallout in North and South America, HASL, July–December 1968

Sampling location		Type of collection	Deposition (nCi/m²)					
			July	August	September	October	November	December
Argentina:	Buenos Aires	column	0.02	0.19	0.04	0.21	0.21	0.33
	Formosa	column	.02	.07	.03	.09	.06	.03
	Malargue	column	[a]	[a]	.02	.02	.03	[a]
Bahamas:	Bimini	column	.08	.06	.01	[b]	[b]	[a]
Bermuda:	Kindley AFB	column	.10	.06	.05	.02	.06	.10
Bolivia:	Chacaltaya	column	.01	.09	.04	.03	.02	.07
	La Paz (city)	column	[a]	.03	.05	.05	.05	.03
	La Paz (Ovejuyo)	column	[a]	.03	.04	.04	.02	.04
Brazil:	Belem	column	[b]	[b]	[b]	[b]	[b]	[b]
	Itaici Sao Paulo	pot	[b]	[b]	[b]	[b]	[b]	[b]
	Nova Friburgo	pot	[b]	[b]	[b]	[b]	[b]	[b]
	Rio de Janeiro	column	.06	.09	.13	.10	.11	.05
	San Jose Dos Compos	pot	[b]	[b]	[b]	[b]	[b]	[b]
	Trindade Island	column	.02	.11	.08	.10	.10	.03
Canada:	Moosonee	column	.24	.27	.04	.10	.02	[a]
Canal Zone:	Balboa	column	.03	.04	.01	.02	.02	[a]
Chile:	Antofagasta	column	[b]	[b]	[b]	[b]	[b]	[b]
	Conception	column	.02	.13	.01	.03	.04	[a]
	Easter Island	column	.03	.04	.09	.02	.06	.02
	Puerto Montt	column	.05	[c].11	[c].12	.08	.11	.16
	Punta Arenas	column	[a]	.03	.02	[a]	.02	.04
	Santiago	pot	.01	.03	.04	[a]	[a]	[a]
	Santiago	column	[a]	.01	.05	.01	[b]	[a]
Colombia:	Bogota	pot	[a]	.01	.01	[a]	[a]	[b]
Costa Rica:	Turrialba	column	.06	.04	.03	.01	[a]	.02
Equador:	Guayaquil	column	[b]	[b]	[b]	[b]	[b]	[b]
	Quito	column	[b]	[b]	[b]	[b]	[b]	[b]
Greenland:	Thule	column	.02	.02	.01	[a]	[a]	[a]
Iceland:	Keflavik	column	.08	.15	.09	.04	.04	.05
Mexico:	Mexico City	column	.07	.04	.03	[a]	.02	[a]
Peru:	Lima	column	[a]	.01	.01	.01	[a]	[a]
Puerto Rico:	San Juan	column	.07	.08	.05	.02	.06	.04
Venezuela:	Caracas (site 1)	column	.05	.02	.01	.02	[a]	[a]
	Caracas (site 2)	column	[a]	.02	[b]	.01	.01	[a]

[a] Zero or trace.
[b] Data not available.
[c] Indicates proportioned from originally consolidated data.

Table 3. Strontium-90 fallout in the United States, HASL, January–June 1969

Sampling location		Type of collection	Deposition (nCi/m²)					
			January	February	March	April	May	June
Ala:	Birmingham	pot	0.08	0.05	0.09	0.07	0.20	0.10
Alaska:	Anchorage	column	.01	[a]	[a]	[a]	.09	.02
	Barrow	column	[a]	[a]	[a]	[a]	[a]	[a]
	Cold Bay	column	.02	[a]	.03	[b]	.08	.10
	Fairbanks	column	[a]	[a]	[a]	[a]	.05	.15
	Juneau	column	.05	.11	.10	.08	[b]	.02
	Nome	column	.11	[a]	[b]	[a]	.02	.05
Calif:	W. Los Angeles	pot	.15	.14	.02	.02	.03	.01
	San Francisco	column	.09	.10	.03	.06	.01	[a]
Colo:	Denver	column	[a]	.01	.02	.08	.46	.12
Fla:	Coral Gables	pot	.03	.03	.04	.14	[c].12	[c].13
	Miami	column	.06	.03	.03	.12	.19	.18
Hawaii:	Hilo	column	.06	.41	.15	.22	[b]	[b]
	Honolulu	pot	[c].20	[c].20	[b]	[b]	[b]	.02
	Lihue	column	.06	.07	.06	.04	.06	.05
	Mauna Loa	column	.02	.02	.04	.02	.02	[a]
Ill:	Argonne	pot	.05	.01	.04	.25	.26	.18
La:	New Orleans	column	.03	.08	.07	.10	.11	.03
Minn:	International Falls	column	.02	[a]	[a]	.10	.09	.20
Mo:	Columbia	column	.06	.05	.05	.15	.26	.36
Mont:	Helena	column	.02	[a]	[a]	.06	.09	.14
N.Y:	New York City	pot	.04	.06	.16	.13	.26	.11
N. Dak:	Williston	column	[a]	[a]	[a]	.07	.09	.20
Ohio:	Wooster	pot	.04	.01	.01	.03	.05	.17
Okla:	Tulsa	pot	.09	.03	.23	.29	[c].11	[c].12
Oreg:	Medford	column	.03	.03	.01	.03	.09	.14
S.C:	Columbia	column	.02	.06	.06	[a]	.10	.18
S. Dak:	Vermillion	pot	.06	.03	.05	.15	.29	.53
Tex:	Dallas	column	.03	.03	.12	.10	.23	.05
	El Paso	column	[a]	[a]	.02	[a]	.07	[a]
	Houston	column	.03	.10	.09	.08	.09	.06
Utah:	Salt Lake City	pot	.07	.06	.12	.14	.08	.36
Va:	Sterling	column	.04	.02	.11	.07	.04	.13
Wash:	Seattle	pot	.17	.04	.05	.15	.10	.06
	Forks	column	.15	.09	.10	.22	.14	.04
Wisc:	Green Bay	column	.04	[a]	.04	.13	.08	.27

[a] Zero or trace.
[b] Data not available.
[c] Proportioned from originally consolidated data.

Table 4. Strontium-90 fallout in North and South America, HASL, January–June 1969

Sampling location		Type of collection	Deposition (nCi/m²)					
			January	February	March	April	May	June
Argentina:	Buenos Aires	column	0.06	0.07	0.08	0.08	0.08	0.05
	Formosa	column	[a]	.05	[b]	[b]	.02	.01
	Malargue	column	.04	.02	.02	.02	.01	.02
Bahamas:	Bimini	column	.01	.11	[a]	.04	.09	.08
Bermuda:	Kindley AFB	column	.02	[c].08	[c].08	.12	.06	[b]
Bolivia:	Chacaltaya	column	.04	[a]	.02	[a]	[a]	[a]
	La Paz (city)	column	.03	.01	[a]	[a]	[a]	[a]
	La Paz (Ovejuyo)	column	.03	.02	[a]	[a]	[a]	[a]
Brazil:	Belem	column	[b]	[b]	[b]	[b]	[b]	[b]
	Rio de Janeiro	column	[b]	[b]	.06	.06	.02	.02
	Trindade Island	column	.04	[b]	.02	[b]	.04	[b]
Canada:	Moosonee	column	.02	[a]	[a]	.04	.09	.26
Canal Zone:	Balboa	column	[a]	[a]	[a]	.02	.02	.01
Chile:	Antofagasta	column	[b]	[b]	[b]	[b]	[b]	[b]
	Conception	column	[a]	[a]	.01	.09	.11	.18
	Easter Island	column	.05	.03	.05	.02	.06	.03
	Puerto Montt	column	.07	[b]	.07	.17	[b]	.18
	Puerta Arenas	column	.04	.07	.03	.04	.04	.07
	Santiago	pot	[a]	[a]	[a]	.02	.06	.02
	Santiago	column	.02	.01	[a]	.03	.08	.04
Colombia:	Bogota	pot	[b]	[a]	[b]	[a]	[a]	[b]
Costa Rica:	Turrialba	column	.01	[a]	.02	.02	.03	.01
Equador:	Guayaquil	column	[b]	[b]	[b]	[a]	[a]	[a]
Greenland:	Thule	column	[a]	[a]	[a]	[a]	[c].01	[c].02
Iceland:	Keflavik	column	.03	.05	.10	.08	.10	.13
Mexico:	Mexico City	column	[a]	[a]	[a]	[a]	.02	.04
Peru:	Lima	column	[a]	[a]	[a]	[a]	[a]	[a]
Puerto Rico:	San Juan	column	.06	.02	.01	.05	.06	.07
Venezuela:	Caracas (site 1)	column	[a]	[a]	[a]	.02	.01	.01
	Caracas (site 2)	column	[a]	[a]	[a]	.01	.02	[a]

[a] Zero or trace.
[b] Data not available.
[c] Proportioned from originally consolidated data.

Table 5. Strontium-90 fallout in the United States, HASL, July-December 1969

Sampling location		Type of collection	Deposition (nCi/m²)					
			July	August	September	October	November	December
Ala:	Birmingham	pot	0.09	0.04	0.02	0.03	0.01	0.06
Alaska:	Anchorage	column	.05	.04	.02	.03	.01	.01
	Barrow	column	[a]	[a]	[a]	.01	[a]	[a]
	Cold Bay	column	.08	.12	.04	.05	.02	.03
	Fairbanks	column	.06	.14	.03	.01	[a]	[a]
	Juneau	column	.12	[b]	.23	.08	.05	[a]
	Nome	column	.04	.07	[a]	.03	[a]	[a]
Calif:	W. Los Angeles	pot	.02	[b]	[a]	.01	.04	.15
	San Francisco	column	.02	[a]	[a]	.03	.01	.03
Colo:	Denver	column	.12	.03	.04	.07	.01	.01
Fla:	Coral Gables	pot	[c].08	[c].08	.01	.03	.04	.01
	Miami	column	.11	.07	.05	.05	.01	.01
Hawaii:	Hilo	column	[b]	[b]	[b]	[b]	[b]	[b]
	Honolulu	pot	.12	.12	.06	.03	.08	.04
	Lihue	column	.06	.03	[a]	.03	.01	.04
	Mauna Loa	column	[a]	[a]	.02	.01	[a]	.01
Ill:	Argonne	pot	.14	.08	.04	.07	.02	.01
La:	New Orleans	column	.16	.08	[a]	.01	.02	.04
Minn:	International Falls	column	.24	.16	.07	.06	.01	[a]
Mo:	Columbia	column	.15	.06	.06	.09	.01	.03
Mont:	Helena	column	.15	.04	[a]	.03	.01	[a]
N.Y:	New York City	pot	.26	.08	.13	.04	.06	.10
N. Dak:	Williston	column	.26	.09	.04	.03	[a]	[a]
Ohio:	Wooster	pot	.33	.11	.05	.01	.07	.03
Okla:	Tulsa	pot	[b]	.04	.04	.03	.07	[b]
Oreg:	Medford	column	.02	[a]	.02	.04	.01	.03
S.C:	Columbia	column	.06	.10	[a]	.02	.01	.06
S. Dak:	Vermillion	pot	.26	.14	.02	.06	.01	[a]
Tex:	Dallas	column	.02	.08	.02	.06	.02	.05
	El Paso	column	.10	.02	[a]	.02	[a]	.01
	Houston	column	.11	.06	.09	.03	.02	.03
Utah:	Salt Lake City	pot	.08	.05	.05	.09	.04	.06
Va:	Sterling	column	.09	.12	.05	[a]	.03	.04
Wash:	Seattle	pot	.03	.01	.08	.03	.06	.02
	Forks	column	.03	.10	.10	.13	.11	.14
Wisc:	Green Bay	column	.13	.06	.05	.06	.01	[a]

[a] Zero or trace.
[b] Data not available.
[c] Proportioned from originally consolidated data.

Table 6. Strontium-90 fallout in North and South America, HASL, July–December 1969

Sampling location		Type of collection	Deposition (nCi/m²)					
			July	August	September	October	November	December
Argentina:	Buenos Aires	column	0.04	[a]	0.10	0.08	0.10	0.08
	Formosa	column	[a]	0.02	.05	.04	.05	.03
	Malargue	column	[a]	[a]	.02	.06	.03	.12
Bahamas:	Bimini	column	.06	[a]	[b]	.01	.02	.01
Bermuda:	Kindley AFB	column	.09	.03	.04	.06	.03	.06
Bolivia:	Chacaltaya	column	[a]	[a]	[a]	.02	.01	.02
	La Paz (city)	column	[a]	[a]	[a]	[a]	.01	.02
	La Paz (Ovejuyo)	column	.01	[a]	[a]	.01	[a]	.01
Brazil:	Belem	column	[b]	[b]	[b]	[b]	[b]	[b]
	Rio de Janeiro	column	.02	[b]	[b]	[b]	[b]	.13
	Trindade Island	column	[a]	[b]	.05	.04	.08	.02
Canada:	Moosonee	column	.15	.30	.08	.04	.02	.01
Canal Zone:	Balboa	column	.03	.04	[a]	.01	.01	[a]
Chile:	Antofagasta	column	[b]	[b]	[b]	.01	[b]	[b]
	Conception	column	.16	.07	.07	.07	.03	.02
	Easter Island	column	.08	.07	.07	.05	.02	.04
	Puerto Montt	column	.06	.24	.21	.10	.09	[b]
	Punta Arenas	column	[b]	.05	.05	.01	.02	[b]
	Santiago	pot	.05	.05	.03	.01	[a]	.01
	Santiago	column	.05	.05	.01	.04	.01	.01
Colombia:	Bogota	pot	[b]	[a]	[a]	.01	[a]	[a]
Costa Rica:	Turrialba	column	.02	.02	.02	.01	.01	.02
Equador:	Guayaquil	column	[a]	[b]	[a]	[a]	[a]	.09
Greenland	Thule	column	[c].01	[c].02	[c].03	[a]	[a]	[a]
Iceland:	Keflavik	column	.10	.12	.03	.10	.02	.02
Mexico:	Mexico City	column	.04	[a]	.02	.01	[a]	[a]
Peru:	Lima	column	[a]	[a]	.02	.01	[a]	[a]
Puerto Rico:	San Juan	column	.08	.06	.02	.08	.06	.01
Venezuela:	Caracas (site 1)	column	.02	.01	[a]	.02	[a]	.02
	Caracas (site 2)	column	.02	[a]	[b]	[b]	[b]	[b]

[a] Zero or trace.
[b] Data not available.
[c] Proportioned from originally consolidated data.

the end of the column, which is then sealed and sent to HASL for analysis. It has been shown that at the 95-percent confidence level there was no significant difference in the strontium-90 measurements obtained from samples collected by the two methods (*1*).

The analysis of monthly precipitation samples for other radionuclides that were being performed at laboratories in Westwood, N.J., New York City, N.Y., and Seattle, Wash., has been discontinued as of July 1967.

REFERENCE

(*1*) ONG, L. D. Y. Homogeneity between pot and ion exchange column strontium-90 measurements. Fallout Program Quarterly Summary Report, HASL-135:256–269. Office of Technical Services, Department of Commerce, Washington, D.C. (April 1, 1963).

Surface Air Sampling Program-80th Meridian Network[1] January-December 1967

Health and Safety Laboratory
Atomic Energy Commission

The Health and Safety Laboratory began its Surface Air Sampling Program in January 1963, as a continuation of the 80th Meridian Program conducted by the U.S. Naval Research Laboratory. The objectives of this program are to determine the concentrations of radioactivity in surface air in both hemispheres as a function of latitude and season and to interpret these data in terms of the dominant meteorological and physical processes which are operative.

The basic network consists of a line of sites approximately along the 80th Meridian extending from about 76° N to 63° S latitudes (figure 1). Since 1963, a number of sites have been added to investigate the possible effects of longitude, elevation and proximity to coastlines; and in late 1965, samplers were placed on four Atlantic Ocean weather ships to extend the surface air study over the marine environment. The Environmental Science Services Administration, Department of Commerce, maintains a collection site at the Amundson Scott Station in Antarctica. The air filter samples from this site are also included in the HASL program.

Sampling and analytical procedures

Approximately 1,400 cubic meters of ambient air per day are drawn through an 8-inch diameter microsorban filter for the land stations. For the ocean stations, about 2,200 cubic meters of air per day are filtered by an 8- by 10-inch microsorban filter. Each filter is changed on the 1st, 8th, 15th, and 22nd of the month or more frequently if the filter becomes clogged with debris suspended in the air. Under normal conditions, the filters from each station are compressed into a monthly composite and the gamma spectrum of the composite is obtained by an 8- by 4-inch NaI (Tl) crystal approximately 2 weeks after the last collection in the month.

The integrated response between 100 keV and 3.0 MeV is corrected by the average detection efficiency (35 percent) of the gamma photons present in fallout and the total gamma radioactivity is reported in units of photons/min/1,000 m^3. Average monthly gamma concentrations are calculated by weighting the concentrations in each sampling interval by the relative period of time in the interval. After the gamma measurements have been completed, monthly composites from each site are submitted to contractor laboratories for radiochemical analyses.

Since the last major nuclear weapon test series occurred at the end of 1962, only the longer lived artificially produced radionuclides were present in the filters analyzed prior to May 1966. Consequently, emphasis was given to the determination of manganese-54, iron-55, strontium-90, cadmium-109, cesium-137, cerium-144, plutonium-238, and plutonium-239. Strontium-89, zirconium-95, and cerium-141 were analyzed in samples collected after the foreign nuclear test of May 1966.

The longer lived fission products and plutonium-239 concentrations should describe the general distribution in surface air of all previous nuclear weapon debris which was transferred from the lower stratosphere to the troposphere during the collection period of this report. Other tracer radionuclides can be associated with debris from a single detonation or a series of detonations. Manganese-54 and iron-55 were produced in large quantities in the 1961 and 1962 test series. Cadmium-109 was generated by the U.S. high altitude test over Johnston Island on July 9, 1962. While plutonium-238 is present in low concentrations in nuclear weapons debris, about 17,000 curies of plutonium-238 were disseminated at high altitude in the stratosphere on April 21, 1964 during the reentry burnup of a SNAP-9A power source. As the levels of any of the radionuclides drop to

[1] Summarized from "Fallout Program Quarterly Summary Report," HASL 204, 237 from the National Technical Information Service, 5285 Port Royal Road, Springfield, Va. 22151.

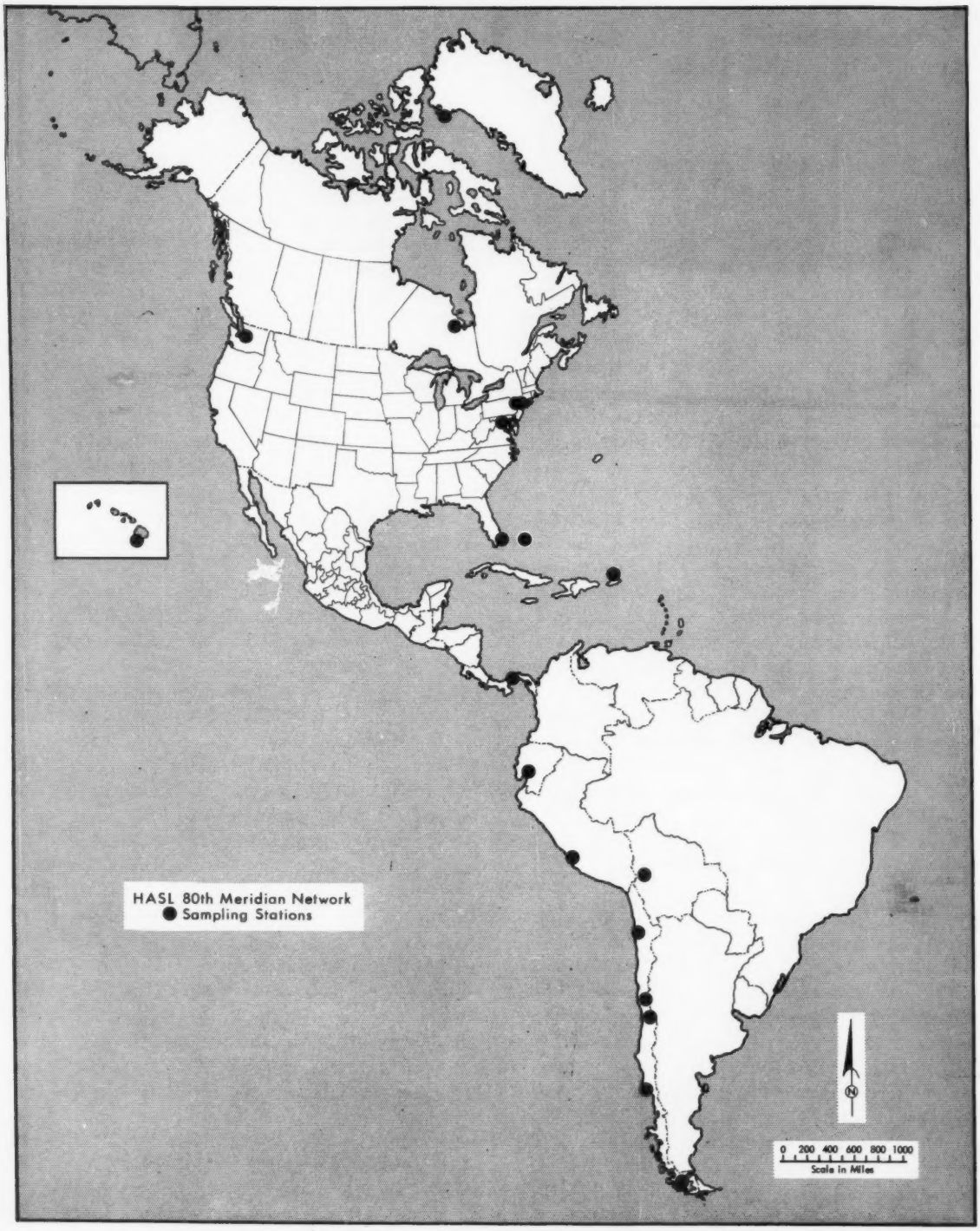

Figure 1. 80th Meridian Network sampling stations

Table 1. Station location

Site		Latitude	Longitude (west)	Elevation (meters)
Greenland:	Thule	76° 36′ N	68° 35′	259
Ocean station:	Bravo	56° 30′ N	51° 00′	0
	Charlie	52° 45′ N	35° 30′	0
Ontario:	Moosonee	51° 16′ N	80° 30′	10
Wash:	Seattle	47° 36′ N	122° 20′	3
Ocean station:	Delta	44° 00′ N	41° 00′	0
N.J.:	Westwood	41° 00′ N	74° 01′	38
N.Y:	New York	40° 48′ N	73° 58′	38
Va:	Sterling	38° 58′ N	77° 25′	76
Ocean station:	Echo	35° 00′ N	48° 00′	0
Fla:	Miami	25° 49′ N	80° 17′	4
Bahamas:	Bimini	25° 46′ N	79° 22′	3
Hawaii:	Mauna Loa	19° 28′ N	155° 36′	3,401
P.R:	San Juan	18° 26′ N	66° 00′	10
Panama:	Balboa	8° 58′ N	79° 34′	23
Ecuador:	Guayaquil	2° 10′ S	79° 52′	7
Peru:	Lima	12° 01′ S	77° 08′	13
Bolivia:	Chacaltaya	16° 21′ S	68° 07′	5,220
Chile:	Antofagasta	23° 37′ S	70° 16′	31
	Portillo	32° 50′ S	70° 08′	2,850
	Santiago	33° 27′ S	70° 42′	520
	Puerto Montt	41° 27′ S	72° 57′	7
	Punta Arenas	53° 08′ S	70° 53′	35

Table 2. Iron-55 concentrations in surface air, January-December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	50.67	63.85	75.51	79.04	17.24	18.40	104.75	[a]18.01	[a]9.23	17.95	18.14	21.44
Ocean station:	Bravo	NS	NS	28.04	49.65	25.67	NS	16.77	[a]7.00	12.84	[a]6.46	8.17	15.76
	Charlie	NS	30.88	35.45	45.73	43.40	42.64	NS	9.94	18.94	NS	[a]7.16	[a]10.37
Ontario:	Moosonee	20.34	NS	66.59	78.69	47.76	54.26	207.03	28.52	17.85	18.91	[a]10.14	13.73
Wash.:	Seattle	[a]36.52	48.15	[a]7.41	[a]37.30	78.76	58.05	37.12	NS	NS	NS	NS	NS
Ocean station:	Delta	NS	NS	51.27	56.70	67.27	56.65	26.53	19.15	26.98	20.33	19.40	10.08
N.J.:	Westwood	33.24	43.93	47.06	82.26	126.03	100.53	[a]47.94	NS	NS	NS	NS	NS
N.Y.:	New York	44.23	79.67	71.57	99.54	116.48	163.03	167.63	34.60	28.64	15.91	33.53	17.40
Va.:	Sterling	37.19	NS	NS	100.00	56.56	104.44	191.42	[a]17.48	NS	[a]26.39	[a]11.09	[a]11.75
Ocean station:	Echo	NS	39.81	55.65	200.80	47.52	76.35	43.88	[a]57.07	16.09	14.94	12.38	17.74
Fla.:	Miami	52.42	NS	84.90	179.29	75.46	36.34	202.54	15.78	[a]7.09	34.33	37.66	32.64
Bahamas:	Bimini	71.00	NS	111.38	231.92	119.48	49.73	[a]35.98	[a]15.98	[b]	[a]28.49	[a]30.30	[a]26.89
Hawaii:	Mauna Loa	106.90	NS	64.69	121.57	78.73	60.88	44.03	15.71	[a]28.49	[a]13.29	21.57	22.27
P.R.:	San Juan	31.19	31.04	77.17	116.26	36.95	17.78	[a]17.83	[a]30.96	[a]20.38	[a]9.60	[a]20.16	[a]17.22
Panama:	Balboa	22.92	NS	51.27	35.66	48.77	[a]4.85	[a]5.29	[a]5.38	[a]8.31	[b]	[a]23.04	[a]4.58
Ecuador:	Guayaquil	[a]9.07	NS	11.73	7.03	6.83	[a]5.61	[a]10.50	[b]	[b]	[b]	[b]	[b]
Peru:	Lima	32.45	NS	13.68	11.49	26.28	12.39	[b]	[a]21.32	[a]25.90	[a]63.59	[a]22.02	[b]
Bolivia:	Chacaltaya	11.32	NS	5.23	9.74	7.67	11.76	19.94	29.02	17.35	23.18	30.71	[b]
Chile:	Antofagasta	42.33	NS	14.35	13.97	10.65	62.53	[a]25.48	[a]19.78	24.97	33.73	[a]17.39	[a]20.51
	Portillo	64.10	NS	59.52	40.48	39.83	NS	NS	NS	NS	NS	NS	NS
	Santiago	[a]42.77	[a]58.86	26.85	31.92	13.21	38.69	[b]	[a]18.55	[a]13.76	[b]	[b]	[b]
	Puerto Montt	NS	35.23	42.52	18.29	24.77	[a]8.12	[a]14.62	[a]6.59	[a]7.68	[a]13.66	[a]9.11	[a]9.44
	Punta Arenas	[a]11.58	NS	42.79	11.59	10.32	6.16	[a]5.03	[a]5.12	[a]39.07	[b]	[b]	[b]

Errors are less than 20 percent except:
[a] Error between 20–100 percent.
[b] Error greater than 100 percent.
NS, no sample.

below practical determination limits, they are eliminated from the radiochemical program; thus cadmium-109 was not analyzed after the end of 1967. Manganese-54 and cesium-137 were not analyzed for 1967.

During 1966, Food Chemical and Research Laboratories, Incorporated (FCRL) of Seattle, Wash., performed the analyses of most of the samples in this program. Isotopes, Incorporated of Westwood, N.J., analyzed most of the samples collected at Westwood, N.J., and a few samples from Chacaltaya, Bolivia, Antofagasta and Santiago, Chile. Since the beginning of 1967 the analyses were contracted as follows: January and February, 1967, Tracerlab, Inc., Richmond, Calif.; March and April 1967, Custom Nuclear Co., Stanford, Calif.; May and June, 1967, Isotopes, Inc., Palo Alto, Calif.; July through December, 1967, Tracerlab Inc., Richmond, Calif.

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Table 3. Strontium-89 concentrations in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	92.63	55.16	[a]66.33	[a]50.60	88.94	[b]	[b]	[b]	[b]	[b]	[a]8.54	[a]3.65
Ocean station:	Bravo	12.35	NS	[a]25.64	[a]22.38	18.33	NS	[a]1.09	[b]	[a]1.80	[a]1.49	[a].53	5.56
	Charlie	NS	NS	[a]28.25	[a]30.49	37.94	15.93	NS	[a]0.72	[a].75	NS	[a]1.14	NS
Ontario:	Moosonee	27.93	37.41	[a]45.02	[a]53.86	104.24	22.74	[b]	[b]	[b]	[a]2.31	[a]7.13	[a]1.90
Wash.:	Seattle	40.05	39.68	29.19	71.08	86.83	14.91	[a]2.67	NS	NS	NS	NS	NS
Ocean station:	Delta	29.25	44.44	[a]31.75	[a]42.06	44.74	13.19	[a]1.80	[b]	[a]1.91	[a]1.56	[a]2.12	NS
N.J.:	Westwood	NS	NS	NS	NS	80.29	9.66	[a]1.65	NS	NS	NS	NS	NS
N.Y.:	New York	46.89	63.79	86.53	72.94	82.27	22.06	0.99	1.40	1.78	2.25	2.09	9.45
Va.:	Sterling	51.25	45.50	60.41	[a]98.32	40.72	38.32	[b]	[b]	[a]1.01	[b]	[a]5.92	[a]3.17
Ocean station:	Echo	65.29	52.23	[a]58.30	[a]104.42	33.06	6.63	[a]2.41	[a]3.78	[a]1.46	[a].90	[a]1.47	[a]2.35
Fla.:	Miami	17.08	90.03	[a]116.34	[a]107.14	133.25	[b]	[b]	[a]3.14	[a]1.44	[a]1.20	[a]30.57	3.97
Bahamas:	Bimini	73.50	NS	[a]100.27	[a]98.87	NS	8.69	3.37	[a]1.55	[b]	[a]3.45	[a]5.79	5.79
Hawaii:	Mauna Loa	309.66	110.88	[a]52.81	[a]62.09	30.79	18.81	18.68	NS	[a]3.41	[a]2.15	[a]58.62	[a]4.69
P.R.:	San Juan	8.89	NS	[a]89.33	[a]93.60	42.61	[b]	[a]2.23	[a]1.80	[a]1.04	[a]2.33	[a]1.54	[a]2.99
Panama:	Balboa	8.97	NS	[a]53.52	[a]25.84	24.09	NS	4.66	5.13	[a]1.48	[a].80	[a]17.80	[a]1.31
Ecuador:	Guayaquil	5.01	NS	[b]	[b]	[b]	NS	78.30	19.88	[b]	4.66	5.48	[a]1.55
Peru:	Lima	[a]20.24	NS	[b]	[b]	[b]	NS	172.89	NS	37.18	21.12	6.45	[a]2.24
Bolivia:	Chacaltaya	13.48	42.64	[b]	[b]	[b]	120.61	284.95	252.55	36.59	13.39	[a]2.09	[a].33
Chile:	Antofagasta	NS	12.65	[b]	[b]	[b]	28.71	234.33	75.52	25.28	[a]12.59	[a]6.69	[b]
	Portillo	57.05	NS	[b]	[b]	54.39	NS	NS	NS	NS	NS	NS	NS
	Santiago	23.08	NS	[a]24.69	[b]	[b]	15.26	40.63	61.80	30.39	15.03	6.96	[a]8.43
	Puerto Montt	20.16	NS	[b]	[b]	[b]	9.26	19.45	16.03	10.52	6.74	[a]5.38	[a]2.81
	Punta Arenas	[a]3.53	NS	[a]18.02	[b]	[b]	4.51	1.19	5.76	[a]6.22	3.69	[a]2.46	2.01

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

Results

The updated radioactivity concentrations in surface air during 1967 are presented in tables 2 through 10. The sites are listed according to latitude beginning with the most northern site at Thule, Greenland (table 1). The results of the blank and standard analyses are reported in tables 11 through 13.

The concentrations are reported at the midpoint of the collection month for the plutonium isotopes and the fission products and on the following dates

Table 4. Strontium-90 concentration in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	5.98	7.00	8.93	8.70	5.17	2.00	1.83	1.60	0.94	0.86	[a]0.75	1.96
Ocean station:	Bravo	1.37	NS	2.61	2.98	2.75	3.26	1.73	1.03	1.11	.79	.57	1.54
	Charlie	NS	NS	2.94	4.16	4.34	2.49	NS	1.12	1.12	1.07	.88	1.50
Ontario:	Moosonee	2.21	5.08	6.87	8.06	6.42	5.22	3.79	3.09	2.43	[a]1.62	[a]1.60	.75
Wash.:	Seattle	2.90	4.13	5.92	7.62	6.72	4.46	3.91	NS	NS	NS	NS	NS
Ocean station:	Delta	4.02	3.90	3.52	5.19	[a]6.50	4.20	2.68	2.24	3.53	1.88	1.39	1.40
N.J.:	Westwood	3.72	6.97	6.50	11.26	9.73	6.70	4.61	NS	NS	NS	NS	NS
N.Y.:	New York	4.19	7.60	6.66	7.61	8.51	7.61	4.00	3.39	3.21	2.43	1.06	1.64
Va.:	Sterling	4.13	5.57	4.66	10.55	6.65	6.31	3.75	2.58	2.10	1.12	[b]	1.43
Ocean station:	Echo	5.36	4.43	4.95	15.86	4.83	5.77	4.51	1.99	1.89	1.71	1.28	.92
Fla.:	Miami	1.71	7.98	8.96	15.31	6.78	1.92	3.44	1.61	1.09	1.53	[b]	2.41
Bahamas:	Bimini	6.27	NS	11.46	17.91	NS	3.32	4.21	1.60	1.59	2.48	2.74	2.86
Hawaii:	Mauna Loa	10.79	11.23	6.50	11.31	5.52	3.61	2.82	NS	1.59	1.36	[a]7.68	1.97
P.R.:	San Juan	1.95	NS	7.17	10.27	4.19	2.10	3.45	2.47	.91	.82	.89	2.26
Panama:	Balboa	1.97	NS	5.32	3.18	3.11	NS	1.26	.72	.22	.15	[a]1.98	.76
Ecuador:	Guayaquil	.59	NS	1.06	.55	.59	NS	[a]1.02	.63	[a].04	.63	1.61	1.02
Peru:	Lima	3.14	NS	1.56	.93	1.85	NS	2.46	NS	2.87	3.22	1.77	1.42
Bolivia:	Chacaltaya	1.74	4.08	.25	.95	.70	2.88	3.63	5.37	2.02	2.63	1.89	.98
Chile:	Antofagasta	NS	3.28	1.60	NS	1.61	8.27	4.17	3.57	2.29	[a]2.32	2.26	2.56
	Portillo	7.47	NS	2.60	2.60	3.97	NS	NS	NS	NS	NS	NS	NS
	Santiago	3.57	NS	2.52	1.38	1.60	4.38	2.58	3.16	3.37	2.49	2.53	[a]1.89
	Puerto Montt	2.87	NS	2.26	1.84	1.06	1.07	2.00	1.15	.88	1.13	1.16	1.67
	Punta Arenas	2.08	NS	2.00	.99	.95	.56	.70	.66	[a].47	.82	.65	.95

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

Table 5. Zirconium-95 concentrations in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	NS	44.13	[b]	[b]	[b]	2.03	[b]	26.56	[a]0.46	6.09	9.77	22.97
Ocean station:	Bravo	9.24	NS	[b]	[b]	2.77	2.67	[a]0.30	[a].12	9.03	8.28	11.42	14.55
	Charlie	NS	NS	[b]	[b]	3.54	2.55	NS	[a].17	3.52	NS	8.62	14.63
Ontario:	Moosonee	NS	NS	[b]	[b]	3.01	3.90	[b]	2.80	3.94	12.17	5.54	14.69
Wash.:	Seattle	30.98	27.25	27.30	19.08	4.57	4.17	[a].74	NS	NS	NS	NS	NS
Ocean station:	Delta	NS	22.54	[b]	[b]	4.95	3.47	.65	24.22	18.88	20.53	16.59	23.69
N.J.:	Westwood	NS	NS	NS	NS	7.30	7.23	[a].99	NS	NS	NS	NS	NS
N.Y.:	New York	NS	47.00	11.41	12.43	6.33	8.42	1.91	1.45	22.36	.18	16.33	92.78
Va.:	Sterling	NS	30.25	NS	[b]	7.01	7.52	[b]	[a]1.50	4.07	13.58	18.63	18.82
Ocean station:	Echo	34.94	NS	[b]	[b]	2.85	5.50	1.53	[a]1.88	21.28	19.63	11.68	19.59
Fla.:	Miami	NS	NS	[b]	[b]	5.44	1.86	[b]	3.63	5.94	25.64	36.79	34.73
Bahamas:	Bimini	184.25	NS	[b]	[b]	7.16	2.67	4.66	7.62	5.50	35.62	35.26	41.56
Hawaii:	Mauna Loa	NS	NS	[b]	[b]	3.46	3.54	[b]	32.94	14.10	12.09	29.56	28.81
P.R.:	San Juan	9.08	NS	[b]	[b]	3.57	2.34	3.02	3.40	7.92	8.91	12.53	24.07
Panama:	Balboa	NS	15.20	[b]	[b]	3.48	1.16	9.03	9.78	4.21	[a]1.43	3.43	8.51
Ecuador:	Guayaquil	NS	5.35	[b]	[b]	[b]	4.51	149.80	58.88	52.10	21.07	17.28	10.17
Peru:	Lima	NS	NS	[b]	[b]	[b]	10.27	380.35	NS	80.51	57.86	22.69	14.82
Bolivia:	Chacaltaya	20.74	3.33	[b]	[b]	[b]	22.65	447.65	506.57	101.26	35.67	12.16	[a]2.58
Chile:	Antofagasta	NS	NS	[b]	[b]	1.63	5.45	462.59	178.15	85.03	37.05	[b]	[b]
	Portillo	NS	NS	[b]	[b]	2.13	NS	NS	NS	NS	NS	NS	NS
	Santiago	NS	NS	[b]	[b]	[b]	5.84	80.17	NS	82.35	[a]1.72	22.91	9.87
	Puerto Montt	NS	19.74	[b]	[b]	.91	3.94	35.80	32.05	25.35	17.44	10.41	7.00
	Punta Arenas	17.78	8.50	[b]	[b]	.81	[b]	2.51	12.62	11.65	9.37	4.85	4.72

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

Table 6. Cadmium-109 concentrations in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth) ×100											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	[b]	[b]	[b]	[b]	[a]26.44	[b]	[b]	[b]	[b]	[a]52.14	[b]	NS
Ocean station:	Bravo	[b]	[b]	[b]	[b]	[b]	NS	[b]	[b]	[b]	[b]	[b]	NS
	Charlie	NS	[b]	[b]	[b]	NA	[b]	NS	[b]	[b]	NS	[b]	NS
Ontario:	Moosonee	[b]	[a]44.71	[b]	[a]60.89	52.94	[b]	[b]	[b]	[b]	[b]	[b]	NS
Wash.:	Seattle	[b]	NS	[b]	[a]57.57	[a]25.81	[b]	[b]	NS	NS	NS	NS	NS
Ocean station:	Delta	[b]	[b]	[b]	[b]	NS	[a]12.39	[b]	[b]	[b]	[b]	[b]	NS
N.J.:	Westwood	[a]54.44	[b]	[b]	[b]	[a]29.44	[b]	[b]	NS	NS	NS	NS	NS
N.Y.:	New York	NS	NS	9.56	24.82	NS	61.88	9.77	34.89	36.94	NS	12.37	NS
Va.:	Sterling	[b]	[b]	[b]	[a]26.38	NS	[b]	[b]	[b]	[b]	[b]	[b]	NS
Ocean station:	Echo	[b]	[b]	[b]	[b]	NS	[a]19.13	[b]	[b]	[b]	[b]	[b]	NS
Fla.:	Miami	[b]	[b]	[b]	[a]38.10	[b]	[b]	[b]	[b]	[b]	[b]	[b]	NS
Bahamas:	Bimini	[b]	[b]	[b]	70.62	[b]	[b]	[b]	[b]	[b]	[b]	[b]	NS
Hawaii:	Mauna Loa	[b]	[b]	[b]	[b]	NS	[a]13.61	[b]	[b]	[b]	[b]	[b]	NS
P.R.:	San Juan	[b]	[b]	[b]	[a]32.02	[a]16.01	[b]	[b]	[b]	[b]	[a]84.41	[b]	NS
Panama:	Balboa	[b]	[b]	[b]	[b]	NS	[b]	[b]	[b]	[b]	NS	[b]	NS
Ecuador:	Guayaquil	[b]	[b]	[b]	[b]	43.38	[b]	[b]	[b]	[b]	NS	[b]	NS
Peru:	Lima	[b]	[b]	[b]	[b]	NS	[b]	[b]	[b]	[b]	NS	[b]	NS
Bolivia:	Chacaltaya	[b]	[b]	[b]	[b]	[b]	[b]	[b]	[b]	[b]	[b]	[b]	NS
Chile:	Antofagasta	[b]	[b]	[b]	[b]	NS	[b]	[b]	[b]	NS	[b]	[b]	NS
	Portillo	[b]	[b]	[b]	[b]	97.07	NS	NS	NS	NS	NS	NS	NS
	Santiago	[b]	[b]	[b]	[b]	NS	[b]	[b]	NS	[b]	[b]	[b]	NS
	Puerto Montt	[b]	[b]	[b]	[b]	[b]	NS	[b]	[b]	[b]	[b]	[b]	NS
	Punta Arenas	[b]	[b]	[a]24.77	[b]	[a]14.50	[b]	[b]	[b]	[b]	[b]	[b]	NS

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

for the induced nuclides:

Manganese-54 and iron-55	October 15, 1961
Cadmium-109	July 9, 1962

One standard deviation of the counting error for these data is always less than ±20 percent unless otherwise indicated.

Data reliability

The quality of the radiochemical analysis is monitored through the use of various "knowns" submitted to the contractor each month along

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Table 7. Cerium-141 concentrations in surface air, January-December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	466.52	105.87	[b]	[b]	[b]	[b]	[b]	[a]4.19	NS	NS	NS	[a]13.04
Ocean station:	Bravo	[a]11.64	[a]42.60	[b]	[b]	[b]	[b]	[b]	[b]	NS	NS	NS	[b]
	Charlie	NS	76.14	[b]	[b]	[b]	[b]	NS	[b]	NS	NS	NS	NS
Ontario:	Moosonee	[b]	[a]72.71	[b]	[b]	[b]	[b]	[b]	[a]2.78	NS	NS	NS	[a]9.98
Wash.:	Seattle	[b]	[b]	[b]	[a]195.41	[b]	[b]	[a]3.01	NS	NS	NS	NS	NS
Ocean station:	Delta	128.83	[a]48.79	[b]	[b]	[b]	[b]	[a]3.03	[b]	NS	NS	NS	NS
N.J.:	Westwood	NS	NS	NS	NS	[b]	[b]	[a]3.96	NS	NS	NS	NS	NS
N.Y.:	New York	238.23	96.33	NS	NS	NS	NS	1.11	20.47	NS	NS	NS	115.53
Va.:	Sterling	172.79	[a]40.88	[a]33.48	[b]	[b]	[b]	[b]	[a]2.62	NS	NS	NS	[b]
Ocean station:	Echo	74.48	[a]59.42	[b]	[b]	[b]	[b]	[a]3.51	NS	NS	NS	NS	[a]8.99
Fla.:	Miami	243.96	[a]54.22	[b]	[b]	[b]	[b]	[b]	[b]	NS	NS	NS	[a]9.61
Bahamas:	Bimini	400.00	[a]121.80	[b]	[b]	[b]	[b]	[a]4.03	[a]9.75	NS	NS	NS	[a]42.54
Hawaii:	Mauna Loa	[a]306.55	98.60	[b]	[b]	[b]	[b]	[b]	[a]9.56	NS	NS	NS	[b]
P.R.:	San Juan	48.02	[a]37.40	[b]	[b]	[b]	[b]	[b]	[b]	NS	NS	NS	[a]33.07
Panama:	Balboa	[a]23.63	[a]30.62	[b]	[b]	[b]	[b]	16.14	[a]12.80	NS	NS	NS	[a]14.31
Ecuador:	Guayaquil	[a]16.31	[a]8.50	[b]	[b]	[b]	[b]	279.47	96.07	NS	NS	NS	[b]
Peru:	Lima	[a]116.89	[a]27.14	[b]	[b]	[b]	[b]	690.30	[a]66.51	NS	NS	NS	[a]15.99
Bolivia:	Chacaltaya	[a]21.69	[b]	[b]	[b]	[b]	[b]	731.05	658.76	NS	NS	NS	[b]
Chile:	Antofagasta	[a]81.16	[a]56.08	[b]	[b]	[b]	NS	999.51	234.51	NS	NS	NS	[a]11.66
	Portillo	[a]258.01	[a]26.53	[b]	[b]	[b]	[b]	NS	NS	NS	NS	NS	NS
	Santiago	[a]79.38	[a]78.60	[b]	[b]	[b]	[b]	133.67	155.40	NS	NS	NS	[a]12.88
	Puerto Montt	[a]69.57	[b]	[b]	[b]	[b]	[b]	63.13	41.33	NS	NS	NS	[b]
	Punta Arenas	[a]107.41	[a]20.79	[b]	[b]	[b]	[b]	[a]3.48	[a]13.19	NS	NS	NS	[a]10.70

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

Table 8. Cerium-144 concentrations in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth)											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	57.14	21.06	22.58	22.05	8.61	3.76	1.54	3.57	0.67	3.25	6.82	17.09
Ocean station:	Bravo	3.63	10.73	7.61	7.90	4.93	4.49	1.29	.79	3.89	3.29	5.78	7.97
	Charlie	NS	8.99	11.79	10.78	7.44	3.64	NS	.72	2.52	[a]3.12	3.94	10.89
Ontario:	Moosonee	6.41	13.65	23.08	23.35	12.16	8.76	2.88	3.05	2.88	5.93	4.69	9.04
Wash.:	Seattle	12.67	11.67	16.46	13.41	10.19	4.27	3.12	NS	NS	NS	NS	NS
Ocean station:	Delta	11.66	11.11	15.67	12.04	10.86	5.36	[a]2.07	5.69	8.02	8.33	11.55	16.06
N.J.:	Westwood	17.22	18.84	18.61	21.39	13.48	8.18	3.31	NS	NS	NS	NS	NS
N.Y.:	New York	31.54	22.98	17.91	28.85	14.18	16.31	3.70	4.67	10.28	8.55	9.82	39.27
Va.:	Sterling	28.57	15.43	19.07	23.36	12.53	8.36	3.02	2.10	3.16	6.76	9.88	12.45
Ocean station:	Echo	15.15	11.75	16.25	39.80	8.37	6.54	3.36	1.81	6.71	10.42	7.86	15.24
Fla.:	Miami	36.47	24.02	22.70	46.19	12.61	2.63	4.06	2.54	2.36	11.55	21.11	24.86
Bahamas:	Bimini	53.75	30.79	33.88	90.68	40.69	3.98	4.43	4.29	3.05	14.44	21.07	29.10
Hawaii:	Mauna Loa	50.69	20.77	10.66	16.41	6.76	4.83	5.09	4.07	7.41	7.23	17.87	22.30
P.R.:	San Juan	5.84	7.79	16.80	21.90	6.87	3.97	3.08	2.90	3.97	5.12	6.81	18.28
Panama:	Balboa	4.80	8.82	10.87	5.89	3.33	.37	3.16	4.04	1.76	.86	2.10	6.41
Ecuador:	Guayaquil	6.61	3.63	2.79	1.48	.92	7.08	36.53	21.03	18.25	11.85	13.03	5.94
Peru:	Lima	35.88	11.42	6.11	2.87	2.10	20.88	98.38	41.11	35.90	34.91	17.98	7.77
Bolivia:	Chacaltaya	11.91	3.00	.69	1.72	1.10	56.23	105.31	170.47	42.59	17.60	8.61	2.73
Chile:	Antofagasta	26.05	16.30	5.80	3.60	3.37	11.80	117.95	64.18	34.52	21.85	15.33	8.82
	Portillo	54.49	17.18	7.82	8.50	6.82	NS	NS	NS	NS	NS	NS	NS
	Santiago	40.31	27.73	10.46	14.42	3.09	8.83	23.69	42.42	36.24	23.88	20.23	12.19
	Puerto Montt	18.89	18.08	7.62	4.47	2.24	2.19	10.51	10.93	10.52	9.98	7.89	7.44
	Punta Arenas	10.69	6.73	7.18	4.08	1.55	1.23	1.20	4.89	5.11	5.96	3.79	4.37

[a] Errors are less than 20 percent except error between 20–100 percent.
NS, no sample.

Table 9. Plutonium-238 concentrations in surface air, January-December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth) ×100											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	1.26	2.31	4.06	4.63	2.88	2.94	1.64	[a]1.94	[a]0.66	[a]1.93	[a]1.05	2.03
Ocean station:	Bravo	[a].30	1.62	1.17	2.04	1.78	NS	[a]1.58	1.14	1.29	1.45	.60	.90
	Charlie	NS	[a]1.44	1.74	2.91	2.47	NS	NS	1.20	1.38	8.08	.59	1.73
Ontario:	Moosonee	1.04	2.28	2.37	4.15	4.94	6.95	3.72	2.81	2.58	2.08	NS	1.07
Wash.:	Seattle	[b]	[a].45	[b]	[b]	NS	NS	4.54	NS	NS	NS	NS	NS
Ocean station:	Delta	1.53	1.72	3.02	3.63	4.68	NS	3.49	NS	[b]	[b]	1.82	2.45
N.J.:	Westwood	4.07	4.86	3.93	6.14	NS	NS	5.32	NS	NS	NS	NS	NS
N.Y.:	New York	1.91	3.02	3.41	5.12	4.89	15.00	5.52	5.31	3.63	2.78	2.24	4.78
Va.:	Sterling	1.47	3.03	3.05	6.88	7.33	8.18	4.92	2.47	2.51	1.33	1.89	1.41
Ocean station:	Echo	2.94	2.04	3.53	11.41	5.32	NS	4.03	2.17	[a]1.82	2.27	1.87	2.45
Fla.:	Miami	2.30	2.92	5.92	11.57	8.36	3.36	3.61	2.03	1.17	2.22	3.08	NS
Bahamas:	Bimini	3.75	6.08	7.89	15.56	8.37	NS	4.22	2.70	1.66	2.67	4.74	3.50
Hawaii:	Mauna Loa	5.03	4.95	5.51	8.73	6.22	6.09	NS	2.16	1.75	1.31	NS	1.74
P.R.:	San Juan	[a]1.90	2.34	5.24	7.12	4.21	4.49	3.81	2.14	[a]1.16	[a].77	[a]1.80	2.02
Panama:	Balboa	[a]1.44	2.35	5.32	3.00	2.57	.58	1.12	[a].66	[a].41	[b]	[a].37	[a].86
Ecuador:	Guayaquil	[a]2.96	3.50	4.44	2.63	2.51	3.56	3.30	[a]1.98	4.36	[b]	8.90	4.58
Peru:	Lima	17.70	15.72	8.08	5.56	.51	[a]6.73	8.38	10.00	10.97	14.64	10.88	7.48
Bolivia:	Chacaltaya	5.17	2.93	1.58	4.27	2.36	10.81	5.55	13.08	5.52	8.93	[a]5.80	3.53
Chile:	Antofagasta	19.09	13.47	8.19	7.57	5.96	16.74	11.73	13.15	8.69	11.02	11.01	9.60
	Portillo	30.19	21.26	13.37	13.61	17.53	NS	NS	NS	NS	NS	NS	NS
	Santiago	21.78	[a]16.72	18.89	14.23	6.30	17.66	10.85	8.58	10.67	12.15	12.58	10.75
	Puerto Montt	10.99	13.52	11.68	10.07	6.87	NS	7.69	4.34	2.94	4.15	NS	5.60
	Punta Arenas	7.70	7.14	8.69	5.49	4.26	NS	2.29	2.54	3.04	[a]2.36	[a]2.18	4.48

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

Table 10. Plutonium-239 concentrations in surface air, January–December 1967

Site		Concentration (dpm/1,000 standard m³ at midmonth) ×100											
		Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Sept.	Oct.	Nov.	Dec.
Greenland:	Thule	9.11	10.00	13.72	14.27	7.74	NS	5.27	3.08	1.56	[a]1.70	2.19	4.32
Ocean station:	Bravo	1.92	5.09	3.97	5.68	5.07	NS	2.57	1.72	2.13	1.14	1.03	1.72
	Charlie	NS	5.75	5.69	7.65	6.25	NS	NS	1.95	2.06	3.83	1.16	2.63
Ontario:	Moosonee	3.38	7.60	10.28	11.36	12.45	13.95	6.89	4.84	4.53	2.79	NS	2.39
Wash.:	Seattle	[a]5.42	8.57	[a]6.78	6.89	NS	NS	6.97	NS	NS	NS	NS	NS
Ocean station:	Delta	6.23	5.76	8.14	9.66	11.85	NS	4.42	5.20	6.12	[a]2.97	NS	NS
N.J.:	Westwood	4.64	9.10	9.04	14.68	NS	NS	7.28	NS	NS	NS	NS	NS
N.Y.:	New York	8.80	11.83	11.97	15.24	14.21	34.36	8.21	9.35	7.24	6.66	3.73	4.09
Va.:	Sterling	6.85	7.62	9.95	18.03	12.67	10.98	7.54	4.27	4.17	2.49	1.86	2.97
Ocean station:	Echo	8.30	7.11	9.95	31.81	9.09	NS	6.94	3.77	3.13	3.10	2.60	3.51
Fla.:	Miami	9.37	13.20	15.79	26.67	19.10	23.15	6.20	3.56	2.07	4.14	5.34	NS
Bahamas:	Bimini	11.17	18.17	19.43	35.59	18.28	NS	7.05	4.10	2.53	5.01	5.92	5.18
Hawaii:	Mauna Loa	17.10	15.12	10.17	18.56	12.13	10.48	NS	3.05	3.22	2.22	NS	4.84
P.R.:	San Juan	5.05	5.55	12.23	17.04	8.82	8.10	6.74	5.57	3.49	1.85	3.03	3.81
Panama:	Balboa	4.07	6.77	8.82	4.96	4.26	.54	1.95	1.60	.99	[b]	[a].59	1.39
Ecuador:	Guayaquil	[a]2.12	[a]1.69	2.27	1.22	1.94	15.15	10.86	4.88	5.83	[a]3.59	[a]5.52	3.37
Peru:	Lima	11.79	6.94	3.52	2.00	2.28	11.65	25.02	14.77	11.40	13.92	9.82	5.02
Bolivia:	Chacaltaya	3.07	[a].91	.68	2.09	[a]1.41	27.48	NS	35.62	11.01	8.03	[a]6.33	1.97
Chile:	Antofagasta	9.81	7.51	3.26	3.01	2.88	12.73	31.44	18.76	9.72	9.88	9.85	5.71
	Portillo	17.05	7.60	4.84	5.99	7.57	NS	NS	NS	NS	NS	NS	NS
	Santiago	13.97	[a]16.79	7.87	5.22	[a]3.70	8.91	7.70	13.37	14.89	10.24	10.88	7.90
	Puerto Montt	7.55	7.41	4.60	3.82	2.10	NS	4.73	3.79	4.15	4.11	NS	3.93
	Punta Arenas	3.93	3.05	3.74	1.78	1.68	NS	1.05	1.70	[a]2.90	[a]1.81	[a]2.03	3.12

Errors are less than 20 percent except:
[a] error between 20–100 percent.
[b] error greater than 100 percent.
NS, no sample.

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Table 11. Quality control results on blank samples—1967 (average values in dpm per sample)

Month	Iron-55	Strontium-89	Strontium-90	Zirconium-95	Cadmium-109	Cerium-141	Cerium-144	Plutonium-238	Plutonium-239
January	0	0	0	0	0	0	4.7	0	0
February	0	0	0	0	0	0	0	0	0
March	0	0	0	0	0	0	0	0	0
April	0	0	0	0	0	0	0	0	0
May	604	<380	< 3	534	NS	0	1.7	.01	0
June	NS	<300	70	< 30	NS	<300	7	.01	.01
July	44	4.2	.9	3.5	3.7	12.6	.1	0	0
August	81	NS	NS	6.7	0	84	5.4	0	.03
September	48	1.8	.5	108	0	NA	1.5	.01	−.02
October	96	39	1.5	6.1	0	NA	.3	.03	.03
November	37	24	2.1	.5	0	NA	2.4	.30	.20
December	7.1	13	1.9	8.1	NA	6.9	.4	.17	.02

NA, no analysis.
NS, no sample.

Table 12. Quality control results on standard samples—1967 (average values in percent deviation)

Month	Iron-55	Strontium-89	Strontium-90	Zirconium-95	Cadmium-109	Cerium-141	Cerium-144	Plutonium-238	Plutonium-239
January	−21	− 7	0	−11	−39	− 7	+ 6	− 8	− 2
Feburary	− 2	+20	+70	− 7	−32	− 4	− 7	−13	−16
March	+ 5	+ 9	− 5	−12	− 8	+28	+ 4	− 2	− 1
April	+19	+14	− 8	−15	−17	+113	0	− 5	− 6
May	+ 1	+14	NA	−48	−99	− 45	− 5	−15	−13
June	+32	−12	+19	+ 8	−16	−100	−29	− 5	− 7
July	+11	+ 3	− 2	−12	−31	− 52	+ 8	−16	−14
August	− 3	+ 5	−18	−50	−33	− 27	+ 2	− 8	− 8
September	−15	− 4	−19	−40	−27	NA	+ 3	− 6	− 4
October	−13	+29	−14	− 4	−25	NA	+ 2	− 6	−14
November	+18	− 5	− 3	+20	−26	NA	− 3	+37	−11
December	+ 9	−14	−14	−16	NA	NA	+ 2	− 1	− 5

NA, no analysis.

Table 13. Quality control results on duplicate samples—1967 (values are ± percent standard deviation)

Month	Iron-55	Strontium-89	Strontium-90	Zirconium-95	Cadmium-109	Cerium-141	Cerium-144	Plutonium-238	Plutonium-239
January	NA	NA	NA	NA	NA	25	1	20	9
February	NA	NA	NA	NA	NA	4	0	17	12
March	16	2	7	NA	NA	NA	5	4	12
April	1	0	1	NA	15	NA	30	0	7
May	5	17	11	19	NA	NA	1	45	29
June	23	NA	NA	36	NA	NA	37	NA	NA
July	65	NA	8	NA	NA	NA	1	5	1
August	22	12	7	NA	NA	63	32	6	19
September	5	NA	2	48	NA	NA	NA	7	5
October	14	15	12	NA	NA	NA	2	25	21
November	25	5	14	2	NA	NA	0	12	11
December	NA	NA	NA	NA	NA	NA	NA	27	19

NA, no analysis.

with regular samples. Basically there are three kinds of quality control samples: blanks, standards and duplicates.

A blank is prepared by sprinkling a small amount of pre-1945 soil (10 mg) onto an unexposed microsorban filter. Then carbon soot from burning naphthalene is filtered onto the microsorban paper for a few moments under laboratory conditions. This procedure produces blanks which are very similar in appearance, and in some respects, composition to the routine filters collected by this program. When the blanks are coded, as are all the filters submitted to radiochemistry, they are not readily distinguishable from routine samples. Table 11 lists the results of analyses of blanks in 1967. The table summarizes the average dpm

reported from blanks sent out each month. It is evident from a review of these data that the contamination problem has been quite variable, and consequently low level concentrations in the real samples must be considered with care.

A standard is prepared at HASL by evaporating a known amount of radioactivity onto a blank filter. A few drops of wetting agent are applied to the blank prior to the addition of the standard to permit permeation of the solution between the polyethylene fibers of the microsorban filter. Table 12 presents the results of these analyses for each nuclide in 1967. The values are average percent deviations of the reported analyses from the known. Again the results are seen to be extremely variable and generally non-systematic.

Duplicates consist of collections made monthly at the New York site. Prior to December 1967 samples were composited and split before analyses thereby providing a measure of the reproducibility of the radiochemical procedures. The results of these tests are shown as the percent standard deviation from the mean, in table 13.

Recent coverage in *Radiological Health Data and Reports*:

Period	Issue
January–December 1966	October 1970

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SECTION IV. OTHER DATA

This section presents results from routine sampling of biological materials and other media not reported in the previous sections. Included here are such data as those obtained from human bone sampling, Alaskan surveillance, and environmental monitoring around nuclear facilities.

Strontium-90 in Human Bone, January-March 1970[1]

Bureau of Radiological Health
U.S. Public Health Service

To obtain data on the concentration of strontium-90 in man by age and geographical region, the Public Health Service began collecting human bone specimens in late 1961. Analyses of selected samples of people in older-age groups have shown their bone strontium-90 content to be low and age independent (*1*). Consequently, the target population includes children and young adults up to 25 years of age.

Although a few samples come from living persons as a result of surgical procedures, the majority are obtained post mortem. In the latter case, the specimens are limited to accident victims or persons who have died of an acute disease process that was not likely to impair bone metabolism. For analytical purposes, a sample of at least 100 grams of wet bone is desired. Generally, this amount is readily available from older children, but it presents some difficulties from the standpoint of infants and children under 5 years of age. Most specimens received to date have been vertebrae and ribs.

[1] Period during which death or surgical procedure occurred.

Laboratory procedures

The bones are analyzed at the Northeastern Radiological Health Laboratory of the Bureau of Radiological Health at Winchester, Mass. Sample collection and preparation are explained elsewhere (*2*). Strontium-90 is measured by tributyl phosphate extraction of its yttrium daughter, which is precipitated as an oxalate. The strontium-90 content is then calculated (*3*) from the yttrium-90 activity. For the purpose of maintaining analytical

Figure 1. Geographical regions for human bone sampling

Table 1. Strontium-90 in human bone, January–March 1970

Bone region and State	Bone type[a]	Age[b] (years)	Sex	Strontium-90 concentration[c] (pCi/kg bone)	Calcium concentration (g/kg bone)	$^{90}Sr/Ca$ (pCi/g)
Northeast:						
Pennsylvania	V	1	F	176.0±17.0	37.3	4.71
	V	2	M	78.9± 9.1	32.3	2.44
Massachusetts	V	3	M	94.2± 8.4	28.5	3.30
Pennsylvania	V	3	F	47.2± 6.9	32.8	1.43
New Hampshire	V	3	M	136.0±13.0	41.1	3.30
New York	V	3	M	55.3± 9.8	29.1	1.90
Pennsylvania	V	5	M	55.1± 6.6	21.8	2.52
Massachusetts	V	5	M	96.9± 9.9	31.1	3.11
Rhode Island	S	5	M	33.3± 6.5	20.6	1.61
Massachusetts	V	5	M	58.2± 8.1	20.9	2.78
New York	V	6	M	41.3± 7.2	23.4	1.76
	V	6	F	50.6± 6.3	28.7	1.76
Massachusetts	V	6	M	72.9±10.0	29.6	2.46
New Jersey	V	6	M	57.3± 5.9	26.2	2.18
Pennsylvania	V	9	M	57.9± 6.5	38.5	1.50
New York	V	10	F	52.0± 6.9	27.0	1.92
	V	11	M	64.1± 7.1	32.6	1.96
	V	12	F	103.0±12.0	72.5	1.42
	V	12	F	153.0±13.0	60.6	2.52
	V	13	M	71.8± 8.5	35.8	2.00
Pennsylvania	V	13	M	98.3±13.0	56.6	1.73
	V	13	F	106.0± 8.6	53.6	1.97
	V	14	M	72.8± 9.3	47.0	1.54
New York	V	15	F	118.0±11.0	57.4	2.05
Pennsylvania	V	15	F	147.0±18.0	61.7	2.38
	V	16	M	180.0±13.0	64.0	2.81
	V	16	M	77.1± 8.7	44.7	1.72
	V	16	M	102.0±12.0	52.7	1.93
New Jersey	V	16	M	105.0± 8.6	39.4	2.66
Pennsylvania	V	16	M	135.0±11.0	46.8	2.88
	V	18	M	113.0±10.0	54.1	2.08
Massachusetts	V	18	M	130.0± 8.3	39.9	3.25
New York	V	19	F	112.0± 9.3	44.3	2.52
Massachusetts	V	19	F	90.1±11.0	59.1	1.52
Pennsylvania	V	19	M	114.0±12.0	47.8	2.38
	V	19	M	126.0±10.0	57.2	2.20
New York	V	20	F	112.0±13.0	52.9	2.11
Massachusetts	V	20	M	125.0±12.0	49.1	2.54
New York	V	21	M	86.4± 8.4	53.3	1.62
Pennsylvania	V	21	M	106.0±12.0	54.3	1.95
	V	22	M	120.0±13.0	59.4	2.02
	V	23	M	45.7± 6.4	49.9	.91
Massachusetts	V	23	M	147.0±11.0	32.7	4.49
Pennsylvania	V	23	M	72.7± 9.4	54.2	1.34
Pennsylvania	V	24	F	51.3± 6.9	31.4	1.63
Southeast:						
South Carolina	V	0	F	92.7±10.0	26.3	3.52
	V	1	F	92.9± 9.9	25.8	3.60
Maryland	V	4	M	89.5±12.0	36.8	2.43
North Carolina	V	7	M	76.1± 8.4	27.6	2.75
Maryland	V	9	F	64.1± 9.1	33.2	1.93
South Carolina	V	10	M	91.3±11.0	42.2	2.16
Tennessee	V	15	M	78.8±10.0	49.7	1.58
South Carolina	V	16	M	94.8± 8.3	52.4	1.80
Alabama	V	18	M	113.0±12.0	46.2	2.44
Maryland	V	19	M	113.0±11.0	62.3	1.81
	V	20	M	78.8±10.0	58.5	1.34
South Carolina	V	21	M	77.3± 8.8	37.8	2.04
Maryland	V	21	F	81.0±10.0	55.8	1.45
South Carolina	V	23	F	70.5± 9.1	45.7	1.54
Maryland	V	24	M	85.3±11.0	48.5	1.75
	V	24	F	91.3±11.0	47.7	1.91
South Carolina	V	25	M	50.8± 8.0	43.3	1.17
Maryland	V	25	M	98.1±12.0	43.1	2.27
Central:						
Wisconsin	V	0	F	61.7± 7.5	33.5	1.84
	V	1	M	54.4± 8.1	28.7	1.89
Ohio	V	1	M	63.1± 8.6	30.3	2.08
Wisconsin	V	1	M	45.9± 8.0	28.3	1.62
	V	1	F	43.5± 9.1	26.1	1.66
	V	2	M	76.7± 9.0	36.8	2.08
	V	3	M	42.6± 6.1	27.2	1.56
Nebraska	V	3	F	72.1±10.0	30.7	2.34
Ohio	V	3	M	106.0±12.0	35.3	3.00
	V	3	F	127.0±10.0	35.0	3.62
Minnesota	V	6	M	41.1± 6.4	34.3	1.19
Michigan	V	6	M	75.4±11.0	34.0	2.21
	V	7	F	108.0±10.0	47.5	2.27
Minnesota	V	7	M	103.0±10.0	31.9	3.22
Ohio	V	8	M	95.2± 9.9	39.5	2.41
	V	8	F	39.3± 5.0	31.8	1.23
	V	8	M	67.5± 8.1	30.4	2.22
	V	9	M	59.5± 9.9	42.1	1.41
	V	11	F	86.2± 8.5	43.8	1.96
Minnesota	V	12	F	92.1± 7.2	39.6	2.32

See footnotes at end of table.

Table 1. Strontium-90 in human bone, January–March 1970—Continued

Bone region and State	Bone type[a]	Age[b] (years)	Sex	Strontium-90 concentration[c] (pCi/kg bone)	Calcium concentration (g/kg bone)	$^{90}Sr/Ca$ (pCi/g)
Central: (Continued)						
Ohio	V	12	M	70.8± 7.4	37.4	1.89
Michigan	V	13	F	79.8± 9.3	49.7	1.60
Ohio	V	13	M	62.8± 6.1	37.8	1.66
	V	13	F	162.0± 9.8	48.1	3.36
	V	15	M	63.1± 8.0	46.7	1.35
	V	15	M	126.0±11.0	57.1	2.20
	V	16	M	102.0±12.0	51.1	1.99
	V	17	M	96.3±11.0	60.8	1.58
Wisconsin	V	17	M	101.0±10.0	57.4	1.75
Ohio	V	17	F	93.0±10.0	48.8	1.90
	V	17	F	81.9±10.0	51.5	1.59
	V	17	M	83.5±11.0	61.2	1.36
	V	19	F	78.7±10.0	53.9	1.46
	V	19	M	72.1±11.0	67.4	1.06
	V	20	F	73.4± 9.1	48.9	1.50
	V	20	M	80.0± 9.7	52.9	1.51
	V	21	F	90.1± 8.0	44.4	2.02
	V	21	F	101.0± 9.1	63.8	1.58
	V	21	M	137.0±14.0	60.2	2.27
	V	22	M	82.7± 8.2	53.3	1.55
	V	23	M	101.0±11.0	62.2	1.62
Delta:						
Louisiana	V	3	M	75.5± 8.5	33.2	2.27
	V	17	F	135.0±11.0	48.4	2.78
	V	20	M	119.0±14.0	76.1	1.56
	V, R	24	F	119.0±14.0	69.5	1.71
Northwest:						
Oregon	V	2	M	50.3± 7.6	30.8	1.63
	V	2	M	34.7± 6.6	29.8	1.16
	V	6	M	51.5± 7.0	32.7	1.57
	V	8	M	54.7± 7.0	30.0	1.82
	V	12	M	50.3± 6.6	35.8	1.40
Washington	V	16	M	55.5± 8.4	40.2	1.38
Oregon	V	21	M	79.5±10.0	56.5	1.40
	V	21	M	47.2± 6.4	43.2	1.09
Southwest:						
Texas	V	21	F	91.0±10.0	51.1	1.78
Utah	V	23	M	46.8± 5.3	33.3	1.40
North:						
Alaska	V	0	M	101.0±11.0	32.7	3.08
	R	15	M	419.0±36.0	115.0	3.64

[a] V—vertebra; S—sternum; R—rib.
[b] Age given as of last birthday prior to death.
[c] Two sigma counting error.

reproducibility, "blind" duplicate analyses are performed on 10 to 20 percent of the samples.

The analytical results for strontium-90 in individual bones from persons dying during the first quarter (January–March) of 1970 are presented in table 1 in order of increasing age within each geographical region. These regions are indicated in figure 1. Reported values are given in picocuries of strontium-90 per kilogram of bone (as a rough indication of dose) and per gram of calcium (for comparison with other data and for purposes of model development). Two-sigma counting errors are reported for the bone concentration.

Following the pattern of earlier reports, subsequent articles will continue to provide interpretation of the data at appropriate stages in the program (*2–6*).

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Environmental Levels of Radioactivity at Atomic Energy Commission Installations

The U.S. Atomic Energy Commission (AEC) receives from its contractors semiannual reports on the environmental levels of radioactivity in the vicinity of major Commission installations. The reports include data from routine monitoring programs where operations are of such a nature that plant environmental surveys are required.

Releases of radioactive materials from AEC installations are governed by radiation standards set forth by AEC's Division of Operational Safety in directives published in the "AEC Manual."[1]

Summaries of the environmental radioactivity data follow for the Argonne National Laboratory and the National Reactor Testing Station.

[1] Title 10, Code of Federal Regulations, Part 20, "Standards for Protection Against Radiation" contains essentially the standards published in Chapter 0524 of the AEC Manual.

1. Argonne National Laboratory July-December 1969

University of Chicago
Lemont, Ill.

The radioactivity of the environment is determined on a continuing basis by measuring the radioactivity in naturally-occurring materials collected both on and off the Argonne National Laboratory site. Since radioactivity is usually spread by air and water, the environmental monitoring program at Argonne has concentrated on these materials. The sampling locations discussed in this report are shown in figures 1 and 2.

Air monitoring

The radioactivity of particulate matter in air was determined by drawing air through filter paper at a known rate and measuring the radioactivity of the particles collected by the paper. The samples were collected continuously at seven locations on the Argonne site and at five locations off the site. At one location on the site the filter paper was changed daily; at all other locations the filter papers were changed at weekly intervals. The daily samples record short-term changes in radioactivity, while the weekly samples are used to compare onsite and offsite activities. Higher activities on the site are indicative of radioactivity released by Argonne if the differences are greater than the 10 to 20 percent error in sampling and counting. The alpha and beta radioactivity in air-filter samples are summarized in table 1; the average concentration of gamma-ray emitters as determined by gamma ray spectrometry are summarized in table 2. The alpha radioactivity concentrations were due principally to naturally-occurring radionuclides and were in the range found in previous years. As in the past, much of the beta and gamma radioactivity was due to fission and neutron activation products from nuclear detonations, although about one-third of the gamma radioactivity and a smaller fraction of the beta radioactivity was due to beryllium-7, produced continuously in the upper atmosphere by cosmic-ray interactions. The total beta radioactivity and all the gamma-ray emitters for which positive results were obtained, including the naturally-occurring beryllium-7, increased steadily from January to August 1969, then decreased during the remainder of the year. This behavior is attributed to the usual "spring maximum" in stratospheric fallout, although the peak was broader than in the past several years. Most of the fission products in the samples were injected into the atmosphere several months before they were observed, probably about December 1968, since the shortest-lived fission product detected in the spring was cerium-141 and the most abundant fission products were zirconium-95 and niobium-95. The presence of a very small amount of barium-140 ($\sim$1 fCi/m^3) in the air during the middle of January supports the suggestion that most of the fission products were produced in December 1968. The delay in its appearance implies high-altitude injection, or possibly an increased spring fallout rate from the troposphere by rain washout.

The average beta radioactivity for the year (0.23 pCi/m^3) was about 10 percent higher than

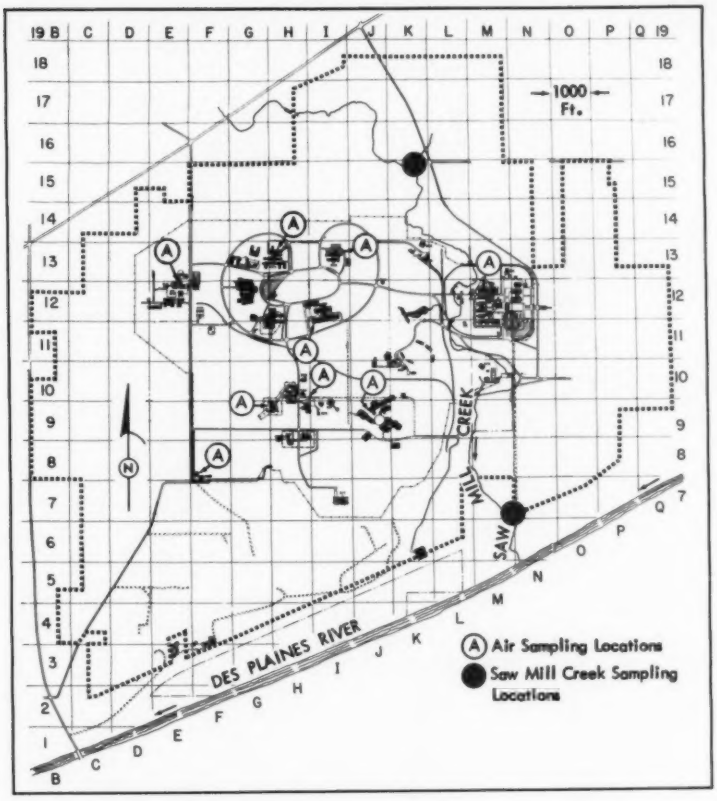

Figure 1. Onsite sampling locations at Argonne National Laboratory

during 1968. Most of the gamma-ray emitters were also present in slightly higher concentrations this year.

Precipitation collected on the site contained the same fission products found in air samples. Composite large volume rain samples were also analyzed monthly for iron-55. The average monthly concentrations ranged from 9 to 890 pCi/m² of surface. The average for the year (180 pCi/m²) was a factor of three greater than the 1968 average. However, if the unusually large value of 890 pCi/m² in June 1969 is excluded, the average decreased to 110 pCi/m². Iron-55 is a neutron activation product and was produced mainly in nuclear tests conducted in September and October 1961. Therefore, the above concentrations have been corrected for decay to October 15, 1961, for comparison purposes. Since iron-55 decays by electron capture and does not emit any gamma-

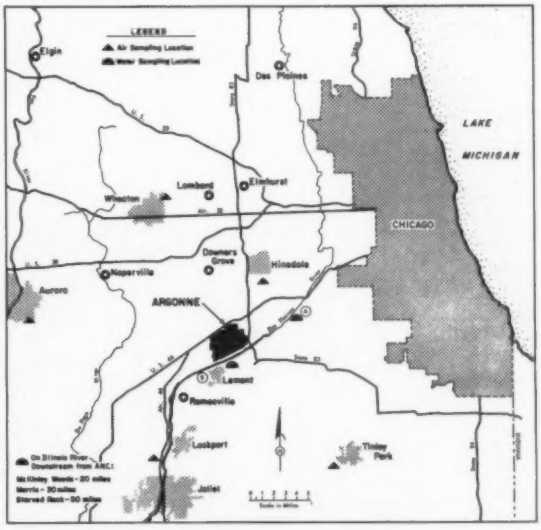

Figure 2. Locations of Argonne National Laboratory (including some offsite sampling stations)

Table 1. Alpha and beta radioactivity in air-filter samples,[a] Argonne National Laboratory, July–December 1969

(1969)	Location	Number of samples	Alpha radioactivity (fCi/m³)			Beta radioactivity (pCi/m³)		
			Average	Minimum	Maximum	Average	Minimum	Maximum
July	Onsite	23	3.0	1.6	5.5	0.38	0.20	0.61
	Offsite	20	4.9	1.6	12.9	.41	.15	.57
August	Onsite	26	4.1	.3	14.5	.51	.13	.80
	Offsite	21	4.8	2.6	15.0	.54	.19	.80
September	Onsite	23	3.5	1.0	7.8	.19	.09	.43
	Offsite	23	5.0	2.5	7.9	.24	.09	.42
October	Onsite	27	3.2	1.1	7.2	.13	.06	.21
	Offsite	22	3.8	1.5	9.4	.16	.10	.26
November	Onsite	25	3.0	1.5	8.1	.08	.04	.12
	Offsite	18	3.4	2.2	5.8	.09	.04	.13
December	Onsite	24	2.8	1.4	4.4	.07	.03	.11
	Offsite	22	3.8	1.6	9.9	.08	.05	.11
Annual summary	Onsite	300	3.3	0.3	14.5	.21	0.03	0.80
	Offsite	255	4.1	0.5	15.0	.24	0.04	0.80

[a] These results were obtained by measuring the samples four days after they were collected in order to avoid counting the natural radioactivity due to radon and thoron decay products.

Table 2. Gamma-ray radioactivity in air-filter samples, Argonne National Laboratory, July–December 1969

Radionuclide	Location	Concentration (pCi/m³)						
		July	August	September	October	November	December	Annual summary
Barium–lanthanum–140	Onsite	<0.01	<0.01	<0.01	<0.01	<0.01	<0.01	<0.01
	Offsite	< .01	< .01	< .01	< .01	< .01	< .01	< .01
Beryllium–7	Onsite	.11	.13	.11	.07	.07	.06	.09
	Offsite	.10	.14	.11	.07	.05	.05	.08
Cerium–141	Onsite	.01	.01	< .01	< .01	< .01	< .01	.01
	Offsite	.01	.01	< .01	< .01	< .01	< .01	.01
Cerium–144	Onsite	.05	.07	.03	.02	.01	.01	.03
	Offsite	.06	.06	.03	.02	.01	.01	.03
Cesium–137	Onsite	< .01	< .01	< .01	< .01	< .01	< .01	< .01
	Offsite	< .01	< .01	< .01	< .01	< .01	< .01	< .01
Iodine–131	Onsite	< .01	< .01	< .01	< .01	< .01	< .01	< .01
	Offsite	< .01	< .01	< .01	< .01	< .01	< .01	< .01
Ruthenium–103	Onsite	.02	.02	< .01	< .01	< .01	< .01	.01
	Offsite	.02	.02	< .01	< .01	< .01	< .01	.01
Ruthenium–rhodium–106	Onsite	.09	.09	.06	.04	.02	.02	.05
	Offsite	.08	.09	.07	.05	.02	.02	.05
Zirconium–niobium–95	Onsite	.15	.17	.07	.03	.02	.02	.06
	Offsite	.17	.18	.07	.03	.02	.01	.06

Table 3. Argon-41 concentrations, near Juggernaut Reactor, July–December 1969

Month (1969)	Number of samples	Number of samples > 20 nCi/m³	Concentration[a] (nCi/m³)		Percent of AEC standard	
			Average	Maximum	Average	Maximum
July	30	8	24	152	60	380
August	28	2	12	40	30	100
September	16	5	25	77	63	193
October	16	2	10	26	25	65
November	8	1	15	49	38	123
December	24	4	22	106	45	265
Annual summary	202	56	22	152	45	380

[a] Minimum detectable concentration, 20 nCi/m³.

rays, this nuclide does not contribute to the beta and gamma activities presented in tables 1 and 2. The neutron activation products, tungsten-181 and 185, were also detected in some precipitation samples.

Air was also sampled continuously for radioiodine with activated charcoal in the 200 and 300 Areas because of the possibility of iodine releases in these areas. No radioactivity above the minimum detectable concentration of 0.1 pCi/m³ was detected at either location.

Air sampling for argon-41 (a beta-particle emitter with a 1.8 hour half-life that is produced

in an operating reactor by the action of neutrons on the stable argon-40 in air) was conducted near the Juggernaut reactor (Building 335) from March through December. The results are tabulated in table 3. These concentrations are based on one-minute grab samples taken at a position downwind from the reactor at a point favorable for argon-41 detection after the reactor has been in operation for several hours. Therefore, the actual average concentrations in the 300 Area were much less than the values in the table. The argon-41 concentration in the exhaust from the Juggernaut is about 260 $\mu Ci/m^3$, so the average dilution by outside air in the samples collected was about a factor of 10^4. In 1968, argon-41 concentrations were measured in the vicinity of the CP-5 reactor (Building 330), which operates at a power level about 20 times higher than Juggernaut. The CP-5 reactor is undergoing rehabilitation and is not being operated this year. A comparison of the argon-41 concentrations near the two reactors is of interest. In 1968, argon-41 was detected in about 46 percent of the samples collected near CP-5. The average concentration was 140 nCi/m^3. The fraction of positive results near the Juggernaut reactor in 1969 was appreciably less (28 percent), while the average concentration (22 nCi/m^3) was six times smaller. In the CP-5 reactor the air that is exposed to neutrons is diluted with a large volume of inactive air before it is discharged. This means that the argon-41 concentrations in the air leaving the two reactor buildings are not greatly different, although the volume of air (containing argon-41) is about 15 times greater from CP-5. As a result, the argon-41 concentration in outside air averaged only six times less than the Juggernaut, although the argon-41 production is about 20 times less than CP-5.

Air was sampled for tritiated (hydrogen-3) water vapor 50-yards west of the CP-5 reactor. The results are given in table 4. The results were positive in all but a few samples, although the concentrations were well below the AEC standards. This nuclide is produced continuously while the reactor is in operation by the action of neutrons on the heavy water used for moderating and cooling. The average concentration this year (400 pCi/m^3) was about three times higher than the 1968 average. Eighteen samples were also collected between May and December 1969 in the east area, 1,829 meters (2,000 yards) from the reactor.

Table 4. Tritium concentrations near CP-5 reactor July-December 1969

Month (1969)	Number of samples	Concentration (pCi/m^3)			Percent of AEC standard	
		Average	Minimum	Maximum	Average	Maximum
July	3	105	89	133	0.05	0.07
August	11	190	109	507	.10	.25
September	11	210	<10	1,804	.11	.90
October	4	95	30	175	.05	.09
November	7	754	41	1,525	.38	.76
December	10	2,330	392	5,750	1.17	2.88
Annual summary	122	400	<10	5,750	0.20	2.88

In these samples, the tritium concentration ranged from less than 10 to 45 pCi/m^3 and averaged 15 pCi/m^3. Samples collected near the reactor at the same time averaged 660 pCi/m^3. The east area samples generally varied in the same manner as the activity in the samples near CP-5. For example, the highest concentrations at both locations, (45 and 5,750 pCi/m^3) were found in samples collected on the same day, and east area samples, (greater than 10 pCi/m^3) were found only on those days when the CP-5 samples contained more than 90 pCi/m^3. The correlation is sufficiently good to conclude that most of the tritium in the east area samples came from the CP-5 reactor. Although the reactor has not been in operation since January 1969, the heavy water containing the long-lived tritium was handled while the reactor was dismantled and small quantities evidently did escape from the building.

Water monitoring

Argonne wastewater is discharged into Sawmill Creek, a small stream that runs through the Argonne grounds and enters the Des Plaines River about 500 yards downstream from the wastewater discharge. Sawmill Creek was sampled upstream from the Argonne site and downstream from the wastewater outfall to determine if radioactivity was added to the stream in Argonne wastewater. The sampling locations are shown in figure 1.

Below the wastewater outfall the creek was usually sampled three to five times weekly. Since it was impractical to analyze all the samples for all the radionuclides and elements desired, equal portions of the individual samples collected each week were combined and analyzed. The results

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Table 5. Non-volatile alpha and beta radioactivity in Sawmill Creek water, Argonne National Laboratory, July–December 1969

Month (1969)	Location[a]	Number of samples	Alpha radioactivity (pCi/liter)			Beta radioactivity (pCi/liter)		
			Average	Minimum	Maximum	Average	Minimum	Maximum
July	Upstream	2	3.3	2.1	4.5	20	15	25
	Downstream	15	1.4	.2	3.0	16	13	21
August	Upstream	2	2.4	1.7	3.2	15	11	19
	Downstream	24	2.4	1.6	3.4	21	14	40
September	Upstream	2	6.2	6.1	6.2	32	29	34
	Downstream	20	2.1	1.3	3.6	19	16	21
October	Upstream	2	1.9	1.4	2.3	22	14	30
	Downstream	25	1.5	.2	3.5	18	13	25
November	Upstream	2	2.5	2.1	2.9	14	11	17
	Downstream	18	1.9	1.5	2.2	13	12	14
December	Upstream	2	2.3	2.2	2.3	6.3	6.1	6.4
	Downstream	23	2.7	1.2	4.5	14	13	15
July–December summary	Upstream	12	3.1	1.4	6.2	18	6.1	34
	Downstream	121	2.0	.2	4.5	17	12	40
Annual summary	Upstream	24	3.0	1.4	6.2	16	6.1	34
	Downstream	214	2.2	.2	4.5	15	8.7	40

[a] Relative sampling location with respect to Argonne wastewater outfall (figure 1).

obtained in this way represent the average concentrations in the weekly samples. Above the site, samples were collected twice monthly and at least one sample each month was analyzed for each radionuclide of interest. The total alpha and beta radioactivities found in Sawmill Creek during July–December 1969 are given in table 5. Upstream from the Argonne site, the alpha activity was due primarily to radioactive nuclides that occur naturally in the stream, and was in the normal range. Any additional activity downstream was due to alpha-emitting nuclides in the wastewater. The concentration of alpha activity in the creek resulting from the presence of Argonne wastewater can be estimated as follows.

Below the outfall, the Argonne wastewater was diluted from one to ten times by creek water during July–December 1969. This increased dilution above previous years is due to the discharge into the creek of water from a new public sewage disposal plant north of the laboratory. Consequently, the natural alpha activity in the creek below the outfall varied from 50 to 90 percent of the upstream concentration and the alpha activity added by Argonne wastewater varied from 0.7 to less than 0.1 pCi/liter. Although Argonne wastewater contained a small quantity of alpha activity, its addition to the creek did not increase the total alpha activity in the stream. On some days, when below-outfall water contained less alpha activity than upstream water, the wastewater evidently contained less alpha activity than natural creek water. The unusually high alpha activity for this location in the September upstream was accompanied by an unusually large amount of dissolved solids. These two samples increased the above site annual average from 2.6 to 3.0 pCi/liter.

The alpha-particle emitters most likely to be present in Argonne wastewater are isotopes of uranium, plutonium, and thorium. The alpha radioactivity in the creek water due to these elements are summarized in table 6. The concen-

Table 6. Alpha-emitting elements in Sawmill Creek water, Argonne National Laboratory, July–December 1969

Element	Location[a]	Number of samples	Concentration (pCi/liter)			Percent of AEC standard	
			Average	Minimum	Maximum	Average	Maximum
Uranium	Upstream	6	2.0	0.9	3.4	0.005	0.009
	Downstream	57	1.6	.7	2.9	.004	.007
Plutonium	Upstream	6	—	—	< .05	—	< .001
	Downstream	58	—	—	< .05	—	< .001
Thorium	Upstream	6	—	—	< .05	—	< .003
	Downstream	58	—	—	< .05	—	< .003

[a] Relative sampling location with respect to Argonne wastewater outfall (figure 1).

trations are low compared to the AEC standards. The uranium content attributable to Argonne below the outfall ranged from 0.6 to less than 0.1 pCi/liter, depending on the dilution (about the same as during 1968). The amount of uranium in the wastewater, however, was too low to increase the concentration in the creek water, and, as was the case for total alpha radioactivity, on some days the wastewater contained less uranium than the creek water. Comparison of the data in table 6 with the total alpha radioactivity indicates that most of the alpha activity in the creek was due to uranium. The alpha to uranium ratio (about 1.5) at both locations is normal for surface water in the area and the remaining alpha activity is usually due to uranium decay products.

In addition to the natural beta radioactivity in the creek, beta activity from nuclear detonations was detected at both sampling locations and beta radioactivity from Argonne wastewater was found in some samples below the outfall. The natural beta radioactivity is approximately 5 pCi/liter above the site and 2 to 5 pCi/liter below the site depending on the amount of dilution by wastewater. It is estimated that fallout activity added about 10 pCi/liter to the nonvolatile beta radioactivity at both locations (fallout is added to the water below the outfall directly from the air as well as by above-site water) and the Argonne contribution below the outfall averaged 2 pCi/liter. The Argonne contribution decreased to one-third of the 1968 levels, while the fallout contribution remained about the same.

The results of analyses of Sawmill Creek water for specific beta-emitting nuclides are summarized in table 7. The principal beta-particle emitter added to the creek by Argonne wastewater was tritium. This radionuclide, present as water (HTO), is not included in the total beta activity in table 5. Separate samples were analyzed specifically for tritium. Although small amounts are produced by cosmic-ray interactions with the atmosphere, the bulk of the tritium in above-site water originated in nuclear detonations. Although some of the other nuclides listed were also added to the creek by fallout, the total concentration regardless of source, must be used in assessing the health hazard of a radioactive nuclide not naturally present in the environment. The percent of the AEC standard given in the table was calculated on this basis. As indicated in the table, the beta activities in the creek were very low compared to the AEC standards.

Since Sawmill Creek empties into the Des Plaines River, which in turn flows into the Illinois River, the radioactivity in the latter two streams is important in assessing the contribution of Argonne wastewater to the environmental radioactivity. The Des Plaines River was sampled monthly above the mouth of Sawmill Creek and weekly below the mouth to determine if the radioactivity in the creek has any effect on the radio-

Table 7. Beta-emitting radionuclides in Sawmill Creek water, Argonne National Laboratory, July-December 1969

Radionuclide	Location[a]	Number of samples	Beta radioactivity (pCi/liter)			Percent of AEC standard	
			Average	Minimum	Maximum	Average	Maximum
Barium-140	Upstream	6	—	—	< 2	—	<0.007
	Downstream	57	—	—	< 2	—	< .007
Cerium-144	Upstream	12	—	—	<10	—	< .1
	Downstream	121	—	—	<10	—	< .1
Cesium-137	Upstream	6	—	—	< .5	—	< .003
	Downstream	121	1.0	< 0.5	3.1	0.005	.016
Cobalt-58	Upstream	12	—	—	<10	—	< .01
	Downstream	121	—	—	<10	—	< .01
Cobalt-60	Upstream	12	—	—	<10	—	< .02
	Downstream	121	—	—	<10	—	< .02
Tritium	Upstream	6	500	<400	1,000	.02	.03
	Downstream	121	2,400	600	18,600	.08	.62
Iodine-131	Upstream	0	—	—	—	—	—
	Downstream	121	—	—	< 3	—	<1
Strontium-89	Upstream	6	—	—	< 2	—	< .07
	Downstream	57	—	—	< 2	—	< .07
Strontium-90	Upstream	6	1.7	< 0.5	3.4	.6	1.1
	Downstream	57	1.5	.8	1.9	.3	.6
Technetium-99	Upstream	6	—	—	.5	—	< .0002
	Downstream	121	< .5	< .5	1.0	< .0002	.0003
Thorium-protactinium-234	Upstream	6	1.6	.7	2.7	.008	.014
	Downstream	57	1.3	.5	2.3	.007	.012

[a] Relative sampling location with respect to Argonne wastewater outfall (figure 1).

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Table 8. Average radioactivity in Des Plaines and Illinois River water Argonne National Laboratory, 1969

Location	Concentration (pCi/liter)			
	Nonvolatile alpha radioactivity	Uranium	Nonvolatile beta radioactivity	Tritium
Des Plaines River[a] (above Sawmill Creek)	[b] 3.9 (12)	1.9 (3)	14 (12)	<400 (6)
Des Plaines River[c] (below Sawmill Creek)	3.4 (53)	1.9 (13)	15 (53)	2,300 (24)
Illinois River[d]	2.4 (4)	2.1 (2)	8 (4)	<400 (4)

[a] Sampled near Route 45, upstream from the mouth of Sawmill Creek.
[b] Number in parenthesis is number of samples.
[c] Sampled near Lemont, downstream from the mouth of Sawmill Creek.
[d] Sampled at McKinley Woods State Park, Morris, and Starved Rock State Park.

Table 9. Radioactivity in Des Plaines River water, Argonne National Laboratory July–December 1969

Radionuclide	Location[a]	Number of samples	Concentration (pCi/liter)		
			Average	Minimum	Maximum
Barium-lanthanum-140	Upstream	4	—	—	< 2
	Downstream	6	—	—	< 2
Cerium-144	Upstream	6	—	—	< 10
	Downstream	27	—	—	< 10
Cobalt-58-60	Upstream	6	—	—	< 10
	Downstream	27	—	—	< 10
Tritium	Upstream	3	<400	<400	620
	Downstream	12	5,700	<400	11,000
Plutonium (alpha)	Upstream	3	—	—	< .05
	Downstream	6	—	—	< .05
Strontium-89	Upstream	3	—	—	< 2
	Downstream	6	—	—	< 2
Strontium-90	Upstream	3	1.5	1.3	1.8
	Downstream	6	1.9	1.3	2.9
Thorium-protactinium-234	Upstream	2	1.3	1.1	1.4
	Downstream	7	1.6	1.0	2.2
Uranium	Upstream	2	1.8	1.5	2.0
	Downstream	7	2.0	1.2	2.8

[a] Relative sampling location with respect to Argonne wastewater outfall (figure 1).

activity in the river. The total radioactivity is summarized in table 8, and the results of analysis for specific elements and radionuclides are given in table 9. Tritium was the only nuclide whose concentration was higher below the mouth of Sawmill Creek, and the additional activity at this location evidently originated in Argonne wastewater. However, the average and maximum concentrations below the creek amounted to only 0.08 and 1.1 percent, respectively, of the AEC standards. The concentration of tritium in the river above the creek is similar to that found last year in the Des Plaines River and to that found in other streams in the area. This nuclide occurs naturally and is also present in debris from nuclear tests.

The concentrations of the other radionuclides and of the total alpha and beta radioactivities were in their normal ranges at both locations. The annual average beta radioactivity (15 pCi/liter) was about the same as 1968. The natural nonvolatile beta radioactivity in the river is 5 to 10 pCi/liter, and the excess (about 7 pCi/liter) was due to fallout. The total radioactivity in samples of Illinois River water were normal and similar to those found in other streams in the area. No evidence of radioactivity in Argonne was detected. The tritium concentrations in the Illinois River samples were less than 400 pCi/liter.

Radioactivity in milk

Raw milk was collected monthly from a local dairy farm and analyzed for several fission products. Strontium-89, iodine-131, and barium-140 were not present in concentrations greater than the minimum detectable amounts of 20 pCi/liter for iodine-131 and 5 pCi/liter for the other two

radionuclides. The strontium-90 and cesium-137 concentrations are given in table 10. These two radionuclides are long-lived fission products from past nuclear tests and their presence in milk is not related to Argonne operations. The average strontium-90 content changed little compared to last year, while the cesium-137 concentrations were about 35 percent lower.

Table 10. Fission product concentration in milk July-December 1969

Date collected (1969)	Concentrations (pCi/liter)	
	Cesium-137	Strontium-90
July 3	20	9.5
August 6	33	6.6
September 2	36	3.1
October 8	30	6.5
November 5	30	8.0
December 3	27	5.8
Annual average	24	6.4

Conclusion

Argonne's contribution to the environmental radioactivity was primarily limited to tritium and argon-41 in the air and to tritium in water from Sawmill Creek and the Des Plaines River. The only activity detected off the site were the low levels of tritium in the two streams.

Recent coverage in *Radiological Health Data and Reports:*

Period	Issue
July–December 1968	February 1970
January–June 1969	July 1970

2. National Reactor Testing Station[3] January-June 1969

Health Services Laboratory
U.S. Atomic Energy Commission
Idaho Falls, Idaho

Data from the environmental monitoring network on and around the National Reactor Testing Station (NRTS) in eastern Idaho revealed that NRTS operations during the first half of 1969 did not contribute significantly to environmental radiation or radioactivity concentration levels. These levels remained well below the Atomic Energy Commission (AEC) standards as set forth in AEC Manual Chapter 0524. These standards are based on FRC recommendations. The concentrations of radioactivity reported include contributions from all sources, such as fallout and natural radioactivity. Samples of air, water and milk are collected routinely at stations shown in figure 3. The results of the analyses performed on the air, water, and milk samples are shown in table 11 for January–June 1969.

[3] Summarized from "Environmental Monitoring Report No. 24, January–June 1969," U.S. Atomic Energy Commission, Idaho Operations Office, Health Service Laboratory.

Water monitoring

Water containing small quantities of radioactivity is returned to the ground through ponds and wells at the NRTS. The radioactivity in the water is measured before discharge to make sure that standards are not exceeded in the environment. Radioactivity is also monitored in the ground water from onsite and offsite sampling locations. Samples are analyzed for gross alpha and gross beta radioactivity and tritium. Offsite surface water samples are also collected from the Snake River at Idaho Falls and Bliss. With the exception of tritium at some onsite sampling locations, the radioactivity in both onsite and offsite water was usually not subtracted from the levels measured onsite. The average concentrations of all types of radioactivity in onsite and offsite water were less than 11 percent of their respective AEC guides.

Air monitoring

Radioactive materials in the atmosphere are collected by drawing air through filters. The samplers operate continuously, and the filters are changed weekly. The filters are analyzed for gross alpha, gross beta and iodine-131 radioactivity.

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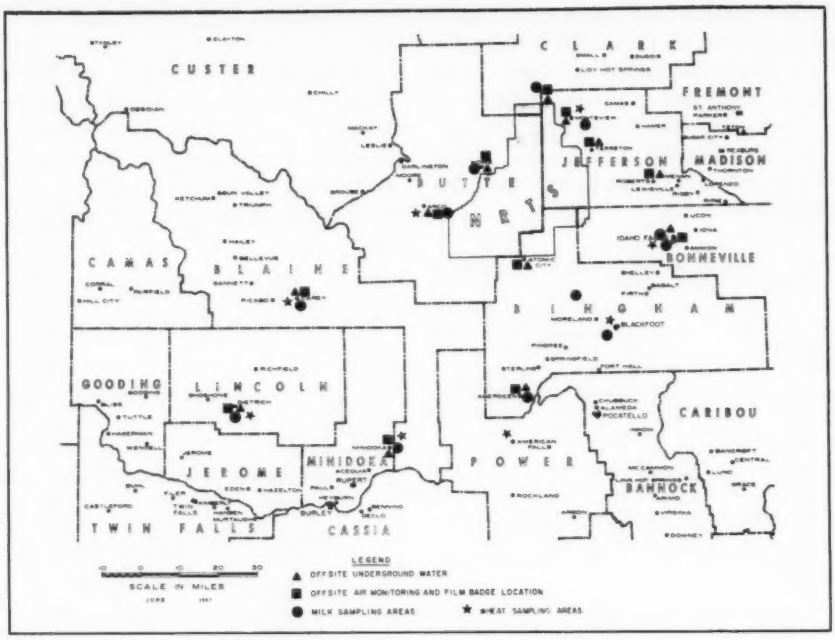

Figure 3. Environmental monitoring stations, National Reactor Testing Station

Table 11. Environmental monitoring data for the National Reactor Testing Station, January–June 1969

Type of sample and units	Number of samples	Type of radioactivity	Minimum level of detection	Maximum radioactivity of single sample	Average radioactivity per sample	AEC standard
Onsite production well water (pCi/liter)	244	Alpha	3	6	< 3	400
		Beta	5	66	< 6	3,000
		Tritium	[a] 2	[a] 131	[a] 8.9	[a] 100,000
Offsite underground water (pCi/liter)	12	Alpha	3	4	< 3	30
		Beta	5	5	< 5	100
		Tritium	[a] 2	[a] < 2	[a] < 2	[a] 3,000
Surface water (pCi/liter)	2	Alpha	3	3	< 3	30
		Beta	5	5	< 5	100
		Tritium	[a] 2	[a] < 2	[a] < 2	[a] 3,000
Onsite air (pCi/m³)	208	Alpha	.001	.023	.003	.6
		Beta	.003	1.9	.47	300
		Iodine–131	.01	.11	< .02	9,000
Offsite air (pCi/m³)	52	Alpha	.001	.013	.004	.2
		Beta	.003	1.6	.46	10
		Iodine–131	.01	.14	< .02	100
Offsite milk (pCi/liter)	91	Iodine–131	20	<20	<20	100
		Strontium–90	1	5	2	200
Offsite thermoluminescent dosimeters (mRem/6 months)	6	Gamma	5	6	< 5	85
On and offsite public highways:						
Highway 20	38	Gamma	5	18	< 5	85
Highway 88	19	Gamma	5	7	< 5	85
Highway 22	13	Gamma	5	< 5	< 5	85
Highway 28	8	Gamma	5	< 5	< 5	85

[a] Concentration of tritium in nCi/liter.

Samplers are located at eight locations onsite as shown in figure 3. Background concentrations are measured in Idaho Falls. The average gross beta concentration at the NRTS was not significantly different from background. This was also true for the average iodine-131 concentrations. On the other hand, the average concentration of gross alpha radioactivity continued to be significantly higher in Idaho Falls than at the NRTS. The average gross alpha concentration (including background) was less than 20 percent of the maximum permissible concentrations, if one assumes all the radioactivity was due to the most hazardous isotope that could possibly be in the air.

Because the alpha radioactivity levels observed are not much larger than the concentrations present from natural sources, it is very difficult to determine what isotope or isotopes are present. However, analyses made to date indicate that the increased levels in Idaho Falls are due to polonium-210 which occurs naturally in rocks and soil as a daughter product of natural uranium. If heated, polonium is readily volatilized into the atmosphere. In fact, some polonium is normally present in the air from decay of radon-222, which, being a gas, regularly diffuses from the ground.

Partly to find an explanation for the higher alpha radioactivity levels and partly to obtain a more representative background for NRTS measurements, another sampler was set up at Fanning Field in northwestern Idaho Falls. Also, samplers were set up northwest of Blackfoot and in Arco and Butte City.

The average gross alpha radioactivity measured in Idaho Falls was the same as that measured in Blackfoot. On the other hand, the levels in Butte City and Arco were lower, being roughly equal to those measured at the NRTS. This indicates that the higher gross alpha radioactivity levels are probably restricted to an area east of the NRTS. This radioactivity is not related to NRTS operations. All air monitoring results are listed in table 11.

Offsite milk monitoring

Offsite milk samples are collected from 12 locations routinely and analyzed for strontium-90 and iodine-131. A composite grade A sample is collected weekly from an Idaho Falls dairy. This composite sample contains milk from areas north, south and east of NRTS. Milk samples are collected monthly from all sampling locations. The sampling locations are shown in figure 3, and the results are given in table 11. No iodine-131 was detected at any time during the year. The strontium-90 concentrations are attributed to worldwide fallout and have shown a decline in recent years.

Gamma radiation levels

The external gamma radiation is measured with thermoluminescent dosimeters. The exposure results reported in table 11 represent net values after subtracting natural background exposures. Aberdeen, Carey, Deitrich, Idaho Falls, and Minidoka were used to determine background radiation exposure. The offsite locations are Arco, Atomic City, Howe, Monteview, Reno Ranch, and Roberts. No significant radiation exposures were observed.

Recent coverage in *Radiological Health Data and Reports:*

Period	Issue
January–June 1968	August 1969
July–December 1968	June 1970

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Reported Nuclear Detonations, December 1970

(Includes seismic signals presumably from foreign detonations)

Seismic signals, presumably from a Soviet underground nuclear explosion, were recorded by the United States on December 12, 1970. The signals originated at approximately 2:00 a.m., EST, in the Kazakh Desert, south of the Urals, and were equivalent to those of an underground nuclear test in the intermediate yield range (200 kilotons to 1 megaton TNT equivalent).

The U.S. Atomic Energy Commission conducted two underground nuclear tests at its Nevada Test Site on December 16, 1970. The first test was in the low yield range of less than 20 kilotons TNT equivalent. The second test was in the low-intermediate yield range (20 to 200 kilotons TNT equivalent).

There was some release of radioactivity from the first test of December 16, 1970. Radioactivity at very low levels, (less than 1 milliroentgen per hour) was recorded on the ground along State Highway 25, north of the test site. Monitors and aircraft of the Environmental Protection Agency (formerly the U.S. Public Health Service) monitored the situation continuously. The radioactive air mass moved slowly north and northwest, widening gradually as it moved. Aircraft kept the air mass defined and instruments recorded levels of a few milliroentgens per hour several thousand feet above the ground. None of the readings indicated that there would be a health hazard to persons living downwind.

On December 17, 1970, the United States recorded seismic signals which originated from the Soviet nuclear test area in the Semipalatinsk region. The signals were equivalent to those of a nuclear test in the low-intermediate yield range (20–200 kilotons TNT equivalent).

Also on December 17, 1970, the Atomic Energy Commission announced that it had conducted an underground nuclear test in the intermediate yield range (200 kilotons to 1 megaton TNT equivalent) at its Nevada Test Site.

Seismic signals were recorded by the United States on December 23, 1970, presumably from a Soviet underground nuclear explosion. The signals originated from the Kazakh Desert south of Urals, and were equivalent to those of an underground nuclear test in the intermediate yield range of 200 kilotons to 1 megaton.

Information in this section is based on data received during the month, and is subject to change as additional information may become available. Persons requiring information for purposes of compiling announced nuclear detonation statistics are advised to contact the Division of Public Information, U.S. Atomic Energy Commission, Washington, D.C. 20545.

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SYNOPSES

Synopses of reports, incorporating a list of key words, are furnished below in reference card format for the convenience of readers who may wish to clip them for their files.

ESTIMATES OF POTENTIAL DOSES TO VARIOUS ORGANS FROM X-RADIATION EMISSIONS FROM COLOR TELEVISION PICTURE TUBES. *Robert H. Neill, Harry D. Youmans,* and *John L. Wyatt. Radiological Health Data and Reports, Vol. 12, January 1971,* pp. 1–6.

Using the exposure rates determined in the metropolitan Washington color television survey, doses to various organs of children and adults were calculated. For a television receiver with a picture tube emitting 0.5 mR/h (the maximum exposure rate allowed under the Department of Health, Education and Welfare performance standards) and with other specified assumptions, the calculated dose to a child ranged from 5 mrem/a (female gonads) to 42 mrem/a (skin). For an adult, the doses were 2 mrem/a to 18 mrem/a, respectively. Using the same assumptions but for a picture tube emitting 0.043 mR/h (the average exposure rate observed in the metropolitan Washington survey), the respective doses were 0.4 mrem/a to 3.6 mrem/a for a child and 0.2 mrem/a to 1.6 mrem/a for an adult.

KEYWORDS: Doses, color television, organs, x radiation.

SURVEYS OF RADIUM CONTAMINATION IN AIRCRAFT INSTRUMENT REPAIR FACILITIES, 1966–1967. *Gail D. Schmidt, Jerome A. Halperin, Eric Geiger,* and *Walter Livingstone. Radiological Health Data and Reports, Vol. 12, January 1971,* pp. 7–15.

A survey of 30 aircraft instrument repair facilities revealed many cases of significant radium contamination, resulting from repair and dial stripping operations on radium luminous dial instruments. Five out of 15 facilities surveyed in Wichita, Kans. showed alpha contamination exceeding 200,000 dpm/100 cm^2, with a maximum of 1,650,000 dpm/100 cm^2. The direct gamma radiation ranged from background to 60 mR/h range. Whole body counts of 33 persons working in these facilities showed radium body burdens in two people that were about 40 percent of the maximum permissible body burden. A similar survey of 15 repair facilities in Dade County, Florida, showed lower alpha contamination levels by a factor of 10.

Based upon these findings, a radiation control program including contamination guides has been recommended for instrument repair facilities.

KEYWORDS: Radium, repair facilities, aircraft instruments.

ENVIRONMENTAL SURVEY OF URANIUM MILL TAILINGS PILE, MEXICAN HAT, UTAH. *Robert N. Snelling. Radiological Health Data and Reports, Vol. 12, January 1971,* pp. 17–28.

At the request of the Navajo Tribal Council through the PHS Division of Indian Health at Window Rock, Ariz., an environmental radiological survey was conducted on the A–Z Minerals Corporation uranium tailings pile, Mexican Hat, Utah, in May 1968. The purpose of the survey was to identify any radiation hazards which might exist and recommend methods for their control. The survey included evaluation of external gamma radiation and airborne and waterborne radioactivity.

The results of the survey indicate that the external radiation levels on the tailings area exceed recommended exposure limits for individuals in the general population. Therefore, the area should not be released for public use in its present state. Action which would permit the release of the area would involve covering the tailings with uncontaminated soil to an extent that would diminish the external radiation to an acceptable level and to stabilize the covering against wind erosion.

Radiation levels in air and water do not exceed recommended exposure limits. However, to minimize the possibility of increased activity from weather conditions different from those existing during the survey, it is recommended that the tailings be stabilized against wind erosion or periodic monitoring will be necessary.

KEYWORDS: Air, environmental survey, exposure limits, gamma radiation, mill tailings, uranium, Utah, water.

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